



Acetate turnover and methanogenic pathways in Amazonian lake sediments

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1 **Abstract**

2 Lake sediments in Amazonia are a significant source of CH₄, a potential greenhouse gas.
3 Previous studies of sediments using ¹³C analysis found that the contribution of
4 hydrogenotrophic versus acetoclastic methanogenesis to CH₄ production was relatively high.
5 Here, we determined the methanogenic pathway in the same sediments (n = 6) by applying
6 [¹⁴C]bicarbonate or [2-¹⁴C]acetate, and confirmed the high relative contribution (50-80%) of
7 hydrogenotrophic methanogenesis. The respiratory index (RI) of [2-¹⁴C]acetate, which is
8 ¹⁴CO₂ relative to ¹⁴CH₄ + ¹⁴CO₂, divided the sediments into two categories, i.e., those with an
9 RI < 0.2 being consistent with the operation of acetoclastic methanogenesis, and those with an
10 RI > 0.4 showing that a large percentage of the acetate-methyl was oxidized to CO₂ rather
11 than reduced to CH₄. Hence, part of the acetate was probably converted to CO₂ plus H₂ via
12 syntrophic oxidation, thus enhancing hydrogenotrophic methanogenesis. This happened
13 despite the presence of potentially acetoclastic *Methanosaetaceae* in all the sediments.
14 Alternatively, acetate may have been oxidized with a constituent of the sediment organic
15 matter (humic acid) serving as oxidant. Indeed, apparent acetate turnover rates were larger
16 than CH₄ production rates except in those sediments with a R < 0.2. Our study demonstrates
17 that CH₄ production in Amazonian lake sediments was not simply caused by a combination of
18 hydrogenotrophic and acetoclastic methanogenesis, but probably involved additional acetate
19 turnover.
20



21 **1. Introduction**

22 Acetate is an important intermediate in the anoxic degradation of organic matter and is
23 produced by fermentation processes and chemolithotrophic homoacetogenesis (Conrad, 2019;
24 Conrad et al., 2014; Ye et al., 2014). The contribution of these two processes to acetate
25 production is difficult to determine, but seems to be quite different for different environments
26 (Fu et al., 2018; Hädrich et al., 2012; Heuer et al., 2010). The degradation of acetate requires
27 a suitable oxidant such as oxygen, nitrate, ferric iron or sulfate. If such oxidants are not or no
28 longer available, such as in many freshwater environments (e.g., paddy fields, lake sediments,
29 peat) acetate sometimes accumulates until suitable electron acceptors become again available.
30 Temporal accumulation and subsequent oxidative consumption has for example been
31 observed in peatlands during increase and decrease, respectively, of the water table
32 (Duddleston et al., 2002). Normally, however, acetate degradation in the absence of inorganic
33 electron acceptors is accomplished by aceticlastic methanogenesis (Conrad, 2019). If
34 aceticlastic methanogenesis is operative, the methyl group of the acetate is converted to CH₄.

35 If methanogenesis is the exclusive final step in the anaerobic degradation of organic
36 matter, polysaccharides (one of the most important compounds from primary production) will
37 be dismutated to equal amounts of CH₄ and CO₂. Furthermore, acetate usually accounts for
38 more than two third of total methane production, especially if polysaccharides are the
39 predominant degradable organic matter (Conrad, 2019). However, CO₂ has often been found
40 to be the main product in many anoxic environments despite the absence of inorganic electron
41 acceptors (O₂, nitrate, ferric iron, sulfate) (Keller et al., 2009; Yavitt and Seidmann-Zager,
42 2006). Such results have been explained by the assumption that organic substances (e.g.
43 humic acids) may also serve as electron acceptors (Gao et al., 2019; Keller et al., 2009;
44 Klüpfel et al., 2014). Organic electron acceptors also allow the oxidation of acetate (Coates et
45 al., 1998; Lovley et al., 1996). The role of organic electron acceptors during anaerobic
46 degradation of organic matter is potentially important, but still not well known (Corbett et al.,
47 2013)

48 There are also many reports that methane production in lake sediments is dominated by
49 hydrogenotrophic rather than aceticlastic methanogenesis (Conrad et al., 2011; Ji et al., 2016).



50 Such observations were explained (1) by incomplete degradation of organic matter producing
51 predominantly H_2 and CO_2 without concomitant acetate production (Conrad et al., 2010;
52 Hodgkins et al., 2014; Liu et al., 2017), (2) by acetate oxidation coupled to the reduction of
53 organic substances (see above), or (3) by syntrophic acetate oxidation coupled with
54 hydrogenotrophic methanogenesis (Lee and Zinder, 1988; Vavilin et al., 2017). If syntrophic
55 acetate oxidation is operative, the methyl group of the acetate is converted to CO_2 , similarly
56 as found during acetate oxidation with external inorganic or organic electron acceptors.
57 However, syntrophic acetate oxidation does not require a chemical compound (other than H^+)
58 as electron acceptors, since it is the hydrogenotrophic methanogenesis that eventually accepts
59 the electrons released during acetate oxidation.

60 Syntrophic acetate oxidation can replace aceticlastic methanogenesis and thus, has been
61 found when aceticlastic methanogenic archaea were not present in the microbial community
62 of lake sediment (Nüsslein et al., 2001). This may also happen in other anoxic environments
63 when conditions are not suitable for aceticlastic methanogens, e.g., at elevated temperatures
64 (Conrad et al., 2009; Liu and Conrad, 2010; Liu et al., 2018), in the presence of high
65 concentrations of ammonium (Müller et al., 2016; Schnürer et al., 1999; Zhang et al., 2014),
66 or phosphate (Conrad et al., 2000). However, syntrophic acetate oxidation has also been found
67 in lake sediments that contained populations of putatively aceticlastic methanogens (Vavilin
68 et al., 2017). It is presently unknown under which conditions syntrophic acetate oxidizers can
69 successfully compete with aceticlastic methanogens and co-occur with acetate oxidation that
70 is coupled to the reduction of organic substances.

71 As a further step in understanding the ecology of syntrophic acetate oxidizers versus
72 aceticlastic methanogens, we attempted to document their coexistence by studying lake
73 sediments, which had been reported containing 16S rRNA genes of putatively aceticlastic
74 *Methanosaetaceae* (*Methanotrichaceae* (Oren, 2014)) (Ji et al., 2016). We used these
75 sediments and measured the fractions of hydrogenotrophic methanogenesis and of the methyl
76 group of acetate being oxidized to CO_2 rather than reduced to CH_4 , and compared the
77 turnover of acetate to the production rate of CH_4 .

78



79 **2. Materials and Methods**

80 The sediment samples were obtained from floodplain lakes in the Amazon region and have
81 already been used for a study on structure and function of methanogenic microbial
82 communities (Ji et al., 2016). In particular, these sediment have been assayed for the
83 percentage of hydrogenotrophic methanogenesis using values of $\delta^{13}\text{C}$ of CH_4 , CO_2 and
84 acetate-methyl, and for the percentage contribution of putatively aceticlastic methanogens to
85 the total archaeal community (Ji et al., 2016). Here, we used six of these sediments for
86 incubation experiments with radioactive tracers. These sediments are identical to those listed
87 in our previous publication (Ji et al., 2016). The identity of the lake sediments and the
88 percentage content of putatively aceticlastic methanogens is summarized in Table 1.

89 The experiments were carried out at the same time as those in our previous publication and
90 were basically using the same incubation techniques. Briefly, for determination of CH_4
91 production rates and the fractions of hydrogenotrophic methanogenesis about 10-15 ml of
92 each replicate ($n = 3$) were filled into 27-ml sterile tubes, flushed with N_2 , closed with butyl
93 rubber stoppers, and incubated at 25°C . After preincubation for 12 days (in order to deplete
94 eventually present inorganic oxidants), 0.5 ml of a solution of $\text{NaH}^{14}\text{CO}_3$ (about $1 \mu\text{Ci}$) was
95 added, the tubes flushed again with N_2 , and incubation was continued at 25°C for about 100
96 days. Partial pressures of CH_4 and CO_2 as well as their contents of ^{14}C were measured at
97 different time points after mixing the slurries by heavy manual shaking. The gas partial
98 pressures were measured by gas chromatography with a flame ionization detector (Ji et al.,
99 2016), the radioactivities were analyzed with a radiodetector (RAGA) (Conrad et al., 1989).
100 For determination of acetate turnover, the same conditions were used, except that
101 preincubation was for 25 days, 0.5 ml of a solution of carrier-free $\text{Na}[2\text{-}^{14}\text{C}]\text{acetate}$ (about 2
102 μCi) was added, and incubation was continued for about 8 h. During this time gas samples
103 were repeatedly taken and the radioactivities in CH_4 and CO_2 were analyzed in a gas
104 chromatograph with a radiodetector (RAGA) (Conrad et al., 1989). In the end, the sediment
105 samples were acidified with 1 ml of 1M H_2SO_4 to liberate CO_2 from carbonates, and the
106 radioactivities in CH_4 and CO_2 were analyzed again.



107 The data were used to calculate the fractions of hydrogenotrophic methanogenesis (f_{H_2}),
108 the acetate turnover rate constants (k_{ac}) and the respiratory index (RI) values as described by
109 Schütz et al. (1989). The RI is defined as $RI = \frac{^{14}CO_2}{(^{14}CO_2 + ^{14}CH_4)}$. The acetate turnover
110 rates were calculated as the product of k_{ac} time the acetate concentration, which was analyzed
111 in the sediments at the end of the incubation using high pressure liquid chromatography. The
112 acetate concentrations are summarized in Table 1.

113

114 3. Results

115 Six different lake sediments from Amazonia were incubated in the absence and the
116 presence of $H^{14}CO_3$. Although the rates of CH_4 production were different in the two different
117 incubations, the orders of magnitude were similar for the different lake sediments (Fig. 1A).
118 The incubations in the presence of $H^{14}CO_3$ were used to follow the specific radioactivities of
119 CH_4 (Fig. 2A) and CO_2 (Fig. 2B) over the incubation time. The specific radioactivities of CH_4
120 changed only little but were slightly different for the different lake sediments. The specific
121 radioactivities of CO_2 decreased with time as expected due to the production of non-
122 radioactive CO_2 . Both specific radioactivities were used to calculate the fractions of
123 hydrogenotrophic methanogenesis (f_{H_2}), which increased with incubation time and eventually
124 reached a plateau. The values of f_{H_2} averaged between 30 and 60 d of incubation are
125 summarized in Fig. 1B. Only the incubations of sediment “Grande” did not reach a plateau
126 but still increased after 260 d of incubation due to the continuously decreasing specific
127 radioactivities of CO_2 (data not shown). Averaging these values over the 4 data points
128 between 160 and 260 d resulted in f_{H_2} of about 60% (Fig. 1B). The thus determined values of
129 f_{H_2} were comparable to those determined in the absence of $H^{14}CO_3$ using values of $\delta^{13}C$,
130 which have already been published (Ji et al. 2016) (Fig. 1B).

131 The same sediments were used to determine the turnover of $[2-^{14}C]$ acetate by measuring
132 the increase of radioactive CH_4 (Fig. 3A) and CO_2 (Fig. 3B). These data were used to
133 determine the rate constants of acetate turnover (Fig. 3C), which ranged between 0.02 and 1.7
134 h^{-1} . The respiratory indices (RI) were generally larger than 0.2 except those of the sediments
135 Tapari and Verde, which were smaller than 0.2 (Fig. 4A). The RI values and the acetate



136 turnover rate constants were used to calculate the rates of CH₄ production from acetate in
137 comparison to the rates of total CH₄ production (Fig. 4B). Interestingly, acetate-dependent
138 CH₄ production was always larger than total CH₄ production, except in those sediments
139 exhibiting a RI <0.2.

140

141 **4. Discussion**

142 The RI value quantifies the fraction of the methyl group of acetate that is oxidized to CO₂
143 rather than reduced to CH₄. Since some oxidation of acetate methyl is also happening in pure
144 cultures of aceticlastic methanogens (Weimer and Zeikus, 1978), and since a RI of around 0.2
145 has often been found in environments where acetate turnover was dominated by aceticlastic
146 methanogenesis (Phelps and Zeikus, 1984; Rothfuss and Conrad, 1993; Winfrey and Zeikus,
147 1979), an RI value of 0.2 may in practice be used as the threshold for the change of
148 methanogenic to oxidative acetate turnover. Based on this criterion, i.e. RI < 0.2, the lake
149 sediments of Tapari and Verde behaved as when acetate turnover was exclusively caused by
150 aceticlastic methanogenesis. The percentage of acetate-dependent CH₄ production was fairly
151 consistent with the fraction of hydrogenotrophic methanogenesis, which made up the
152 remainder of total CH₄ production. In conclusion, the acetate turnover and CH₄ production in
153 these lake sediments behaved as expected as when aceticlastic methanogenesis was the sole
154 process of acetate consumption (Fig. 5).

155 However, the sediments of Jua and in particular those of Jupinda, Cataldo, and Grande
156 exhibited RI values >0.2, showing that a substantial fraction of the acetate-methyl was
157 oxidized to CO₂. Hence, acetate was not exclusively consumed by aceticlastic
158 methanogenesis, but it was oxidized, for example by syntrophic acetate oxidation producing
159 H₂ and CO₂. The H₂ and CO₂ may subsequently have been used as methanogenic substrates,
160 thus supporting CH₄ production (Fig. 5). Such support would be consistent with the relatively
161 high fractions (f_{H2}) of hydrogenotrophic methanogenesis observed in these sediments.
162 However, it would not explain why acetate turnover rates were higher than necessary for
163 supporting the observed rates of total CH₄ production. A possible conclusion is that acetate
164 was converted to CO₂ without concomitant production of H₂. Possibly, electrons from acetate



165 were transferred to organic electron acceptors (Fig. 5), such as suggested in the literature
166 (Coates et al., 1998; Lovley et al., 1996). In conclusion, these lake sediments behaved as
167 when acetate consumption was accomplished not only by aceticlastic methanogenesis, but
168 also by oxidative consumption.

169 Our conclusions are mainly based on radiotracer measurements, which may be biased. For
170 example, acetate turnover rate constants are calculated from acetate concentrations and
171 turnover rate constants. Acetate concentrations were only measured at the end of incubation
172 and thus, may not have been representative for much of the incubation time. Furthermore,
173 acetate in the sediment may occur in several pools with different turnover (Christensen and
174 Blackburn, 1982). Therefore, acetate turnover rates and acetate-dependent CH₄ production
175 rates may be overestimated, if the actual acetate turnover depends on a pool size that is
176 smaller than that analyzed. Such overestimation in sediments of Jua, Jupinda, Cataldo and
177 Grande cannot be completely excluded, although rates in sediments of Tapari and Verde were
178 in a realistic range.

179 The determination of fractions of hydrogenotrophic methanogenesis (f_{H_2}) depends on the
180 specific radioactivity of the dissolved CO₂ pool that is involved in CH₄ production. However,
181 it is the pool of gaseous CO₂ that is analyzed in the assay, assuming that its specific
182 radioactivity is identical to that of the active dissolved pool. Since non-radioactive CO₂ is
183 permanently produced from oxidation of organic matter, there may be disequilibrium.
184 Nevertheless, determinations of f_{H_2} using radioactive bicarbonate exhibited the same
185 tendencies as those based on $\delta^{13}C$ values, and thus are probably quite reliable.

186 Despite these reservations, our results collectively demonstrated that acetate turnover in
187 tropical lake sediments did not necessarily follow a canonical pattern with aceticlastic
188 methanogenesis as sole or predominant process of acetate turnover, despite the fact that all
189 these sediments contained populations of putative aceticlastic methanogenic archaea. Acetate
190 consumption in *Methanosaeta* species is known to have a relatively high affinity and a low
191 threshold for acetate (Jetten et al., 1992). Therefore, the question arises why oxidative
192 processes, including syntrophic acetate oxidation, could successfully compete with
193 aceticlastic methanogenesis.



194

195 **5. Author contribution**

196 RC designed the experiments, evaluated the data and wrote the manuscript; MK did the
197 experiments; AEP provided the samples and contributed to the discussion of the data.

198

199 **6. Competing interests**

200 The authors declare that they have no conflict of interest.

201

202 **7. Acknowledgements**

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205

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307
308



309 Table 1: Identity of sediment samples (same as those in Ji et al. (2016)),
310 percentage content of putatively aceticlastic methanogens (*Methanosaetaceae*)
311 relative to total archaea, and concentrations of acetate; mean \pm SE.

312

Lake #	Name	Type	<i>Methanosaetaceae</i> (%)	Acetate (nmol g ⁻¹ dry weight)
P1	Jua	clear water	21 \pm 1	93 \pm 5
P8	Tapari	clear water	19 \pm 3	261 \pm 39
P9	Verde	clear water	19 \pm 11	126 \pm 12
P10	Jupinda	clear water	27 \pm 4	110 \pm 6
A1	Cataldo	white water	42 \pm 1	50 \pm 3
A2	Grande	white water	36 \pm 3	35 \pm 1

313

314



315 **Figure captions**

316 Fig. 1: Methane production in sediments of different Amazonian lakes: (A) rates of CH₄
317 production, and (B) fractions of hydrogenotrophic methanogenesis, both determined in the
318 absence and the presence of radioactive bicarbonate. The data in the absence of radioactive
319 bicarbonate are the same as published in Ji et al. (2016), when f_{H₂} was determined from
320 values of δ¹³C; mean ±SE.

321 Fig. 2: Conversion of radioactive bicarbonate in sediments of different Amazonian lakes: (A)
322 specific radioactivities in CH₄; (B) specific radioactivities in gaseous CO₂; and (C)
323 fractions (f_{H₂}) of hydrogenotrophic methanogenesis; mean ±SE.

324 Fig. 3: Conversion of [2-¹⁴C]acetate in sediments of different Amazonian lakes: (A)
325 accumulation of radioactive CH₄; (B) accumulation of radioactive gaseous CO₂; and (C)
326 acetate turnover rate constants; mean ±SE.

327 Fig. 4: (A) Rates of total and acetate-derived CH₄ production in sediments of different
328 Amazonian lakes and (B) respiratory indices (RI) of the turned over [2-¹⁴C]acetate; mean
329 ±SE.

330 Fig. 5: Scheme of the pathways involved in acetate turnover in sediments of Amazonian
331 lakes; (1) aceticlastic methanogenesis; (2) syntrophic acetate oxidation; (3)
332 hydrogenotrophic methanogenesis; (4) acetate oxidation with organic electron acceptors.
333

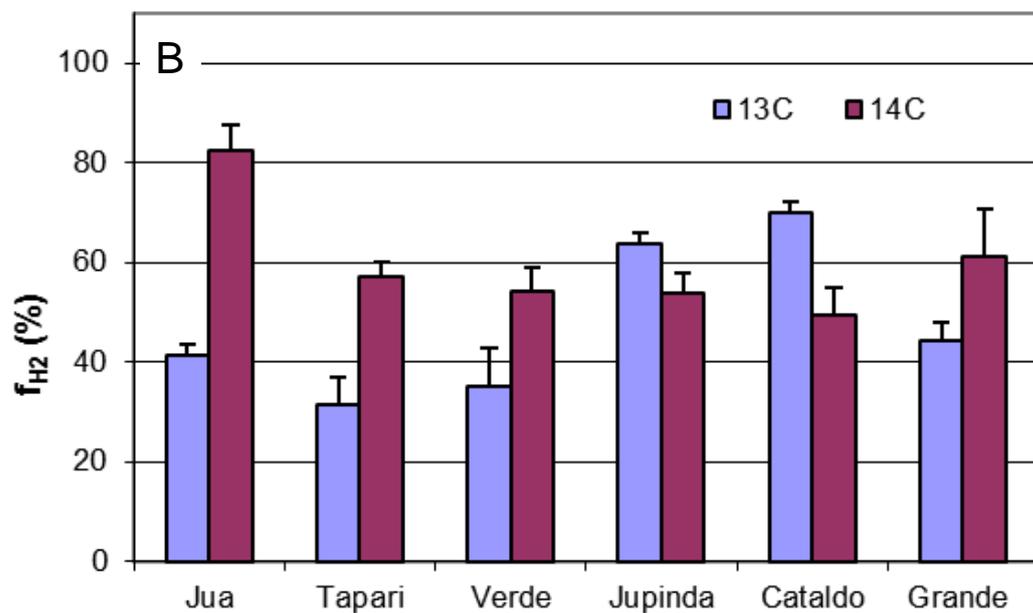
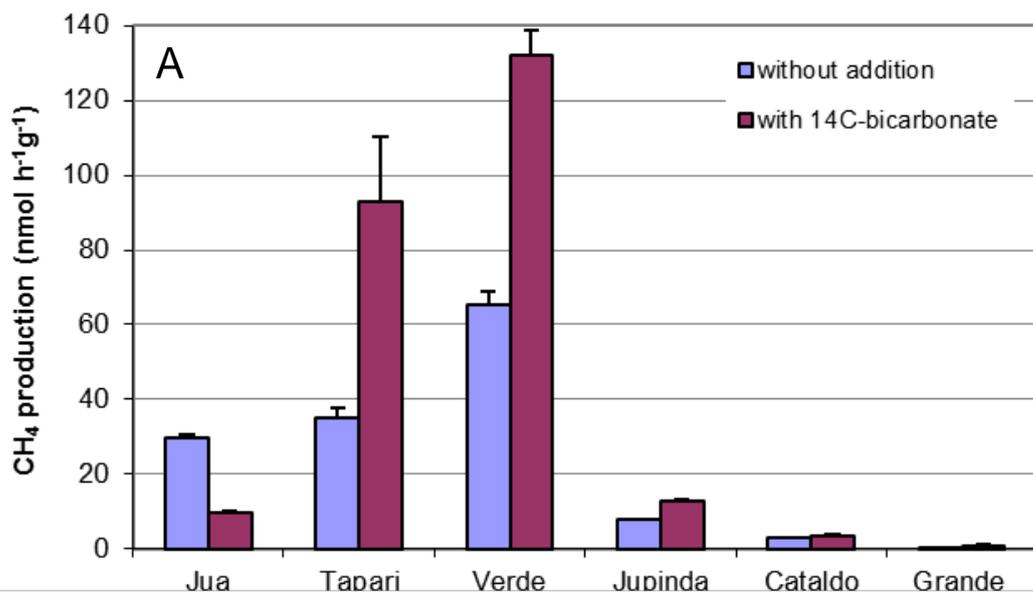


Fig. 1

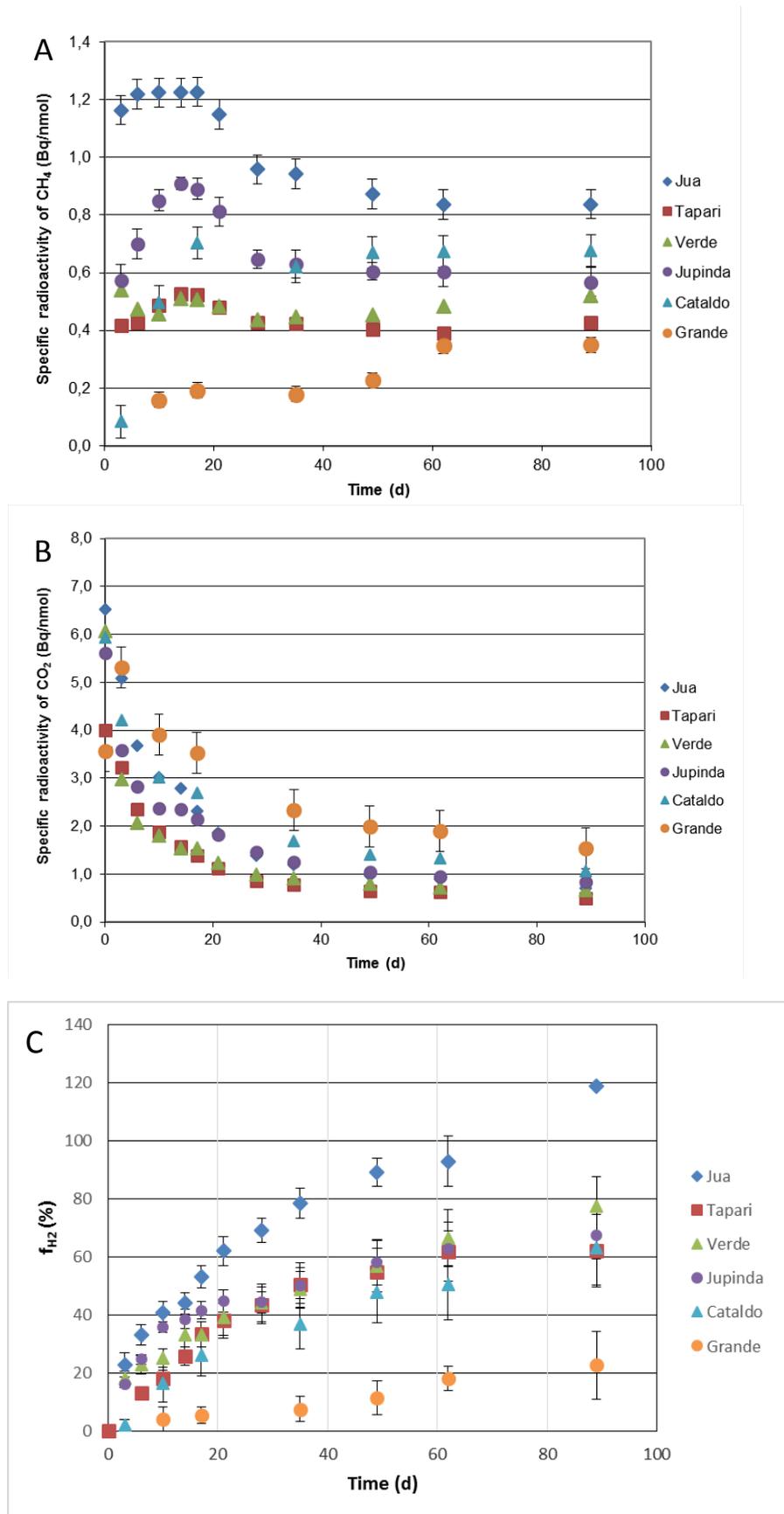


Fig. 2

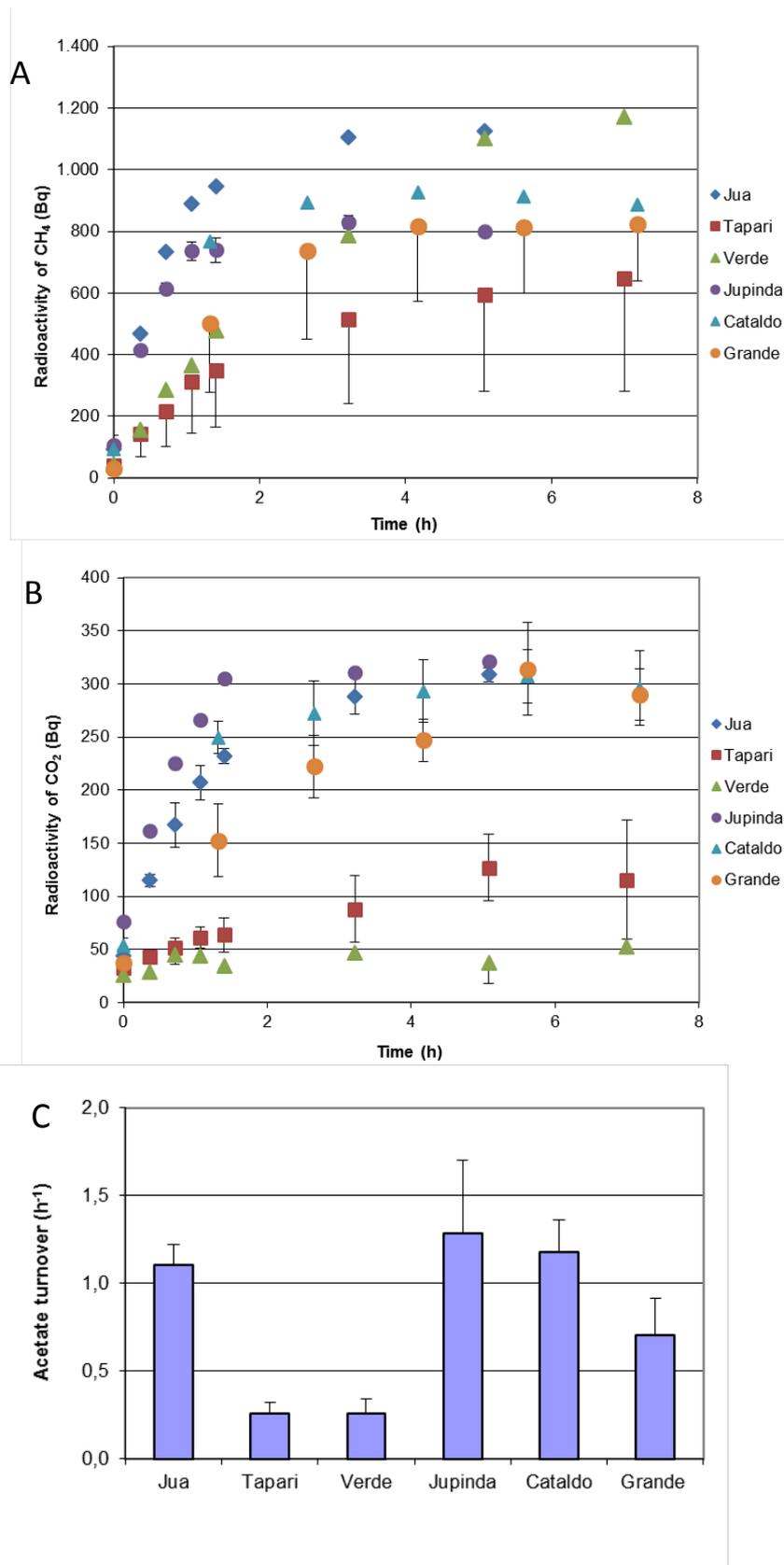


Fig. 3

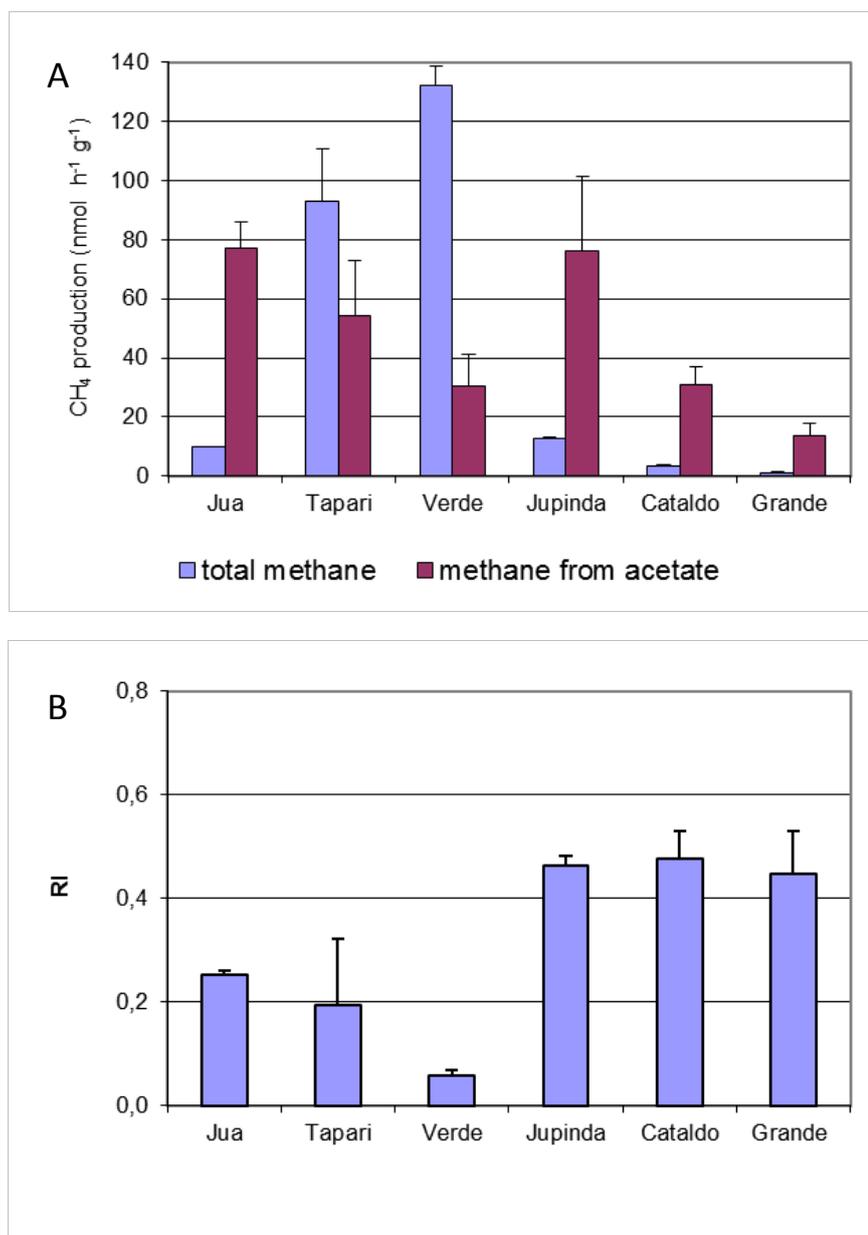


Fig. 4

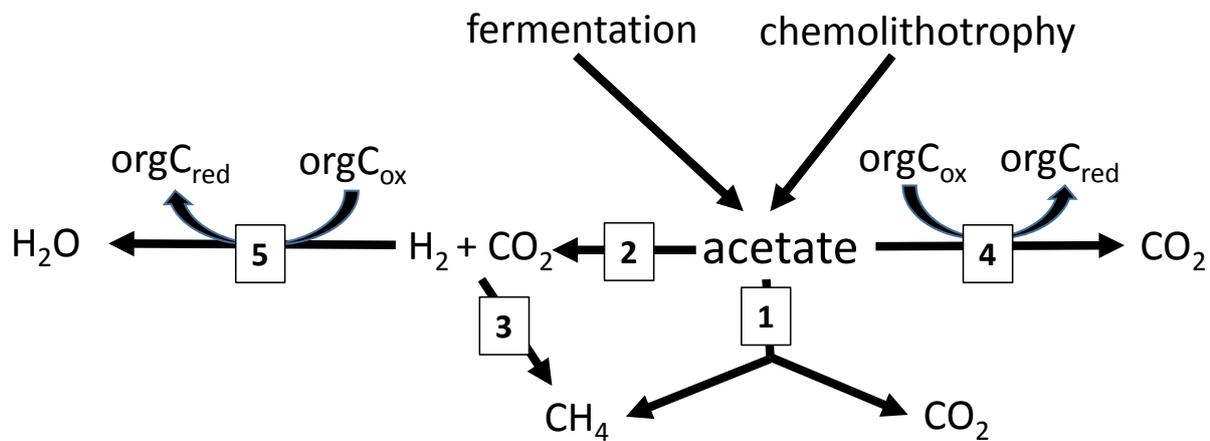


Fig. 5