



1 **Improving maps of forest aboveground biomass: A**
2 **combined approach using machine learning with a**
3 **spatial statistical model**

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22 **Abstract:** Aboveground biomass (AGB) estimates at the plot level plays a major part in connecting
23 accurate single-tree AGB measurements to relatively difficult regional-scale AGB estimates. However,
24 complex and spatially heterogeneous landscapes, where multiple environmental covariates (such as
25 longitude, latitude, and forest structure) affect the spatial distribution of AGB, make upscaling of
26 plot-level models more challenging. To address this challenge, this study proposes an approach that
27 combines machine learning with spatial statistics to construct a more accurate plot-level AGB model.
28 The study was conducted in a *Eucalyptus* plantation in Nanjing, China. We developed, evaluated, and
29 compared the accuracy and performance of three different machine learning models [support vector
30 machine (SVM), random forest (RF), and the radial basis function artificial neural network
31 (RBF-ANN)], one spatial statistics model (P-BSHADE), and three combinations thereof (SVM &
32 P-BSHADE, RF & P-BSHADE, RBF-ANN & P-BSHADE) for forest AGB estimates based on AGB
33 data from 30 sample plots and their corresponding environmental covariates. The results show that the
34 performance indices RMSE, nRMSE, MAE, and MRE of all combined models are substantially
35 smaller than those of any individual models, with the RF & P-BSHADE combined method giving the
36 smallest value. These results demonstrate clearly that combined models, especially the RF &
37 P-BSHADE model, can improve the accuracy of plot-level AGB models and reduce uncertainty on
38 plot-level AGB estimates or even on large-forested-landscape AGB estimates. These research results
39 are important because they reduce the uncertainty in estimates of the regional carbon balance.

40

41 **Keywords:** Aboveground biomass, plot-level model, Machine learning, Spatial statistical model

42



43 **1 Introduction**

44 Accurate maps of aboveground biomass (AGB) provide a solid foundation for sound decision-making in
45 sustainable forest management scenarios, such as reducing deforestation, forest degradation, and
46 greenhouse-gas emissions (Bustamante et al., 2016; Houghton et al., 2009; Mendoza-Ponce and Galicia,
47 2010). Most AGB maps are constructed based on plot-level estimation models, which are challenging to
48 scale up and can ultimately propagate uncertainty to regional AGB maps. The uncertainty of such
49 regional maps can be attributed to two primary sources: (1) the use of inadequate sampling data to
50 construct the plot level prediction models, and (2) model-dependent uncertainty, including
51 unreasonable model-parameter assumptions and improper model structure (Chen et al., 2015; Gao et al.,
52 2016; McRoberts et al., 2016). The present study mainly focuses on reducing the second source of
53 uncertainty.

54 An estimated 18%–103% of the uncertainty in AGB mapping can be attributed to model-dependent
55 uncertainty (Djomo and Chimi, 2017; Malhi et al., 2004). Although the allometric model, which is the
56 most popular plot-level model, has produced useful results for forest AGB estimates (Conti et al., 2019;
57 Huang et al., 2019), selection error in plot-level allometric modeling still leads to over 40% uncertainty
58 (Djomo et al., 2016; Fayolle et al., 2013; Chave et al., 2014), and simple or complex forms of the
59 allometric model account for 20%–60% of the uncertainty (Picard et al., 2015).

60 Many different plot-level prediction models other than allometric models have been applied to
61 constructing accurate AGB maps, including linear models (Andersen et al., 2014; Morel et al., 2012),
62 machine learning models (Chen, 2015; Gleason and Im, 2012), and spatial statistical models (Benitez et
63 al., 2016; Propastin, 2012; Van der Laan et al., 2014). With the development of computer-science
64 techniques and advances in nonlinear biomass modeling, machine learning methods have become
65 prevalent. Traditional parametric methods, which summarize data with a fixed number of parameters
66 based on sample size (e.g., logistic regression and perceptron) (Gao and Hailu, 2012), have difficulty
67 characterizing nonlinear relationships between AGB and multiple environmental covariates. By
68 comparison, nonparametric machine learning algorithms, in which the number of parameters depends
69 on the number of training examples (e.g., K-nearest neighbor, support vector machine, and random
70 forest), are advantageous because they are more elastic and do not restrict variable types, the
71 distribution of predictor variables, or the relationship between response and predictor variables (Lu et



72 al., 2007). In addition, nonparametric machine learning algorithms may offer higher prediction accuracy
73 (Frey et al., 2019; Gleason and Im, 2012).

74 Another group of models frequently used to estimate the relationship between forest AGB and multiple
75 environmental covariates is based on spatial statistical approaches, including geographically weighted
76 regression and Kriging (Du et al., 2010; Van der Laan et al., 2014; Viana et al., 2012). Spatial statistical
77 methods are based on analyses of attribute information, such as spatial location (Schabenberger and
78 Gotway, 2005). Compared with traditional statistical methods, spatial methods integrate spatial factors
79 that affect model responses, thus removing the constraints of traditional statistical methods that assume
80 sample independence (Rangel and Bini, 2010) and improving our understanding of spatial
81 autocorrelation and heterogeneity (He et al., 2011; Rosenberg and Anderson, 2011).

82 Although many studies have integrated ground-based plot data, multi-source remote-sensing data (e.g.,
83 LiDAR and Landsat), and machine learning or spatial statistical methods, the prediction accuracy of
84 current AGB spatial mapping still suffers from uncertainty (McRoberts et al., 2018; Paul et al., 2016;
85 Saatchi et al., 2011; Zheng et al., 2004; Jachowski et al., 2013; Zhang et al., 2014). First, existing
86 studies that used machine learning methods have not considered the spatial heterogeneity of multiple
87 environmental covariates (such as longitude, latitude, and forest structure), which affects the spatial
88 distribution of AGB (Babcock et al., 2015; Fassnacht et al., 2014). Second, the assumptions of the spatial
89 statistical method (e.g., spatial autocorrelation and spatial stratified heterogeneity) may not always apply
90 to forest AGB.

91 AGB estimates at the plot level serve as a bridge to connect single-tree AGB measurements to AGB
92 estimates on a regional scale. Accurate AGB mapping at the plot scale provides a basis for future
93 upscaling to the regional scale. However, the uncertainty and error propagation inherent in different
94 prediction models make this process challenging. Allometric models are most commonly used to
95 construct plot-level AGB models, but they cannot fully capture the complex and spatially
96 heterogeneous landscapes where multiple environmental covariates (such as longitude, latitude, and
97 forest structure) affect the spatial distribution of AGB. The objective of the present study is to develop
98 and evaluate a combined machine learning and spatial statistical method that uses ground-based samples
99 to improve the prediction accuracy of AGB spatial mapping at the plot level. The proposed method



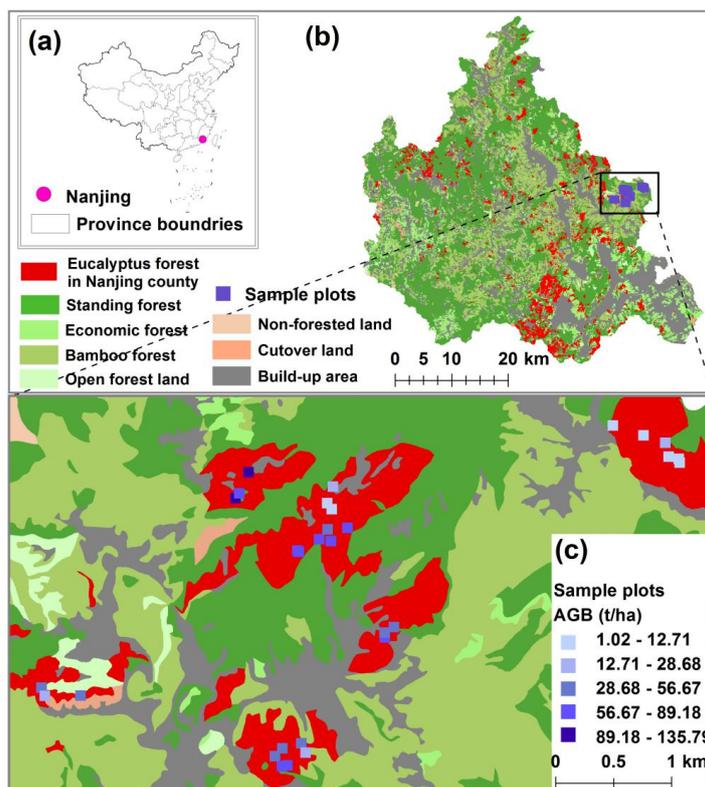
100 integrates the nonlinear mapping capabilities of machine learning algorithms [i.e., radial basis function
101 artificial neural network (RBF-ANN), support vector machine (SVM), and random forest (RF)] with the
102 spatial autocorrelation and stratified heterogeneous advantages of a spatial statistical model (i.e., the
103 point estimation model of biased sentinel hospital-based area disease estimation, P-BSHADE) (Xu et al.,
104 2013). Our aim is to answer two specific questions: (1) What are the differences in prediction accuracy
105 of AGB maps based on different methods? (2) Can the integration of spatial statistical and machine
106 learning methods improve the accuracy of AGB models at the plot level? We explore these two
107 questions by studying an empirical case for predicting an AGB map at a *Eucalyptus* plantation in
108 Nanjing County, China.

109 **2 Materials and Methods**

110 **2.1 Site description**

111 Nanjing County (117°00'–117°36'E, 24°26'–25°00'N, Fig. 1b) is located in the upstream region of the
112 Jiulong River in Fujian Province, China. Seventy-four percent (145 009 ha) of the county comprises
113 forests and 79 346 ha are plantations. The region is affected by the South Asian tropical monsoon climate.
114 In 2014, the average annual temperature in Nanjing County was 21.1°C, with an annual precipitation of
115 1700 mm and 340 frost-free days. The major soil type is red soil.

116 The study area has a complex topography with significantly varying elevation (0–1566 m). Forest
117 composition, structure, and biomass are spatiotemporally heterogeneous. The main tree species are
118 *Eucalyptus grandis x urophylla*, *Pinus massoniana*, and *Cunninghamia lanceolata*. Recently, the area of
119 *Eucalyptus* plantations has increased rapidly, reaching 13 338 ha, which is an increase of 10 862 ha in
120 one decade.



121

122 Figure 1. The study area is a typical example of a non-representative-sample problem. (a)
123 Geographical location of the study area. (b) Spatial distribution of *Eucalyptus* plantations (red) and
124 other major forests. (c) Spatial distribution of the 30 sample plots used in this study (blue).

125 2.2 Data collection

126 2.2.1 Non-destructive sampling in sample plots

127 A total of 30 fixed sample plots were selected in 2012 from the Yongfeng forest farm. The plots were
128 located in the eastern section of the study area (Fig. 1). The 30 sampling plots included ten *Eucalyptus*
129 plantation age groups. In each plot (0.04 ha, 20 m × 20 m), we measured the diameter at breast height
130 (DBH) of all living stems ≥ 8 cm and the tree height (H). In addition, we measured mean plot-level
131 variables, including stand age, density, longitude, latitude, and altitude.



132 **2.2.2 Destructive sampling in sample plots: Tree harvest**

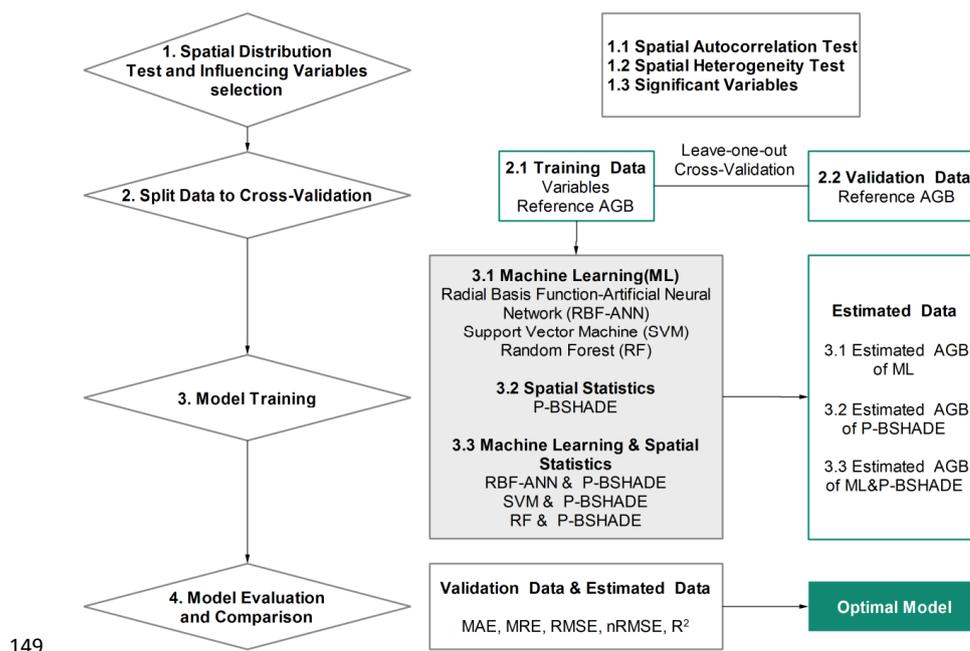
133 Trees were harvested from standard woods in the 30 fixed sample plots. Three trees with a DBH close
134 to mean DBH of trees in each plot were cut down, for a total of 90 trees harvested from the 30 plots.
135 We then measured the H and DBH of each harvested tree, as well as the biomass of each organ (foliage,
136 stems, and branches) to obtain the AGB of each harvested tree. Table B.2 in section S2 of the
137 Supplementary Material presents the data for the 90 harvest trees. Details on selection of the standard
138 wood and the cutting process are provided in section S1 of the Supplementary Material.

139 **2.3 Construction of tree-level allometric models**

140 All analyses were based on the underlying assumption that the relationship between the response and
141 predictor variables in the sample data used to construct the models was the same as the relationship in
142 the entire population. We divided the 90 harvested trees into three age groups (1–2 yr, 3–5 yr, 6–10 yr)
143 for the tree-level allometric models. The allometric models were then applied to each tree in each
144 sample plot according to their age, DBH, and H, thereby producing a true measure of AGB for each
145 sample plot.

146 **2.4 Construction of plot-level models**

147 Processing based on model screening was applied to alleviate uncertainty caused by model-dependence
148 and consisted of the four steps shown in Fig. 2.



149

150 Figure 2. Workflow for screening an optimal model.

151 **2.4.1 Selection of variables and analysis of resulting spatial distribution**

152 To create the plot-level model, we first identified predictor variables. Based on our previous work (Ren
 153 et al., 2017), we selected plot-level environmental covariates including longitude and altitude, and forest
 154 attribute variables including forest distribution density, DBH, H, tree stem volume, and forest age.
 155 Pearson’s correlation coefficient was used to investigate the correlation between these variables and the
 156 true AGB of sample plots.

157 We then analyzed the spatial autocorrelation and spatial heterogeneity of AGB data from the selected
 158 sample plots. We used Moran’s *I* (Cliff and Ord, 1981), a commonly used global spatial autocorrelation
 159 index, to evaluate spatial autocorrelation between the true AGBs of sample plots. The spatial stratified
 160 heterogeneity (which refers to the within-strata variance being less than the between-strata variance; it
 161 is ubiquitous in ecological phenomena, such as AGB) of the true AGB of sample plots was evaluated
 162 by using a *q*-statistic generated by applying the GeogDetector model, which is a software tool proposed
 163 by Wang et al. (2016) that analyzes spatial variation of the geographical strata of variables. First, we
 164 used the K-means algorithm to obtain the strata of true AGB for preprocessing by GeogDetector. Next,



165 we regarded the true AGB as Y , the strata of true AGB as X , and put them into the GeogDetector
166 model to obtain the q -statistic (Wang et al., 2010; Wang et al., 2016).

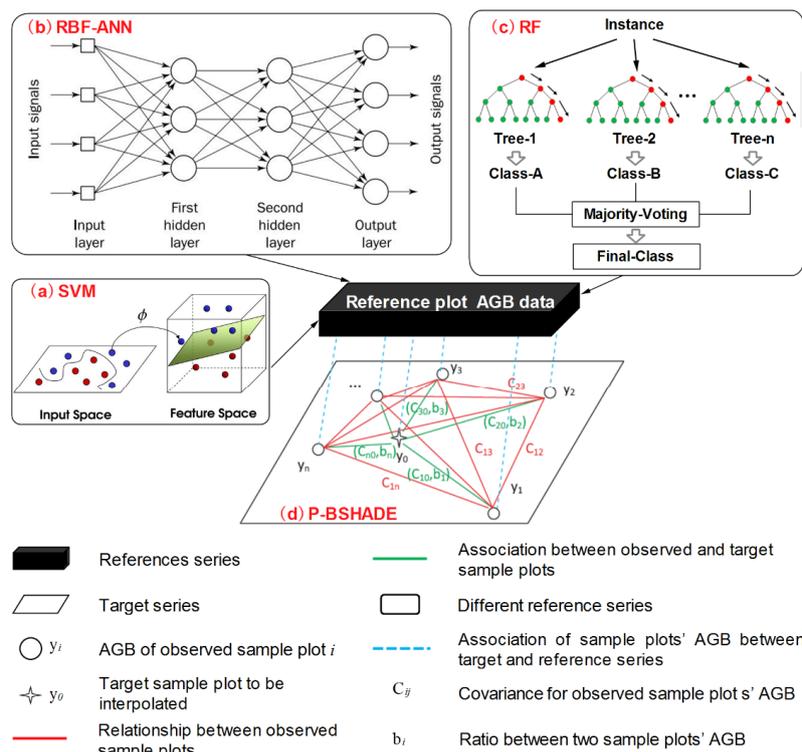
167 **2.4.2 Split datasets**

168 We used the leave-one-out cross-validation method to split the 30 sample plots into 30 sets, with each set
169 containing two groups of data: (1) validation data (the AGB of one plot) and (2) training data (the AGBs
170 and predictor variables of the other 29 plots), see Table B.3. The leave-one-out cross-validation method
171 assumes that, in a dataset containing n samples, each sample serves as a test sample with the other $n - 1$
172 samples serving as training samples. Thus, with n iterations, we can obtain n training datasets and n
173 validation datasets.

174 **2.4.3 Model training**

175 Seven models including three machine learning models [Figs. 3(a–3(c))], one spatial statistical model
176 [Fig. 3(d)], and three combined machine learning and spatial statistical models [Figs. 3(a) and 3(d), 3(b)
177 and 3(d), and 3(c) and 3(d)] were developed and trained to predict the AGB of sample plots. The three
178 machine learning models were (a) SVM, (b) RBF-ANN, and (c) RF.

179 The spatial statistical model (P-BSHADE) required AGB-related variables (reference series). In this
180 case study, we used the reference-plot AGB data as the variables. The allometric model (Qiu et al.,
181 2018) was applied to obtain the AGB of each tree in each sample plot. Next, the reference-plot AGB
182 data consisted of the sum of the AGB of each tree. This method produces the P-BSHADE model shown
183 in Fig. 3(d). For the combined machine learning and spatial statistical models, the reference plot AGB
184 data in P-BSHADE were obtained from the results of the SVM [Fig. 3(a)], the RBF-ANN [Fig. 3(b)], or
185 the RF [Fig. 3(c)]. The three combined models are denoted SVM & P-BSHADE [Figs. 3(a) and 3(d)],
186 RBF-ANN & P-BSHADE [Figs. 3(b) and 3(d)], and RF & P-BSHADE [Figs. 3(c) and 3(d)]. Each
187 model was trained on 30 datasets, yielding a total of 30 predicted AGB datasets for 30 sample plots (see
188 Table B.3, section S2 in the Supplementary Material).



189

190

191 Figure 3. Framework for estimating (a)–(c) the machine learning models, (d) the P-BSHADE model,
 192 and the three models that combine machine learning with the P-BSHADE model (a+d, b+d, c+d).

193

194 (1) Machine learning

195 SVM is a method of supervised learning in machine learning and is often used to solve classification
 196 problems. The basic principle of SVM is to find a hyperplane in the feature space and separate the
 197 positive and negative samples with the minimum misclassification rate (Hearst et al., 1998). RBF-ANN
 198 is a three-layer neural network model, which includes an input layer, a hidden layer, and an output layer.
 199 The transformation from input space to hidden space is nonlinear, whereas the transformation from
 200 hidden space to output space is linear. The function of the hidden layer is to map the vector from the
 201 indivisible low-dimensional linear state to the separable high-dimensional linear state, so as to greatly



202 accelerate the learning and convergence speed and avoid getting stuck in a local optimum (Elanayar and
203 Shin, 1994; Xia and Xiu, 2007). RF is a combination of tree predictors such that each tree depends on
204 the values of a random vector sampled independently and with the same distribution for all trees in the
205 forest. RF is an effective tool in prediction. Because of the Law of Large Numbers, RF does not overfit.
206 Injecting the right type of randomness means that RF makes accurate classifiers and regressors (Breiman,
207 2001).

208 The schematic function for machine learning is

$$209 \quad y_j = f(x_{j,1}, x_{j,2}, x_{j,3}, x_{j,4}) \quad (1)$$

210 where y_j is the AGB of the j th sample plot predicted by a machine learning model, $f(\dots)$ is a machine
211 learning model represented by a function of $x_{j,k}$ ($k = 1, \dots, 4$); and $x_{j,1}$, $x_{j,2}$, $x_{j,3}$, and $x_{j,4}$ are the
212 central longitude, the mean DBH, the mean H, and the forest age of the j th sample plot, respectively. A
213 specific description of the three machine learning models is given in section S1 of the Supplementary
214 Material.

215 (2) Spatial statistical model: P-BSHADE

216 P-BSHADE is an optimal linear unbiased estimation interpolation method based on the assumption of
217 the simultaneous existence of the spatial autocorrelation and heterogeneity of the target object. We use
218 it here to solve the problem of an unrepresentative sample imposed by the spatial location of a
219 convenient sample at the plot level.

220 The core of the model is to minimize the variances between predicted error and unbiased estimation.
221 The prediction process of the P-BSHADE model requires strong spatio-temporal coordination between
222 the predictive variable (forest AGB of target plots) and the reference series (reference forest AGB of
223 target plots), so as to realize the spatial interpolation of the predictive variable. The model is also a data
224 fusion approach that combines the observed samples with the reference series (related variable).

225 P-BSHADE is markedly different from the Kriging and Inverse Distance Weighting (IDW) algorithms.
226 Compared with Kriging and IDW, the application of P-BSHADE to forest AGB interpolation has
227 obvious advantages. The spatial distribution of forest AGB is also characterized by spatial
228 autocorrelation and heterogeneity, which have been taken into account in the P-BSHADE model.



229 Taking into account spatial heterogeneity can effectively solve the difference in forest AGB
230 distribution caused by different terrain or geographical location. However, Kriging and IDW only
231 consider the spatial correlation between plots. In addition, P-BSHADE considers strongly correlated
232 sample plots as neighboring plots, whereas the Kriging and IDW algorithms consider sites that are
233 close in proximity.

234 In brief, the P-BSHADE model includes two steps. First, it obtains reference AGB for all sample plots
235 by using the allometric model. Second, it uses the reference AGB of the target sample plot and the true
236 AGB of other sample plots to obtain the weight relationship between the target sample plot and the
237 other sample plots and puts the true AGB of other sample plots and the weights into Eq. (2) to predict
238 the AGB of the sample plots. Therefore, positions and distances between plots do not apply here. The
239 specific mathematical formula for the P-BSHADE model is now described (Hu et al., 2013; Xu et al.,
240 2013).

241 a. Objective

242 The objective is to interpolate the AGB data of the target sample plot by using data acquired from other
243 sample plots. A theoretical description is

$$244 \hat{y}_j = \sum_{i=1}^n w_{ij} y_i \quad (2)$$

245 where \hat{y}_j is the AGB of the j th sample plot estimated by the P-BSHADE model ($j = 1 - 30, n =$
246 30); y_i is the true AGB of the i th sample plot ($i = 1 - 30, n = 30$); w_{ij} is the weight (contribution)
247 of the true AGB of the i th sample plot to the AGB to be interpolated of the j th sample plot (when $j =$
248 $1, i = 2, 3, \dots, 30$; when $j = 1, i = 1, 3, 5, \dots, 30$); w_{ij} is calculated by the true AGB of the i -th
249 sample plot and the allometric model estimation of the AGB in the j -th sample plot.

250 As expected, the estimates of the two properties in Eq. (2) are unbiased:

$$251 E(y_j) = E(\hat{y}_j) \quad (3)$$

252 Minimum estimation variance is expressed as

$$253 \min_w \left[\sigma_{\hat{y}_j}^2 = E(\hat{y}_j - y_i)^2 \right] \quad (4)$$

254 where E is the statistical expectation.



255 **b. Ratio of data from target sample plot to those from other sample plots**

256 The ratio between data from the target sample plot to those from other sample plots is one of the most
 257 important inputs for estimating the AGB of the target sample plot and is an index of heterogeneity in
 258 the AGB spatial distribution. The relationship between data from the target sample plot and from the
 259 other sample plots is expressed as

$$260 \quad b_{ij} E y_j = E y_i \quad (5)$$

261 In most cases, the AGB of any two plots are not equal, and the relationship between them can be
 262 further expressed as the relative bias b_{ij} between the mathematical expectation of y_j and y_i .
 263 Considering Eq. (2), Eq. (5) can be written as

$$264 \quad \sum_{i=1}^n w_{ij} b_{ij} = 1 \quad (6)$$

265 This equation is generally valid for nonhomogeneous conditions. Clearly, the determination of b_{ij}
 266 requires calculating the coefficients w_{ij} ($i = 1, \dots, n, j = 1, \dots, n$), which is addressed in the following
 267 section.

268

269 **c. Weight estimation**

270 The main challenge in estimation is finding the weights w_{ij} that satisfy the unbiased condition and
 271 that minimize estimation variance:

$$272 \quad \sigma_{\hat{y}_j}^2 = E(\hat{y}_j - y_i)^2 = C(\hat{y}_j \hat{y}_j) + C(y_i y_i) - 2C(\hat{y}_j y_i) \quad (7)$$

273

274 These weights can be calculated by minimizing the estimation variance and taking unbiasedness into
 275 account:

$$276 \quad \begin{bmatrix} C(y_1 y_1) & \dots & C(y_1 y_n) & b_{1j} \\ \vdots & \ddots & \vdots & \vdots \\ C(y_n y_1) & \dots & C(y_n y_n) & b_{nj} \\ b_{1j} & \dots & b_{nj} & 0 \end{bmatrix} \begin{bmatrix} w_{1j} \\ \vdots \\ w_{nj} \\ \mu \end{bmatrix} = \begin{bmatrix} C(y_1 y_j) \\ \vdots \\ C(y_n y_j) \\ 1 \end{bmatrix} \quad (8)$$

277

278 where μ is a Lagrange multiplier. The minimized variance in the estimation error can then be written
 279 as

$$280 \quad \sigma_{\hat{y}_j}^2 = \sigma_{y_i}^2 + \sum_{i=1}^n \sum_{k=1}^n C(y_i y_k) - 2 \sum_{i=1}^n w_{ij} C(y_i y_j) + 2\mu (\sum_{i=1}^n w_{ij} b_{ij} - 1) \quad (9)$$



281

282 The P-BSHADE model is a geospatial model because it has the following characteristics:

283 1. The P-BSHADE model is mainly based on the assumptions of spatial autocorrelation and spatial
284 heterogeneity of forest AGB. Therefore, before using P-BSHADE, we first applied the statistical test of
285 these two theoretical hypotheses (spatial autocorrelation test and spatial differentiation test) for forest
286 AGB.

287 2. The prediction process of the P-BSHADE model requires strong spatio-temporal coordination
288 between the predictive variable (forest AGB of target plots) and the reference sequence (reference
289 forest AGB of target plots), so as to spatially interpolate the predictive variable.

290 3. P-BSHADE is an optimal linear unbiased estimation interpolation method that considers temporal
291 and spatial heterogeneity. Spatial autocorrelation and heterogeneity of AGB data can be added into the
292 model based on prior knowledge (reference AGB data), following which the linear unbiased optimal
293 estimation of the target-plot AGB can be obtained by correcting data from a convenient sample plot.

294 Specifically, for example, the ratio of data from the target sample plot to that from other sample plots is
295 used [see 2.4.3(2)b section]. In the P-BSHADE model, this ratio plays a very important role in
296 estimating the forest AGB of the target plots. This ratio is a manifestation of the spatial heterogeneity
297 of AGB data. P-BSHADE takes into account the reality of the spatial distribution of AGB data and
298 emphasizes that the spatial distribution of AGB data is heterogeneous.

299 (3) Combination of machine learning and spatial statistical models

300 Considering the inherent advantages and disadvantages of P-BSHADE and machine learning, this study
301 investigates whether their combination can improve the accuracy of forest AGB estimates. Therefore,
302 P-BSHADE was separately integrated with the three machine learning methods (SVM, RBF-ANN, and
303 RF) to form three combined models (SVM & P-BSHADE, RBF-ANN & P-BSHADE, and RF &
304 P-BSHADE). The reference AGBs of the 30 sample plots were replaced by the estimates produced by
305 the machine learning models. Each combined model was represented as follows:

306
$$\hat{y}_j = \sum_{i=1}^n w_{ij} y_i \quad (10)$$



307 where \hat{y}_j is the estimated AGB of the j th sample plot using the combined model ($j =$
308 $1, 2, \dots, 30, n = 30$); y_i is the true AGB of the i th sample plot ($i = 1, 2, \dots, 30, n = 30$); w_{ij} is the
309 contribution in weight of the i th true AGB of the sample plot to the j th sample plot AGB to be
310 interpolated (when $j = 1, i = 2, 3, \dots, 30$; when $j = 1, i = 1, 3, 5, \dots, 30$); w_{ij} is calculated by
311 using the true AGB of the i th sample plot and the machine learning estimate of the AGB of the j th
312 sample plot. A detailed description of the combined models and the algorithm formulas is presented in
313 section S1 of the Supplementary Material.

314 2.4.4 Model evaluation and comparison

315 To evaluate the accuracy of the AGB estimates of the seven models (SVM, RBF-ANN, RF, P-BSHADE,
316 SVM & P-BSHADE, RBF-ANN & P-BSHADE, and RF & P-BSHADE), the AGB results were
317 compared to the reference AGBs of the sample-plot groups (AGB group M in Table B.3). We calculated
318 four performance indicators, as given by Eqs. (11)–(14) [mean absolute error (MAE), mean relative
319 error (MRE), root mean square error (RMSE), and normalized root mean square error (nRMSE)]:

$$320 \text{ MAE} = (\sum_{i=1}^n |y_i^p - y_i|) / n \quad (11)$$

$$321 \text{ MRE} = (\sum_{i=1}^n |y_i^p - y_i|) / (y_i \times n) \quad (12)$$

$$322 \text{ RMSE} = \sqrt{(\sum_{i=1}^n (y_i^p - y_i)^2) / n} \quad (13)$$

$$323 \text{ nRMSE} = \frac{\sqrt{(\sum_{i=1}^n (y_i^p - y_i)^2) / n}}{\bar{y}_i} \quad (14)$$

324 where y_i^p is the predictive value of the different models, y_i is the AGB of the i th sample plot, and n
325 is the number of training datasets.

326 We then used the calculated MAE, MRE, RMSE, and nRMSE to identify the optimal model.

327 2.4.5 Robustness of combined models

328 To evaluate the robustness of the combined machine learning and spatial statistical models, we selected
329 22 independent sample plots (see details in S1 and S3 of the Supplementary Material) and made
330 nondestructive measurements of each tree in July 2019. We repeated the workflow used for
331 constructing the plot-level model and evaluated the models. We then evaluated whether the combined



332 models produced higher accuracy than the plot-level models by using the accuracy-assessment indexes
333 (MAE, MRE, RMSE, and nRMSE).

334 **2.5 Model application and upscaling**

335 We treated the irregular polygon forest patches (2980 patches) of the Forest Management and Planning
336 Inventory (FMPI) as a homogenous sample plot and used the optimal plot-level model to upscale forest
337 AGB (see section S1 of the Supplementary Material). We then compared the upscaled forest AGB with
338 the AGB map obtained from the allometric model and calculated the MRE of AGB between the two
339 methods (see Eq. A.15 in section S1 of the Supplementary Material).

340 **3 Results**

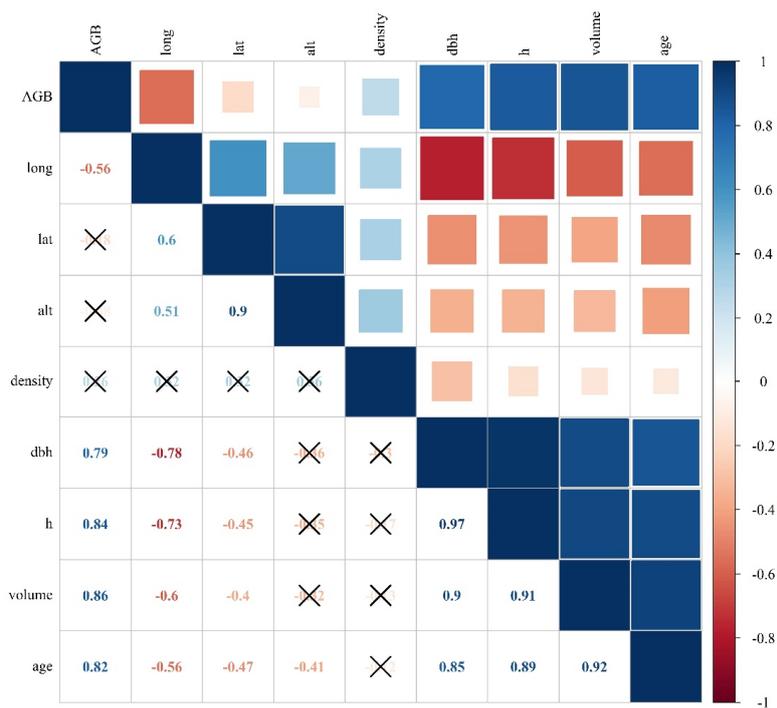
341 **3.1 True AGB of sample plots**

342 The true AGB for the 30 sample plots ranged from 1.02 to 135.79 Mg·ha⁻¹, with an average value of
343 47.34 Mg·ha⁻¹ and a standard deviation of 34.46 Mg·ha⁻¹. The coefficients of variation of the AGB for
344 all the sample plots and for the 10 age categories were 0.73 and 0.07–0.37, respectively.

345 **3.2 Spatial distribution test and the selection of variables**

346 **3.2.1 The effect of different variables**

347 Figure 4 shows the correlation-coefficient matrix of variables. The following variables were strongly
348 correlated with AGB: longitude ($r = -0.56$), DBH ($r = 0.79$), H ($r = 0.84$), trunk volume
349 ($r = 0.86$), and forest age ($r = 0.82$). Timber volume and stem volume were both estimated based on
350 H and DBH, so they were excluded as covariates for the AGB plot-level models. To summarize, four
351 variables (longitude, DBH, H, and forest age) were selected as covariates for the AGB plot-level
352 models of the *Eucalyptus* forest in the Nanjing region. Table B.4 in section S2 of the Supplementary
353 Material lists the statistical descriptions of these covariates and the AGB statistics for the 30 sample
354 plots.



355

356

357 Figure 4. Pearson's correlation coefficients between AGB and other variables represented by numbers
 358 and squares. Negative (red) numbers indicate that the corresponding variables are negatively correlated
 359 and are colored in red, whereas positive (blue) numbers represent positive correlations. Larger absolute
 360 numbers are indicated by darker colors, larger squares indicate stronger correlations, and the symbol "x"
 361 " indicates insignificant correlations.

362 3.2.2 Spatial autocorrelation test

363 The spatial distribution of the true AGBs of the 30 sample plots displayed a pattern of aggregation (see
 364 red regions in Fig. C.1, section S3 of the Supplementary Material and Table 1). In addition, because
 365 less than 1% of the AGB data were randomly distributed (see blue regions in Figs. C.1 and S3 of the
 366 Supplementary Material and Table 1), the possibility of an aggregated distribution was greater than that
 367 of random distribution. Furthermore, the null hypothesis was significantly rejected ($p < 0.01$). These
 368 results suggest that the spatial distribution of the AGB data displays aggregation and a pattern of strong



369 spatial autocorrelation.

370 Table 1. Spatial autocorrelation and heterogeneity test.

| Spatial autocorrelation | | Spatial heterogeneity | | |
|-------------------------|--------|--------------------------------|----------------|----------------|
| Items | Values | Factors | <i>q</i> value | <i>p</i> value |
| Moran I | 0.36 | AGB | 0.87 | <0.01 |
| | | Longitude, long | 0.38 | <0.01 |
| <i>z</i> -score | 4.78 | Diameter at breast height, DBH | 0.54 | <0.01 |
| <i>p</i> -value | 0.00 | Tree height, H | 0.63 | <0.01 |
| | | Age | 0.92 | <0.01 |

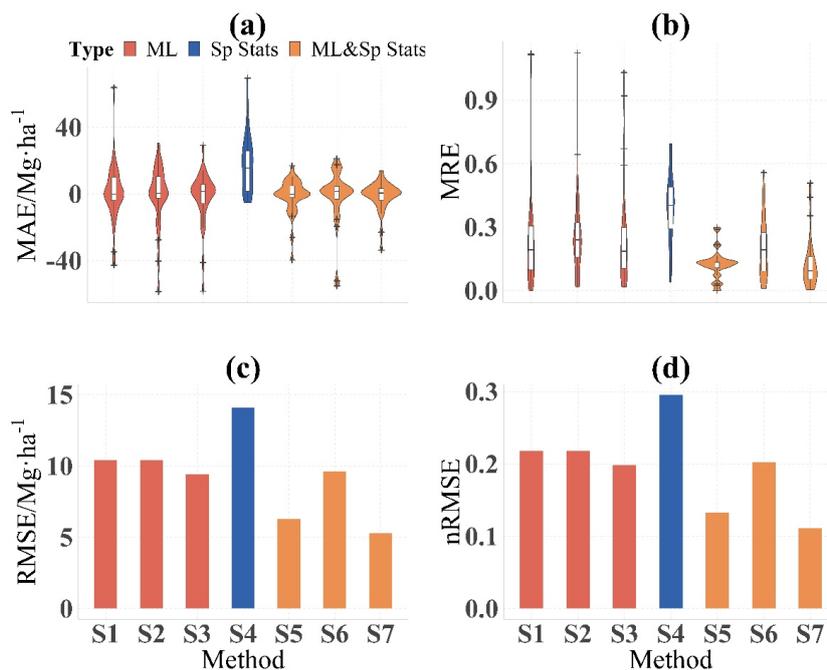
371 3.2.3 Spatial heterogeneity test

372 As shown in Table 1, the true AGBs of the sample plots were divided into three strata by using *k*-means
373 clustering. We then ran the GeogDetector model and obtained a *q* value of 0.87 and a *p* value less
374 than 0.01. These results indicate that the within-layer variances were far less than the sum of variances
375 among different strata. The results also suggest that the reference AGBs of the 30 sample plots were
376 associated with obvious spatially stratified heterogeneity.

377 3.3 Performance of plot-level models

378 We developed seven models for estimating AGB: three machine learning models (SVM, RBF-ANN,
379 and RF), one spatial statistical model (P-BSHADE), and three combined models that integrated each
380 machine learning method with the spatial statistical method (SVM & P-BSHADE, RBF-ANN &
381 P-BSHADE, and RF & P-BSHADE). Furthermore, we used the leave-one-out cross-validation method
382 to split the datasets and evaluated the prediction performance of these seven methods based on the
383 indicators MAE [Fig. 5(a)], MRE [Fig. 5(b)], RMSE [Fig. 5(c)], and nRMSE [Fig. 5(d)].

384



385

386 Figure 5. Prediction performance of the seven different models. (a) MAE and (b) MRE are presented as
387 boxplots for each prediction method, with the median (black horizontal line in the box), inter-quartile
388 range (25%–75% in the box), the range 5%–95% (whiskers), and outliers (asterisks) labeled (S1=SVM,
389 S2=RBF-ANN, S3=RF, S4=P-BSHDE, S5=SVM & P-BSHDE, S6=RBF-ANN & P-BSHDE, S7=RF
390 & P-BSHDE, ML=machine learning, Sp Stats=Spatial statistics). Histogram distributions of RMSE and
391 nRMSE for each prediction method are presented in panels (c) and (d), respectively.

392

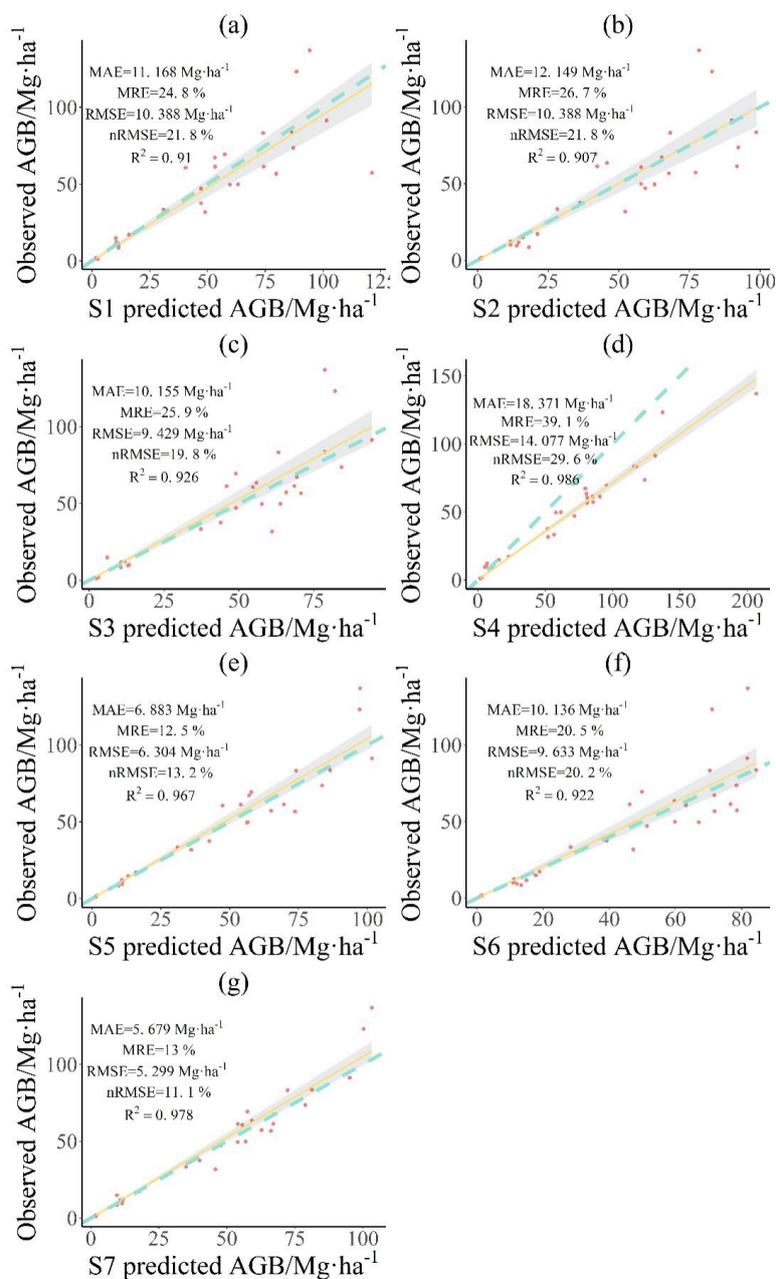
393 The forest AGB estimates obtained by the three machine learning methods were significantly more
394 accurate than those obtained by the spatial statistical method. The performance indicators for
395 P-BSHADE were MAE=18.37 Mg·ha⁻¹, MRE=39.13%, RMSE=14.08 Mg·ha⁻¹, and nRMSE=29.57%,
396 whereas those for the machine learning methods covered the following ranges: MAE 10.16–12.15
397 Mg·ha⁻¹, MRE 24.79%–26.69%, RMSE 9.43–10.39 Mg·ha⁻¹, and nRMSE 19.80%–21.82%.

398 Among the three machine learning methods, the accuracy of RF was highest. The four evaluation
399 indexes (MAE=10.16 Mg·ha⁻¹, MRE=25.93%, RMSE=9.43 Mg·ha⁻¹, and nRMSE=19.80%) were



400 substantially less than those for P-BSHADE and those for the other two machine learning methods
401 (MAE=11.17–12.15 Mg·ha⁻¹, MRE=24.79%–26.69%, RMSE=10.39–10.39 Mg·ha⁻¹, and nRMSE =
402 21.82%). Finally, the combination of machine learning and spatial statistical models produced smaller
403 MAE (5.68–10.14 Mg·ha⁻¹), MRE (12.47%–20.49%), RMSE (5.30–9.63 Mg·ha⁻¹), and nRMSE
404 (11.13%–20.23%) than the single machine learning methods. Of the three combined methods, RF &
405 P-BSHADE produced the highest accuracy with the smallest MAE (5.68 Mg·ha⁻¹), a modest MRE
406 (12.97%), and the smallest RMSE (5.30 Mg·ha⁻¹) and nRMSE (11.13%). In contrast, RBF-ANN &
407 P-BSHADE had the highest MAE (10.14 Mg·ha⁻¹), MRE (20.49%), RMSE (9.63 Mg·ha⁻¹), and
408 nRMSE (20.23%). Compared with the RF model, the RF&P-BSHADE model led to a reduction of the
409 cross-validated prediction error of 43.80%~50.00% (44.08% for MAE, 50.00% for MRE, and 43.80%
410 for RMSE and nRMSE).

411 We also explored the relationship between the observed and predicted AGBs in terms of
412 cross-validation results (Fig. 6). The quantity R² was calculated for the linear regression model applied
413 to the observed and predicted AGBs; R² for every model was greater than 0.9. Although P-BSHADE
414 had the highest R², its distribution of dots in Fig. 6(d) differed quite significantly from the 1:1 line. Of
415 the seven models, the accuracy of RF & P-BSHADE was the highest and the distribution of dots in Fig.
416 6(g) was closest to the 1:1 line. Therefore, we concluded that RF & P-BSHADE was the optimal
417 model.

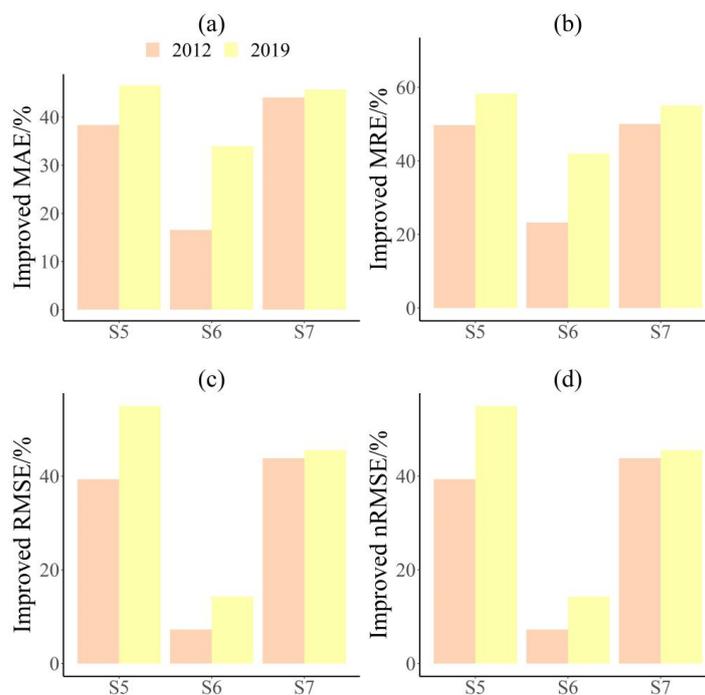


418

419 Figure 6. Comparisons of predicted and observed AGBs for accuracy assessment. Panels (a)–(g) show
420 SVM (S1), RBF-ANN (S2), RF (S3), P-BSHADE (S4), SVM & P-BSHADE (S5), RBF-ANN &
421 P-BSHADE (S6), RF & P-BSHADE (S7), respectively. Green dashed lines represent a 1:1 relationship;
422 dots represent individual sample plots; solid yellow lines indicate trend lines for dots.



423 We compared three machine learning methods with three corresponding combined machine learning
424 and spatial statistical methods by using differences in MAE, MRE, RMSE, and nRMSE during two
425 periods, 2012 and 2019 (Fig. 7). The results suggest that the combined models improved the accuracy
426 of single machine learning models during both years. This suggests that the combined methods are
427 robust.



428

429 Figure 7. The improvement in accuracy assessment indexes of three combined machine learning and
430 spatial statistical methods by comparison with three corresponding machine learning methods. Panels
431 (a)–(d) show the MAE, MRE, RMSE, and nRMSE, respectively; S1-S5 represents RMSE comparison
432 of S5 with S1, S2-S6 represents RMSE comparison of S6 with S2, and S3-S7 represents RMSE
433 comparison of S7 with S3 (S1=SVM, S2=RBF-ANN, S3=RF, S4=P-BSHDE, S5=SVM & P-BSHDE,
434 S6=RBF-ANN & P-BSHDE, S7=RF & P-BSHDE).

435

436 Figure C.3 in section S3 of the Supplementary Material shows the spatial distribution of AGBs



437 predicted by the RF & P-BSHADE model. The predicted AGBs were 7.54–89.93 Mg·ha⁻¹, with an
438 average of 41.21 Mg·ha⁻¹, a median of 43.53 Mg·ha⁻¹, a standard deviation of 18.83 Mg·ha⁻¹, and a
439 coefficient of variation of 45.69%. The total AGB of the Nanjing region (2980 forest patches)
440 estimated by RF & P-BSHADE was 122 812.1 Mg, whereas that estimated by the allometric model
441 was 123 021.5 Mg. The percent difference in total AGB between the two methods was 0.17%.
442 Meanwhile, the AGB MRE between the two methods ranged from 0.04% to 99.8%, with an average of
443 19.93%.

444 **4 Discussion**

445 we developed, evaluated, and compared the accuracy and performance of three different machine
446 learning models [support vector machine (SVM), random forest (RF), and the radial basis function
447 artificial neural network (RBF-ANN)] in this study, which contains one spatial statistics model
448 (P-BSHADE) and three combinations thereof (SVM & P-BSHADE, RF & P-BSHADE, ANN &
449 P-BSHADE) on forest AGB estimates. Those findings suggested that the combined models, especially
450 the RF & P-BSHADE model, could improve the accuracy of plot-level AGB estimates and could reduce
451 the uncertainty of plot-level AGB estimates, owing to its integrated the theoretical advantages of
452 machine learning and spatial statistics.

453 **4.1 Significance of the optimal AGB model at the plot-level**

454 In the past, ecologists converted AGB estimates from forest sample plots into regional AGB estimates
455 by scaling up from the tree-level to the regional scale (Malhi et al., 2004). Plot-level AGB models
456 therefore link tree-level AGB models to regional-scale AGB models. Research by Chen et al. (2015)
457 found that ignoring the uncertainty of plot-level models increased the total uncertainty of pixel-level
458 estimates by 6%. In addition, Marvin et al. (2014) found that the distribution pattern of most AGB is
459 either non-Gaussian, skewed, or multi-modal, especially in tropical and subtropical regions. Different
460 intensity and direction of factors are coupled together, resulting in high heterogeneity and clear
461 nonlinearity in the spatial distribution of forest AGB.

462 Here, we integrated the advantages of machine learning and spatial statistics at the plot level (the key
463 scale linking the tree-level scale to the landscape scale) to construct a plot-level AGB model for a



464 subtropical region. The approach provides a high-precision plot-level AGB model whose estimates can
465 be compared with those obtained from remote sensing, ground observations, and model simulations. It
466 also provides a foundation for making informed forest management decisions (e.g., the method enables
467 quantitative evaluation of carbon emissions from deforestation). Combining the advantages of
468 machine-learning-based quantification of AGB and the complex nonlinear relationships between
469 multiple environmental covariates, in conjunction with the P-BSHADE model, allows the spatial
470 autocorrelation and heterogeneity of multiple environmental covariates to be incorporated into the model.
471 In addition, the sample points are subsequently rectified, thus leading to the best linear unbiased estimate
472 of the target plots.

473 **4.2 Model comparisons**

474 **4.2.1 Machine learning outperforms the spatial statistical model**

475 Regarding the AGB plot-level models, the machine learning methods outperformed the spatial statistical
476 method (P-BSHADE) in terms of prediction accuracy. This may be because machine learning offers an
477 array of supervised learning models capable of relating forest AGB to multi-variables, including forest
478 variables and environmental variables, via complex, potentially nonlinear functional relationships.
479 Machine learning models appear adept at tackling high-dimensional problems, particularly in areas
480 where effective algorithms are lacking and where programs must dynamically adapt to changing
481 conditions (Görgens et al., 2015; Latifi et al., 2010; Stojanova et al., 2010). In addition, the P-BSHADE
482 model yielded negative weights between a small number of plots, which might introduce a slight degree
483 of uncertainty into the results (Xu et al., 2013). Our results were consistent with those of Povak et al.
484 (2014) and Li et al. (2011), who found that a machine learning method (RF) outperformed the spatial
485 statistical method (e.g., Geographically Weighted Regression, Inverse Distance Weighting) in terms of
486 prediction accuracy.

487 **4.2.2 Why a combined model outperforms a single machine learning or spatial statistical model**

488 As expected, the prediction accuracies of the combined methods were higher than those of any single
489 method (either machine learning or spatial statistical). This may due to the advantages of machine
490 learning, which can compensate for the inherent defects of the P-BSHADE model, and vice versa.

491 On the one hand, the P-BSHADE model has its own merits: (1) It takes into account the spatial



492 autocorrelation and spatial heterogeneity of the distribution of the target objects, not only to solve the
493 difference between target objects caused by the different terrain or geographical location but also to
494 solve the problem of strong correlation between target objects with remote geographical locations due
495 to similar terrain condition. (2) The P-BSHADE model calculates the covariance between objects by
496 using a reference sequence between objects (which means the reference AGB data between plots in our
497 study). This method is more reliable because it avoids the second-order stationary hypothesis (i.e.,
498 when using the Kriging algorithm, semi-variograms need this hypothesis), which does not correspond
499 with the actual situation. (3) P-BSHADE regards strongly correlated plots as neighboring plots.
500 However, the P-BSHADE model is also handicapped by the fact that the founding assumption does not
501 conform to reality. The assumption is that estimated AGB is accurate in all sampling plots except the
502 target sampling plot. In other words, the premise behind using only the P-BSHADE model is that the
503 reference AGB data is accurate or strongly correlated with AGB. In reality, the AGB of each sampling
504 plot has a varying degree of uncertainty because it is obtained from the allometric model. Since the
505 P-BSHADE model combined with machine learning uses the results optimized by machine learning as
506 the reference series, it further improves the accuracy of AGB mapping.

507 Machine learning also has its advantages and disadvantages. As we described in the previous section
508 (4.2.2), machine learning has the advantage of being able to handle complex, potentially nonlinear
509 relationships between forest AGB and other variables. However, the initial samples of machine
510 learning are randomly selected, which may lead to differences in the results of each operation of the
511 model. In addition, machine learning uses the average value of all regression trees in the calculation,
512 which may result in overestimating the lower value and underestimating the higher value. As opposed
513 to machine learning, the P-BSHADE model takes into account the spatial autocorrelation and spatial
514 heterogeneity of forest AGB and of environmental covariates, and the bias of the observed values of the
515 sampling plots, which corresponds more to actual situations. A combined model takes the result of
516 machine learning as the reference series of P-BSHADE, so that the fitting process of the combined
517 model takes spatial relationships more into account than is the case for the single machine learning
518 model. The end result is improved accuracy.

519 Machine learning models or the P-BSHADE model have been used to model the uncertainty of
520 temperature measurements obtained by weather stations (Fassnacht et al., 2014; Paul et al., 2016; Xu et



521 al., 2013). However, the methods used in these studies were adopted independently. Conversely, the
522 combination of machine learning and spatial statistics can improve the prediction accuracy of AGB
523 maps, which in turn can be used as criteria for improving the accuracy of LiDAR remote-sensing
524 technology and the results of ecological process models. Eventually, these improvements can promote
525 process-oriented projects that require dynamic AGB predictions for large-scale forests in different
526 forest management scenarios.

527 In addition, we compared the prediction accuracy of AGB mapping obtained by the combined spatial
528 statistical and machine learning models with that reported by recent studies using AGB plot-level
529 models. In the current literature on remote-sensing estimation of forest AGB, nRMSE, RMSE, and R^2
530 were commonly used as indexes for evaluating the prediction performance of models affected by
531 research sample size, data type, and forecasting methods (Fassnacht et al., 2014). In contrast, the
532 present study used four conventional indexes for evaluating prediction performance: nRMSE, RMSE,
533 MAE, and MRE. The criterion for model selection is to choose indexes summarized from sample
534 prediction (such as nRMSE), rather than choosing the goodness-of-fit R^2 (Babcock et al., 2015). Based
535 on calculated nRMSE indexes, the AGB prediction accuracy of the combined RF & P-BSHADE model
536 (11.13%) was higher than that obtained by Babcock et al. (2015) (33.91%) in Colorado, USA. In that
537 study, the authors used a combination of airborne LiDAR, a forest inventory database, and a Bayesian
538 spatial hierarchical framework model and introduced spatial random effects to compensate for the
539 residual spatial dependence and non-stationary model covariates. The AGB prediction accuracy of the
540 method developed in the current work was also greater than that obtained by Ioki et al. (2014)
541 (nRMSE=26%) in northern Borneo using a stepwise linear regression model with airborne LiDAR and
542 a ground survey. Furthermore, it exceeded the accuracy obtained by Hansen et al. (2015) in the tropical
543 submontane rain forest (34.4%) using fusion maps of multi-source databases combined with multiple
544 regression analysis. Our prediction accuracy is close to that obtained by Kim et al. (2016) (9.2%) who
545 studied an intact tropical rain forest by using a voxel-based method based on airborne LiDAR in
546 conjunction with field monitoring in Brunei. Our combined methods produce very small RMSE for the
547 prediction accuracy of AGB, which we attribute to the following reasons: (1) The true AGBs of the 30
548 sample plots were calculated from each tree by using an allometric model constructed from the 90 most
549 accurate harvested trees. There were no differences in the range of true values. (2) Machine learning



550 methods were used to quantify the complex nonlinear relationship between AGB and multiple
551 environmental covariates. (3) We applied a spatial statistical method based on the hypothesis of spatial
552 heterogeneity. Although the nRMSE index was calculated by different studies using different datasets
553 and prediction methods in different locations, most studies agreed that nRMSE was the most
554 commonly used indicator for measuring the AGB prediction errors of plot-level models and for
555 calculating the true AGB of forest sample plots. In contrast to other studies, our work reflects not only
556 a focus on subtropical forests but also the methodological differences in uncertainty mitigation,
557 especially in terms of comprehensively addressing the sources of uncertainty caused by multiple spatial
558 and environmental covariates.

559 **4.2.3 Why RF & P-BSHADE method outperforms other combined methods**

560 The three combined machine learning and spatial statistical methods produced more accurate AGB
561 predictions than any individual method. The accuracy of the RF & P-BSHADE and SVM &
562 P-BSHADE methods were significantly higher than that of the individual methods, but the RBF-ANN
563 & P-BSHADE method was only slightly higher. The accuracies of the combined methods depend on
564 the accuracy of the reference series (machine learning predicted result) (Xu et al., 2013). In other words,
565 the higher the accuracy of the predicted machine learning results, the higher the accuracy of the
566 combined method. Therefore, the different improvements offered by the three combined methods may
567 be attributed to the following two mechanisms: (1) the RF and SVM models are easier to use and
568 optimize than RBF-ANN (Raczko and Zagajewski, 2017). RBF-ANN is sensitive to hyper-parameters
569 and usually requires optimized parameters to obtain better fitting results. However, in the present study,
570 we used no optimized algorithms, such as genetic algorithms, to obtain parameters in the machine
571 learning model. Furthermore, the number of training samples determines the number of nodes in the
572 hidden layer of the RBF-ANN model, and the number of nodes significantly affects the prediction
573 accuracy. With only 30 training samples used in this study, the combined approach may have been
574 unable to strongly improve prediction accuracy. (2) RBF-ANN is more suitable for nonlinear stochastic
575 dynamic systems (Elanayar and Shin, 1994), whereas the relationship between AGB and environmental
576 covariates in this study is likely a monotonically increasing function.



577 **4.3 Comparing upscaling of RF&P-BSHADE with allometric model**

578 We used FMPI data to upscale the optimal plot-level AGB model from plot level to region scale.
579 Because the allometric model offers a fast and simple calculation method, it has been used in many
580 studies as the basis for determining the benchmark map. Nevertheless, spatial heterogeneity caused by
581 multiple environmental covariates is not considered in the allometric model because potential errors in
582 the AGB estimate may be propagated and affect the accuracy of the regional AGB map. Although we
583 regarded the FMPI patches as homogeneous study units in the present study, the area of the forest
584 patches is significantly larger than that of the sample plots. Upscaling results will thus have large
585 uncertainties (see Figs. C.4, S3 of Supplementary Material) (Chen et al., 2015). The current study finds
586 that the relative percent difference in total AGB between RF & P-BSHADE and the allometric model
587 was 0.17%. Meanwhile, the relative error (RE) in AGB between the two models ranged from 0.04% to
588 99.8% with a MRE of 19.93%. This suggests that the two methods are similar in terms of overall
589 estimates of AGB but that the local spatial distribution of AGB differs. Differences in AGB spatial
590 distribution have been reported in many studies of AGB maps. Babcock et al. (2015) asserted that the
591 main reasons for the differences in the spatial distribution of AGB maps between different methods
592 include the following: (1) The structural framework of different research methods and schemes cannot
593 truly reflect actual forest growth. (2) The model is usually a simplification of an ecological process and
594 ignores spatial heterogeneity at the regional scale. (3) The model does not consider the influence of
595 multiple environmental covariates (vegetation, topography, and others) on forest growth in the region.

596 **5 Conclusions**

597 This paper proposes a method to integrate the advantages of machine learning and spatial statistics,
598 different datasets, and multiple environmental covariates to improve the accuracy of plot-level
599 AGB-estimation models. In this study, we explored the prediction performance of different AGB
600 models and found that the model that combines the Random Forest and P-BSHADE models
601 substantially improved estimates of forest AGB. Although data from the sample plots and harvested
602 trees were collected only from *Eucalyptus* forests in the Nanjing region of China, the proposed model
603 and the associated results can provide references for AGB mapping in other countries and in different
604 types of tropical forests.



605 **Data availability.**

606 All data are included in the paper and Supplement.

607 **Author Contributions**

608 Y.R. designed the study. X.Z. carried out the data collection. S.D. carried out the analyses and
609 visualized the data. X.Z. and S.D. wrote the manuscript with help from Y.R. L.G., C.X., S.Z., Q.C., and
610 X.W. provided technical advice and guidance throughout the project implementation and paper-writing
611 stages. S.D. and X.Z. contributed equally to this work.

612 **Competing interests.**

613 The authors declare that they have no conflict of interest.

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629 **References**

- 630 Andersen, H.-E., Reutebuch, S. E., McGaughey, R. J., d'Oliveira, M. V. N., and Keller, M.:
631 Monitoring selective logging in western Amazonia with repeat lidar flights, *Remote Sensing of*
632 *Environment*, 151, 157-165, 10.1016/j.rse.2013.08.049, 2014.
- 633 Babcock, C., Finley, A. O., Bradford, J. B., Kolka, R., Birdsey, R., and Ryan, M. G.: LiDAR based
634 prediction of forest biomass using hierarchical models with spatially varying coefficients, *Remote*
635 *Sensing of Environment*, 169, 113-127, 2015.
- 636 Benitez, F. L., Anderson, L. O., and Formaggio, A. R.: Evaluation of geostatistical techniques to
637 estimate the spatial distribution of aboveground biomass in the Amazon rainforest using
638 high-resolution remote sensing data, *Acta Amazonica*, 46, 151-160, 2016.
- 639 Breiman, L.: Random forests, *Machine Learning*, 45, 5-32, 2001.
- 640 Bustamante, M. M., Roitman, I., Aide, T. M., Alencar, A., Anderson, L., Aragão, L., Asner, G. P.,
641 Barlow, J., Berenguer, E., and Chambers, J.: Towards an integrated monitoring framework to assess the
642 effects of tropical forest degradation and recovery on carbon stocks and biodiversity, *Global Change*
643 *Biology*, 22, 92-109, 2016.
- 644 Chave, J., Réjou-Méchain, M., Búrquez, A., Chidumayo, E., Colgan, M. S., Delitti, W. B. C.,
645 Duque, A., Eid, T., Fearnside, P. M., Goodman, R. C., Henry, M., Martínez-Yrizar, A., Mugasha, W. A.,
646 Muller-Landau, H. C., Mencuccini, M., Nelson, B. W., Ngomanda, A., Nogueira, E. M.,
647 Ortiz-Malavassi, E., Péliissier, R., Ploton, P., Ryan, C. M., Saldarriaga, J. G., and Vieilledent, G.:
648 Improved allometric models to estimate the aboveground biomass of tropical trees, *Global Change*
649 *Biology*, 20, 3177-3190, 10.1111/gcb.12629, 2014.
- 650 Chen, Q.: Modeling aboveground tree woody biomass using national-scale allometric methods
651 and airborne lidar, *ISPRS Journal of Photogrammetry and Remote Sensing*, 106, 95-106,
652 10.1016/j.isprsjprs.2015.05.007, 2015.
- 653 Chen, Q., Laurin, G. V., and Valentini, R.: Uncertainty of remotely sensed aboveground biomass
654 over an African tropical forest: Propagating errors from trees to plots to pixels, *Remote Sensing of*
655 *Environment*, 160, 134-143, 2015.
- 656 Cliff, A., and Ord, V. J.: *Spatial processes: model and applications*, Pion Ltd, London, 1981.
- 657 Conti, G., Gorné, L. D., Zeballos, S. R., Lipoma, M. L., Gatica, G., Kowaljow, E.,
658 Whitworth-Hulse, J. I., Cuchietti, A., Poca, M., Pestoni, S., and Fernandes, P. M.: Developing
659 allometric models to predict the individual aboveground biomass of shrubs worldwide, *Global Ecology*
660 *and Biogeography*, 28, 961-975, 10.1111/geb.12907, 2019.
- 661 Djomo, A. N., Picard, N., Fayolle, A., Henry, M., Ngomanda, A., Ploton, P., McLellan, J.,
662 Saborowski, J., Adamou, I., and Lejeune, P.: Tree allometry for estimation of carbon stocks in African
663 tropical forests, *Forestry: An International Journal of Forest Research*, 89, 446-455,
664 10.1093/forestry/cpw025, 2016.
- 665 Djomo, A. N., and Chimi, C. D.: Tree allometric equations for estimation of above, below and



- 666 total biomass in a tropical moist forest: Case study with application to remote sensing, *Forest Ecology*
667 *and Management*, 391, 184-193, <https://doi.org/10.1016/j.foreco.2017.02.022>, 2017.
- 668 Hearst, M. A., Dumais, S. T., Osman, E., Platt, J., & Scholkopf, B.: Support vector machines,
669 *IEEE Intelligent Systems*, 13, 18-28, 1998.
- 670 Drucker, H., Burges, C. J. C., Kaufman, L., Smola, A., and Vapnik, V.: Support vector regression
671 machines, *Proceedings of the 9th International Conference on Neural Information Processing Systems*,
672 Denver, Colorado, 1996.
- 673 Du, H., Zhou, G., Fan, W., Ge, H., Xu, X., Shi, Y., and Fan, W.: Spatial heterogeneity and carbon
674 contribution of aboveground biomass of moso bamboo by using geostatistical theory, *Plant Ecology*,
675 207, 131-139, 2010.
- 676 Elanayar, V. T. S., and Shin, Y. C.: Radial basis function neural network for approximation and
677 estimation of nonlinear stochastic dynamic systems, *IEEE Transactions on Neural Networks*, 5,
678 594-603, 1994.
- 679 Fasnacht, F. E., Hartig, F., Latifi, H., Berger, C., Hernández, J., Corvalán, P., and Koch, B.:
680 Importance of sample size, data type and prediction method for remote sensing-based estimations of
681 aboveground forest biomass, *Remote Sensing of Environment*, 154, 102-114, 2014.
- 682 Fayolle, A., Doucet, J.-L., Gillet, J.-F., Bourland, N., and Lejeune, P.: Tree allometry in Central
683 Africa: Testing the validity of pantropical multi-species allometric equations for estimating biomass
684 and carbon stocks, *Forest Ecology and Management*, 305, 29-37,
685 <https://doi.org/10.1016/j.foreco.2013.05.036>, 2013.
- 686 Frey, U. J., Klein, M., and Deissenroth, M.: Modelling complex investment decisions in Germany
687 for renewables with different machine learning algorithms, *Environmental Modelling & Software*, 118,
688 61-75, <https://doi.org/10.1016/j.envsoft.2019.03.006>, 2019.
- 689 Gao, L., and Hailu, A.: Ranking management strategies with complex outcomes: An AHP-fuzzy
690 evaluation of recreational fishing using an integrated agent-based model of a coral reef ecosystem,
691 *Environmental Modelling & Software*, 31, 3-18, <https://doi.org/10.1016/j.envsoft.2011.12.002>, 2012.
- 692 Gao, L., Bryan, B. A., Nolan, M., Connor, J. D., Song, X., and Zhao, G.: Robust global sensitivity
693 analysis under deep uncertainty via scenario analysis, *Environmental modelling & software*, 76,
694 154-166, 2016.
- 695 Gleason, C. J., and Im, J.: Forest biomass estimation from airborne LiDAR data using machine
696 learning approaches, *Remote Sensing of Environment*, 125, 80-91, 2012.
- 697 Görgens, E. B., Montaghi, A., and Rodriguez, L. C. E.: A performance comparison of machine
698 learning methods to estimate the fast-growing forest plantation yield based on laser scanning metrics,
699 *Computers and Electronics in Agriculture*, 116, 221-227, <https://doi.org/10.1016/j.compag.2015.07.004>,
700 2015.
- 701 Hansen, H. E., Gobakken, T., Bollandsås, M. O., Zahabu, E., and Næsset, E.: Modeling



- 702 Aboveground Biomass in Dense Tropical Submontane Rainforest Using Airborne Laser Scanner Data,
703 Remote Sensing, 7, 10.3390/rs70100788, 2015.
- 704 He, C., Tian, J., Shi, P., and Hu, D.: Simulation of the spatial stress due to urban expansion on the
705 wetlands in Beijing, China using a GIS-based assessment model, Landscape and Urban Planning, 101,
706 269-277, <https://doi.org/10.1016/j.landurbplan.2011.02.032>, 2011.
- 707 Houghton, R. A., Hall, F., and Goetz, S. J.: Importance of biomass in the global carbon cycle,
708 Journal of Geophysical Research Biogeosciences, 114, G00E03, 2009.
- 709 Hu, M. G., Wang, J. F., Zhao, Y., and Jia, L.: A B-SHADE based best linear unbiased estimation
710 tool for biased samples, Environmental Modelling & Software, 48, 93-97, 2013.
- 711 Huang, H., Liu, C., Wang, X., Zhou, X., and Gong, P.: Integration of multi-resource remotely
712 sensed data and allometric models for forest aboveground biomass estimation in China, Remote
713 Sensing of Environment, 221, 225-234, <https://doi.org/10.1016/j.rse.2018.11.017>, 2019.
- 714 Ioki, K., Tsuyuki, S., Hirata, Y., Phua, M.-H., Wong, W. V. C., Ling, Z.-Y., Saito, H., and Takao,
715 G.: Estimating above-ground biomass of tropical rainforest of different degradation levels in Northern
716 Borneo using airborne LiDAR, Forest Ecology and Management, 328, 335-341,
717 <https://doi.org/10.1016/j.foreco.2014.06.003>, 2014.
- 718 Jachowski, N. R. A., Quak, M. S. Y., Friess, D. A., Duangnamon, D., Webb, E. L., and Ziegler, A.
719 D.: Mangrove biomass estimation in Southwest Thailand using machine learning, Applied Geography,
720 45, 311-321, <https://doi.org/10.1016/j.apgeog.2013.09.024>, 2013.
- 721 Kim, E., Lee, W.-K., Yoon, M., Lee, J.-Y., Son, Y., and Abu Salim, K.: Estimation of Voxel-Based
722 Above-Ground Biomass Using Airborne LiDAR Data in an Intact Tropical Rain Forest, Brunei, Forests,
723 7, 10.3390/f7110259, 2016.
- 724 Latifi, H., Nothdurft, A., and Koch, B.: Non-parametric prediction and mapping of standing timber
725 volume and biomass in a temperate forest: application of multiple optical/LiDAR-derived predictors,
726 Forestry: An International Journal of Forest Research, 83, 395-407, 10.1093/forestry/cpq022, 2010.
- 727 Li, J., Heap, A. D., Potter, A., and Daniell, J. J.: Application of machine learning methods to
728 spatial interpolation of environmental variables, Environmental Modelling & Software, 26, 1647-1659,
729 <https://doi.org/10.1016/j.envsoft.2011.07.004>, 2011.
- 730 Lu, Z., Lin, F., and Ying, H.: DESIGN OF DECISION TREE VIA KERNELIZED
731 HIERARCHICAL CLUSTERING FOR MULTICLASS SUPPORT VECTOR MACHINES,
732 Cybernetics and Systems, 38, 187-202, 10.1080/01969720601139058, 2007.
- 733 Malhi, Y., Phillips, O. L., Chave, J., Condit, R., Aguilar, S., Hernandez, A., Lao, S., and Perez, R.:
734 Error propagation and scaling for tropical forest biomass estimates, Philosophical Transactions of the
735 Royal Society of London. Series B: Biological Sciences, 359, 409-420, 10.1098/rstb.2003.1425, 2004.
- 736 Marvin, D. C., Asner, G. P., Knapp, D. E., Anderson, C. B., Martin, R. E., Sinca, F., and Tupayachi,
737 R.: Amazonian landscapes and the bias in field studies of forest structure and biomass, Proceedings of
738 the National Academy of Sciences of the United States of America, 111, 5224-5232, 2014.



- 739 McRoberts, R. E., Chen, Q., Domke, G. M., Ståhl, G., Saarela, S., and Westfall, J. A.: Hybrid
740 estimators for mean aboveground carbon per unit area, *Forest Ecology and Management*, 378, 44-56,
741 10.1016/j.foreco.2016.07.007, 2016.
- 742 McRoberts, R. E., Chen, Q., Gormanson, D. D., and Walters, B. F.: The shelf-life of airborne laser
743 scanning data for enhancing forest inventory inferences, *Remote Sensing of Environment*, 206,
744 254-259, 10.1016/j.rse.2017.12.017, 2018.
- 745 Mendoza-Ponce, A., and Galicia, L.: Aboveground and belowground biomass and carbon pools in
746 highland temperate forest landscape in Central Mexico, *Forestry: An International Journal of Forest
747 Research*, 83, 497-506, 10.1093/forestry/cpq032, 2010.
- 748 Morel, A. C., Fisher, J. B., and Malhi, Y.: Evaluating the potential to monitor aboveground
749 biomass in forest and oil palm in Sabah, Malaysia, for 2000–2008 with Landsat ETM+ and
750 ALOS-PALSAR, *International Journal of Remote Sensing*, 33, 3614-3639, 2012.
- 751 Paul, K. I., Roxburgh, S. H., Chave, J., England, J. R., Zerihun, A., Specht, A., Lewis, T., Bennett,
752 L. T., Baker, T. G., Adams, M. A., Huxtable, D., Montagu, K. D., Falster, D. S., Feller, M., Sochacki, S.,
753 Ritson, P., Bastin, G., Bartle, J., Wildy, D., Hobbs, T., Larmour, J., Waterworth, R., Stewart, H. T.,
754 Jonson, J., Forrester, D. I., Applegate, G., Mendham, D., Bradford, M., O'Grady, A., Green, D.,
755 Sudmeyer, R., Rance, S. J., Turner, J., Barton, C., Wenk, E. H., Grove, T., Attiwill, P. M., Pinkard, E.,
756 Butler, D., Brooksbank, K., Spencer, B., Snowdon, P., O'Brien, N., Battaglia, M., Cameron, D. M.,
757 Hamilton, S., McArthur, G., and Sinclair, J.: Testing the generality of above-ground biomass allometry
758 across plant functional types at the continent scale, *Global Change Biology*, 22, 2106-2124,
759 10.1111/gcb.13201, 2016.
- 760 Picard, N., Rutishauser, E., Ploton, P., Ngomanda, A., and Henry, M.: Should tree biomass
761 allometry be restricted to power models?, *Forest Ecology and Management*, 353, 156-163,
762 <https://doi.org/10.1016/j.foreco.2015.05.035>, 2015.
- 763 Povak, N. A., Hessburg, P. F., McDonnell, T. C., Reynolds, K. M., Sullivan, T. J., Salter, R. B., and
764 Cosby, B. J.: Machine learning and linear regression models to predict catchment-level base cation
765 weathering rates across the southern Appalachian Mountain region, USA, *Water Resources Research*,
766 50, 2798-2814, 10.1002/2013WR014203, 2014.
- 767 Propastin, P.: Modifying geographically weighted regression for estimating aboveground biomass
768 in tropical rainforests by multispectral remote sensing data, *International Journal of Applied Earth
769 Observation and Geoinformation*, 18, 82-90, 2012.
- 770 Qiu, Q., Yun, G., Zuo, S., Yan, J., Hua, L., Ren, Y., Tang, J., Li, Y., and Chen, Q.: Variations in the
771 biomass of Eucalyptus plantations at a regional scale in Southern China, *Journal of Forestry Research*,
772 29, 1263-1276, 10.1007/s11676-017-0534-0, 2018.
- 773 Raczko, E., and Zagajewski, B.: Comparison of support vector machine, random forest and neural
774 network classifiers for tree species classification on airborne hyperspectral APEX images, *European
775 Journal of Remote Sensing*, 50, 144-154, 10.1080/22797254.2017.1299557, 2017.
- 776 Rangel, T. F., and Bini, L. M.: SAM: A comprehensive application for Spatial Analysis in



- 777 Macroecology, *Ecography*, 33, 46-50, 2010.
- 778 Ren, Y., Zhang, C., Zuo, S., and Li, Z.: Scaling up of biomass simulation for Eucalyptus
779 plantations based on landsense ecology, *International Journal of Sustainable Development & World*
780 *Ecology*, 24, 135-148, 2017.
- 781 Rosenberg, M. S., and Anderson, C. D.: PASSaGE: Pattern Analysis, Spatial Statistics and
782 Geographic Exegesis. Version 2, *Methods in Ecology and Evolution*, 2, 229-232, 2011.
- 783 Saatchi, S. S., Harris, N. L., Brown, S., Lefsky, M., Mitchard, E. T. A., Salas, W., Zutta, B. R.,
784 Buermann, W., Lewis, S. L., Hagen, S., Petrova, S., White, L., Silman, M., and Morel, A.: Benchmark
785 map of forest carbon stocks in tropical regions across three continents, *Proceedings of the National*
786 *Academy of Sciences of the United States of America*, 108, 9899-9904, 10.1073/pnas.1019576108,
787 2011.
- 788 Schabenberger, O., and Gotway, C. A.: *Statistical methods for spatial data analysis*, Chapman &
789 Hall/CRC, Boca Raton, 2005.
- 790 Stojanova, D., Panov, P., Gjorgjioski, V., Kobler, A., and Džeroski, S.: Estimating vegetation
791 height and canopy cover from remotely sensed data with machine learning, *Ecological Informatics*, 5,
792 256-266, <https://doi.org/10.1016/j.ecoinf.2010.03.004>, 2010.
- 793 Van der Laan, C., Verweij, P. A., Quiñones, M. J., and Faaij, A. P.: Analysis of biophysical and
794 anthropogenic variables and their relation to the regional spatial variation of aboveground biomass
795 illustrated for North and East Kalimantan, Borneo, *Carbon Balance and Management*, 9, 8, 2014.
- 796 Viana, H., Aranha, J., Lopes, D., and Cohen, W. B.: Estimation of crown biomass of *Pinus pinaster*
797 stands and shrubland above-ground biomass using forest inventory data, remotely sensed imagery and
798 spatial prediction models, *Ecological Modelling*, 226, 22-35, 2012.
- 799 Wang, J.-F., Zhang, T.-L., and Fu, B.-J.: A measure of spatial stratified heterogeneity, *Ecological*
800 *Indicators*, 67, 250-256, <https://doi.org/10.1016/j.ecolind.2016.02.052>, 2016.
- 801 Wang, J. F., Li, X. H., Christakos, G., Liao, Y. L., Zhang, T., Gu, X., and Zheng, X. Y.:
802 Geographical Detectors-Based Health Risk Assessment and its Application in the Neural Tube Defects
803 Study of the Heshun Region, China, *International Journal of Geographical Information Science*, 24,
804 107-127, 2010.
- 805 Xia, C. L., Xiu, J.: RBF ANN Nonlinear Prediction Model Based Adaptive PID Control of
806 Switched Reluctance Motor, *Proceedings of the CSEE*, 27, 626-635, 10.1007/11893295_69, 2007.
- 807 Xu, C. D., Wang, J. F., Hu, M. G., and Li, Q. X.: Interpolation of Missing Temperature Data at
808 Meteorological Stations Using P-BSHADE*, *Journal of Climate*, 26, 7452-7463, 2013.
- 809 Zhang, J., Huang, S., Hogg, E. H., Lieffers, V., Qin, Y., and He, F.: Estimating spatial variation in
810 Alberta forest biomass from a combination of forest inventory and remote sensing data, *Biogeosciences*,
811 11, 2793-2808, 10.5194/bg-11-2793-2014, 2014.
- 812 Zheng, D., Rademacher, J., Chen, J., Crow, T., Bresee, M., Moine, J. L., and Ryu, S. R.:



- 813 Estimating aboveground biomass using Landsat 7 ETM+ data across a managed landscape in northern
814 Wisconsin, USA, Remote Sensing of Environment, 93, 402-411, 2004.