

1 Response to Referee Comments:

2 *We would like to express our gratitude to Ref #1 for the detailed comments and suggestions*
3 *which helped to improve the manuscript significantly. Our point-by-point responses are*
4 *posted below, with the reviewer's comments being quoted first and our response (R) below*
5 *each comment.*

6

7 **General Comments**

8 This study addresses the underlying factors that may influence the spatial and seasonal
9 distribution of the prokaryotic communities and nutrient dynamics along the Rajang River,
10 South China. Although the results of this study are valid and interesting, there are several points
11 that need to be addressed.

12 1) Dividing sampling cruises into “wet season” and “dry season” may be more beneficial than
13 referring to them individually. Authors mentioned both wet and dry seasons in the
14 Methodology section (section 2.1), however, sampling cruises associated with each are lacking.
15 **R: We agree that it would be beneficial to classify the sampling cruises into “dry season”**
16 **and “wet season”, however as the two “wet” seasons also differ in terms of its microbial**
17 **community composition, we kept the individual cruises in order to prevent confusion**
18 **between the two wet seasons. We have clarified the ‘classification’ of the three cruises as**
19 **wet or dry season in the method section (it now reads: The August 2016 cruise (colored**
20 **red) is classified as the dry season based on the lower mean rainfall value as compared to**
21 **the other two (March 2017 and September 2017), in which the both are classified as the**
22 **wet season (refer to Sup. Fig. 1).**

23

24 2) The site map (Figure 1A) currently shows sampling points throughout the river with source
25 types (Figure 1B), but illustration of anthropogenic activities along the river is missing. It will
26 be helpful to add these as it's not clear which sections of the river are impacted by which
27 activities.

28 **R: Thank you for this. For the anthropogenic activities, the data was extracted from a**
29 **report done by Wetlands International (2015) and is more a qualitative description. This**
30 **description was then used for the classification of land use. The Fig 1(B) was intended**
31 **only to show the zones of peatlands and not for the anthropogenic activities. We have**
32 **added a third map which contains the requested information (Fig. 1C).**

33

34

35 3) Additional statistical analysis, such as PERMANOVA, may be used to infer the impact of
36 anthropogenic activities (e.g. human settlements, effluents, transportation and sand dredging)
37 and source types on beta diversity. Much of the Results and Discussion sections revolve around
38 alpha diversity indices but very little is mentioned about beta diversity.

39 **R: Thank you for pointing this out. Beta diversity was in fact calculated and used for the**
40 **discussion, however, obviously, not clearly pointed out. For example, the plotting of**
41 **nMDS via PRIMER includes Kruskal-Wallis calculations (Kruskal stress formula: 1;**
42 **Minimum stress: 0.01; 2-d: Minimum stress 0.18 occurred 21 times). Furthermore, the**
43 **resemblance matrix was calculated using the Bray-Curtis dissimilarity measure. We have**
44 **included this information in the methods section to reflect its inclusion in our analyses.**
45 **Additional models have been run for the impact of land use on the microbial communities**

46 **and the discussion extended to address this point. New supplementary figures have been**
47 **added as well (10-12)**
48

49 4) Potential functionality inferred from PICRUST showed clear distinction between samples
50 when comparing source type. It would be interesting to see if potential functionality differed
51 too such an extent when samples are compared by anthropogenic perturbations.

52 **R: Thank you for the suggestion. While comparing the potential functionality of the**
53 **anthropogenic perturbations, there was not much variation across the different**
54 **anthropogenic activities, hence this was not included in the results and discussion.**
55

56 5) The general flow of the Discussion section needs improvement. The significance and
57 contribution of the study will have a bigger impact when the Discussion is presented clear and
58 logically. Also, the authors should double check the tense (present/past/passive) for each
59 section.

60 **R: Thank you for pointing this out. We have rearranged the discussion section whereby**
61 **the bulk of the discussion was the drivers of microbial community composition and was**
62 **separated into 3 sections, i.e. spatial and environmental drivers, seasonal drivers and**
63 **anthropogenic drivers.**
64

65 6) Recheck format of in-text references. Not all citations are written in the same format.

66 **R: Thank you. We have checked through the in-text reference and changed those that**
67 **have errors.**
68

69 **Specific Comments**

70

71
72 1.0 Introduction p. 4 paragraph 6: The authors aimed to study microbial diversity and potential
73 function in the Rajang River. Although this study is the first to investigate microbial diversity
74 along a freshwater-marine gradient, with a tropical peatland component, the importance of the
75 river in Malaysia and clear objectives need to stated.

76 **R: Thank you for highlighting this. The last paragraph of the introduction was combined**
77 **with the last few sentences from the previous paragraph. This paragraph now reads:**
78 **“This study focuses on the Rajang River, which is the longest river in Malaysia and one**
79 **of the most socio-economically important peat-draining rivers in South East Asia. It**
80 **transports large amounts of terrestrial material (Müller-Dum et al., 2019) experiences**
81 **two monsoonal seasons (Sa’adi et al., 2017) and is subjected to anthropogenic**
82 **disturbances (Gaveau et al., 2016; Miettinen et al., 2016). Thus, it is fundamental to take**
83 **into consideration both seasonal and anthropogenic influences on the microbial**
84 **communities of the Rajang River. Given the rapid development in Sarawak and the**
85 **importance of microbes in several biogeochemical processes in the Rajang river (Jiang et**
86 **al., 2019; Martin et al., 2018; Müller-Dum et al., 2019; Zhu et al. 2019), it is imperative to**
87 **study the microbial communities to enable future predictions and management responses.**
88 **The Rajang river offers the opportunity to study the microbial diversity along a river to**
89 **sea continuum and at the same time assess influence of natural conditions such as seasons**
90 **(dry vs. wet), different soil types (peat vs. mineral soil), as well as anthropogenic**
91 **disturbances (e.g human settlements and plantations) on microbial succession. This study**
92 **aims to investigate (1) the microbial community structure, diversity and probable**
93 **function across wet and dry seasons in order to (2) understand the underlying factors that**
94 **may influence the spatial and seasonal distribution of the prokaryotic communities and**
95 **the nutrient dynamics involved in the Rajang River.”**

96

97 2.0 Methodology 2.1 Study area and sampling strategy p. 5 line 136-139: “According to
98 Wetlands International (2015), the land surrounding the study sites is characterized by a range
99 of anthropogenic activities, ranging from oil palm and sago plantations to human settlements
100 as well as transportation and sand dredging activities (Fig. 1(B)).” This is not clear from Fig
101 1B. Colours are associated with forested or non-forested lands, however, the map does not
102 depict the different anthropogenic activities along the river.

103 **R: Thank you for this. For the anthropogenic activities, the data was extracted from a**
104 **report done by Wetlands International (2015) and is more a qualitative description. This**
105 **description was then used for the classification of land use. The Fig 1(B) was intended**
106 **only to show the zones of peatlands and not for the anthropogenic activities. We have**
107 **added a new figure highlighting the anthropogenic activities (1C).**

108

109 2.1 Pyrosequencing and Bioinformatics Analyses. Change “Pyrosequencing” to “Illumina
110 sequencing” in the subheading. The authors did not perform 454-pyrosequencing but Illumina
111 sequencing. Did ACE also extract DNA from samples? If that’s the case, the authors should
112 mention this in the beginning of this section.

113 **R: Agreed. “Pyrosequencing” was changed to “Illumina sequencing”. Yes, ACE also**
114 **extracted the DNA samples. This information was placed at the sentence before section**
115 **2.2.**

116

117 3.1 2.4 Statistical Analyses and distLM model. The authors used db-RDA to determine the
118 impact of environmental variables on microbial diversity. The same method can be used to
119 determine which parameters have an influence on specific bacterial taxa. Likewise,
120 Spearman/Pearson correlations can be drawn between environmental parameters and taxa. The
121 information inferred from these additional analyses can help the authors to link certain taxa to
122 specific source types or pollution sources.

123 **R: We agree that this would be a good addition to the existing discussion and have carried**
124 **out the suggested analyses. Spearman’s ranking on the major taxa does support the key**
125 **role of salinity shaping the microbial diversity.**

126

127 I also suggest that anthropogenic inputs should be divided into the following categories: human
128 settlements, effluents (from both palm oil and sago plantations), transportation and sand
129 dredging. Variation partitioning, if possible, may then be used to determine which
130 anthropogenic input, or source type, had the biggest impact on bacterial diversity along the
131 river.

132 **R: Thank you for this suggestion. We did indeed use variation partitioning for the distLM**
133 **models. Unfortunately we do not think that we have sufficient data points from areas**
134 **affected by sand dredging to be included in the model. Transportation of logs and sand**
135 **by boats can be observed throughout the whole river, making it difficult to distinguish its**
136 **impact between different sites.**

137

138 3.0 Results 3.3 Bacterial Distribution according to source type and cruise This section
139 may be improved by organizing it into the following paragraphs: Mention the dominant taxa
140 and their relative abundances. The author mentioned this in the Discussion section (section 4.1
141 line 333-335) but not in the Results section.

142 **R: Agreed, this was moved to the results section 3.3.**

143

144 Which taxa (dominant or specialized) were more predominant at specific source types and/or
145 seasons?

146 In this section, the authors acknowledge a higher Cyanobacterial abundance for the September
147 2017 marine and brackish peat samples. In the Discussion section they refer to the higher
148 Cyanobacterial counts as “blooms” without prior testing (e.g. chlorophyll-a) as a proxy
149 measure to confirm algal blooms. Since Cyanobacteria are more pronounced during warmer
150 months, and anthropogenic activities close to the sampling areas can cause nutrient input and
151 thus proliferation of Cyanobacteria, how valid is the assumption/statement made in the
152 manuscript without additional measurements?

153 **R: Thank you for pointing this out. There were two measurements for validating**
154 **cyanobacterial abundance. One was the respiration experiment, which showed that there**
155 **was greater respiration as compared to oxygen production. The other was phytoplankton**
156 **identification via pigments using a software (CHEMTAX). Only two sampling cruises**
157 **were available for the phytoplankton identification, which unfortunately for September**
158 **2017 is unavailable. However, between the dry (August 2016) and wet (March 2017)**
159 **seasons, the wet season did indeed show greater counts of *Cyanobacteria*.**

160
161 Alpha Diversity Indices: How did the authors calculate the effect of land use and source type
162 on alpha diversity indices? This was not mentioned in the Methodology section.

163 **R: The information “The alpha diversity was calculated using the estimate_richness**
164 **function embedded within the plot_richness function found within the phyloseq package**
165 **utilizing R (v.3.5.3).” is now included in Section 2.2**

166
167 Instead of comparing indices between cruises, the authors can make comparisons between
168 seasons (e.g. compare the entire wet season with the dry season).

169 **R: Referring to the earlier explanation, we kept the individual cruises as the two wet**
170 **seasons exhibited different microbial community composition, which warrants two**
171 **separate cruises instead of the entire wet season.**

172
173 Discussion: I recommend that the authors re-write certain paragraphs of the Discussion section
174 so that it may have a bigger impact on readers. Instead of naming all the different types of taxa
175 in the river, focus on the important ones and what their roles are. How does the environment
176 and different inputs (source types and anthropogenic activities) impact these taxa?

177 **R: Thank you for highlighting this. The sections were further subdivided whereby the**
178 **main discussion was focused on the factors determining bacterial community composition**
179 **and was subdivided into three components which are (4.2.1 Spatial and environmental**
180 **drivers, 4.2.2 Seasonality as a driver of microbial community composition and 4.2.3**
181 **Land-use change and anthropogenic drivers.**

182
183 Possible pathogenic bacteria and/or anthropogenic influence and land-use change. Was
184 *Flavobacterium* the only potential pathogen identified? I would suggest to start the paragraph
185 with anthropogenic influence and land-use change. A second, shorter paragraph can discuss
186 the potential pathogens

187 **R: The possible pathogenic bacteria identified were part of the CFB group. We grouped**
188 **together the other information regarding *Proteobacteria* and *Bacteroidetes* as part of the**
189 **paragraph relating to land use change and anthropogenic drivers.**

190
191

192 **C) Technical comments**

193
194 1.0 Introduction p. 3: Combine paragraphs 2 and 5. Both are discussing lotic environments and
195 nutrient cycling; it will thus make more sense to combine these two.

196 **Response: Agreed, the two paragraphs were combined. This paragraph now reads:**
197 **“Lotic environments are the interface between soil and aquatic environments and aquatic**
198 **environments as terrestrial environments seed microbes into the adjacent water column**
199 **due to flowing waters (Crump et al., 2012). Until not long ago, rivers were thought to be**
200 **passive channels in the global and regional determination of carbon (C) and weathering**
201 **products until it became clear that rivers regulate for example the transfer of nutrients**
202 **from land to coastal areas (Smith and Hollibaugh, 1993). Several studies have shown that**
203 **bacteria are key players in nutrient processing in freshwater systems (Cotner and**
204 **Biddanda, 2002; Findlay, 2010; Madsen, 2011). Zhang et al. (2018a) stated that the**
205 **organic matter composition is strongly modified by bacteria as well as its resistance to**
206 **degradation. Bacteria strongly influence the fluvial organic matter, hence playing a role**
207 **in carbon cycle (Dittmar et al., 2001) and recent studies in the Rajang river have**
208 **demonstrated that as indicated by high concentrations of D-form amino acids (Zhu et al.,**
209 **2019). Moreover, it was demonstrated by Jiang et al. (2019) that Dissolved Organic**
210 **Nitrogen was reduced to NH₄⁺ via mineralization and ammonification, again**
211 **highlighting the biogeochemical activity and the importance of microbes in the Rajang**
212 **River. Until now, there has, however, been no study on their diversity yet; a gap that this**
213 **study aims to fill. Thus, it is essential to understand the dynamics and structure of**
214 **microbial communities in them to assess their contribution towards biogeochemical**
215 **fluxes such as carbon and nitrogen (Battin et al., 2008; Raymond et al., 2013), as well as**
216 **phosphate cycling (Hall et al., 2013). In addition, the fluxes as well as transformations of**
217 **organic matter as well as nutrients in aquatic systems are environmentally driven by**
218 **parameters such as temperature or the availability of nutrients in these ecosystems (Welti**
219 **et al., 2017). In turn, various gradients (i.e physical, chemical, hydrological or even**
220 **biological) contribute to the changes in the microbial diversity and distribution living**
221 **within the lotic environments (Zeglin, 2015).”**

222
223 p. 4 line 93-93: Due to their high diversity and fast generation time, microbial communities are
224 the first responders to environmental changes

225 **R: Agreed and changed to recommended sentence.**

226
227 p. 4 line 96: Liao et al. (2019) showed that p. 4 line 97: delete “further” p. 4 line 97:
228 Bruland et al. (2008) demonstrated that the

229 **R: Agreed, removed “further”**

230
231 p. 4 line 99-102: “Thus, as the Rajang River experiences two monsoonal seasons (Sa’adi et al.,
232 2017) and is subjected to anthropogenic disturbances (Gaveau et al., 2016; Miettinen et al.,
233 2016), it is thus fundamental to take into consideration both seasonal and anthropogenic
234 influences on the microbial communities of the Rajang River.” This forms part of the aim and
235 objective and should rather move to last paragraph

236 **R: Agreed, this was changed to better reflect the aim and importance of the Rajang River.**

237
238 p. 4 line 115: delete “hypothesized”

239 **R: Agreed, removed “hypothesized”**

240
241 p. 4 line 120-121: as well as anthropogenic disturbances (e.g. human settlements and
242 plantations) on microbial succession.

243 **R: Agreed, changed to recommended sentence.**

244
245 p. 4 line 121-122: Delete “Linear models are

246 used to examine the relationship between the microbial community structure and their
247 environment.”

248 **R: Agreed to remove sentence as it was already explained in methodology.**

249

250 2.0 Methodology 2.1 Study area and sampling strategy p. 5 line 130: The region:
251 **R: Agreed. Changed to “The”**

252

253 p.5 line 134: small tributaries
254 **R: Agreed, changed from “distributaries” to “tributaries”**

255

256 p. 5 line 142: Which months were associated with the wet and dry season, respectively?
257 **R: The following sentence was extracted from the caption of Sup. Fig. 1. “ The August
258 2016 cruise (colored red) is classified as the dry season based on the lower mean rainfall
259 value as compared to the other two (March 2017 and September 2017), in which the both
260 are classified as the wet season.” to be placed in-text in the methodology for ease of
261 reference**

262

263 p. 5 line 149: Approximately 250 – 500 mL of water: : .
264 **R: Agreed, changed to “approximately”**

265

266 p. 5 line 153-156: A total of 117 filters were recovered (1 x 3.0 μ m filter was discarded
267 due to contamination) and immediately stored at - 20_C.
268 **R: Agreed. The sentence was changed as recommended.**

269

270 2.2 Pyrosequencing and Bioinformatics Analyses p. 5 line 160: Briefly, fastq files
271 generated: : .
272 **R: Agreed. “In short” changed to “Briefly”**

273

274 p. 6 line 161: quality trimmed with fastqc, primer sequeces: : .
275 **R: Agreed, changed from “processed” to “quality trimmed”**

276

277 p. 6 line 162-163: High quality sequences were subsequently processed using the Mothur
278 pipeline.
279 **R: Agreed. Changed to the recommended sentence.**

280

281 p. 6 line 164: SILVA database
282 **R: Agreed, “alignment” changed to “database”**

283

284 p. 6 line 171: potential functional genes
285 **R: Agreed, added “potential”**

286

287 2.3 Physico-chemical Data and Geochemical Analyses p. 6 line 179: in-between the
288 Cruises
289 **R: Agreed, added “-“**

290 p. 6 line 189-191: Belawai samples (2_13’47.16"N, 111_12’19.04"E) were used
291 in an incubation experiment to study the net primary productivity and respiration rate
292 of the Rajang River. Technical triplicates were incubated in both light and dark set-ups
293 (Refer to Supp. Table 1 for details).
294 **R: Agreed, the sentence was modified to the recommended sentence.**

295

296 2.4 Statistical Analyses and distLM model p. 6 line 195-197: to determine if the various
297 terrestrial source types or different land use impacted bacterial community composition.
298 **R: Agreed, sentence structure was changed to the recommended.**
299

300 p. 7 line 199: what type of normalization method was used?
301 **R: The following sentence was added: “using the “Normalise Variables” function in the**
302 **PRIMER 7 software”.**
303

304 p. 7 line 202-204: “The authors would like to note that the distLM models are based on only
305 the August 2016 and March 2017 cruise as there was a lack of physico-chemical data from the
306 September 2017 cruise due to malfunctioning equipment.” Delete this sentence, no
307 need to mention this twice, at the end of the paragraph (lines 205-208) is sufficient
308 **R: Agreed, the sentence was removed.**
309

310 p. 7 line 215: A total of 74,690 high quality bacterial sequences: : :.
311 **R: Agreed and changed as recommended.**
312

313 3.2 Shifts in bacterial community structure p. 7 line 223-224: Delete this sentence, it’s
314 Redundant
315 **R: Agreed, the sentence was removed.**
316

317 p. 7 line 230: August 2016 (dry season) samples
318 **R: Agreed, added “(dry season)” to the sentence.**
319

320 p. 7 line 231: September 2017 (wet season) samples
321 **R: Agreed, added “(wet season)” to the sentence.**
322

323 p. 7 line 231: There were clear overlaps between samples from:
324 **R: Agreed, changed from “there are apparent” to “there were clear”**
325

326 p. 7 line 232-233: We also observed a gradual shift in bacterial composition from mineral soils
327 and freshwater peat towards brackish and marine samples.
328 **R: Agreed, sentence was changed accordingly as recommended.**
329

330 3.3 Bacterial Distribution according to source type and cruise p. 8 line 240: Delete “Fig
331 3 show that”
332 **R: Removed as recommended but added (Fig. 3) to the end of the sentence.**
333

334 3.4 Alpha Diversity Indices p. 8 line 263-264: Rewrite the sentence
335 **R: Sentence was rewritten as “For the September 17 cruise, we observed increased values**
336 **of Chao1 across the brackish peat, freshwater peat as well as mineral soils.”**
337

338 p. 8 line 265: microbial communities varied significantly along the different source types
339 **R: Agreed, changed sentence as recommended.**
340 p. 8 line 266: to be higher than that of March 2017:
341 **R: Agreed, changed from “found in” to “of”**
342

343 p. 9 line 276: Authors are referring to “upstream” samples in this sentence, which samples are
344 these? They did not clearly differentiate between upstream and downstream samples in the
345 Methodology section which is causing confusion in subsequent text.

346 **R: Agreed. Added explanation at the end of the text: (i.e. Human Settlement, Oil Palm**
347 **and Sago Plantation, Oil Palm Plantation and Secondary Forest).**

348

349 p. 9 line 289-290: Potential KEGG pathways between (i) marine and brackish peat, and (ii)
350 freshwater peat and mineral soil were similar. There were differences between source types
351 and seasons

352 **R: Agreed, the recommended sentence provided more clarity.**

353

354 p. 9 line 290-292: Delete this sentence. It's part of Discussion

355 **R: Agreed and removed.**

356

357 p. 9 line 301: Dissolved Inorganic Phosphate (10.57%)

358 **R: removed "at" and added parenthesis to "10.57%".**

359

360 p. 9 line 304: Delete "lastly"

361 **R: Agreed.**

362

363 p. 9 line 305: Dissolved Inorganic Nitrogen (4.29%) respectively made up the

364 **R: Agreed, changed from "(4.29%, respectively)" to "(4.29%) respectively**

365

366 p. 10 line 308-309: Move this sentence to p. 9 line 300: "Marginal DistLM was performed in
367 order to gauge the extent of physicochemical parameters or environmental variables accounting
368 for a compelling proportion of variation in the bacterial communities. Significant vectors of
369 environmental variables ($R^2 > 0.3892$, $P < 0.001$) were 308 calculated based on a linear model
370 (DistLM) and plotted against the bacterial community composition as shown in Fig 7. Salinity
371 was the single best predictor variable

372 **R: Agreed and changed.**

373

374 p. 10 line 311-320: The distLM model clustered samples from the August 2016 cruise
375 separately from the March 2017 samples. Brackish peat, as well as marine samples from
376 August 2016, correlated more strongly with salinity, irrespective of land use. On the contrary,
377 the March 2017 samples were found to cluster separately with DO. In addition, the August
378 2016 mineral soil samples correlated with silicate.

379 **R: Agreed and changed to suggested sentences.**

380

381 p. 10 line 332: Delete this subheading and move subheading 4.2 to 4.1

382 **R: Agreed, the remaining labels were corrected accordingly.**

383

384 p. 10 line 335-342: in varying abundances, indicating high variation within the system. The
385 majority of bacterial taxa were restricted to a relatively small number of assemblages. However,
386 due to the heterogeneity of the Rajang River, substantial shifts in OTU diversity were shown,
387 while exhibiting successional changes in community composition downstream. We observed
388 abrupt shifts in terms of richness and diversity as well as bacterial distribution, which was
389 structured according to macro-scale source types. Staley et al. (2015) proposed that variability
390 in microbial communities were less due to the presence/absence but likely due to shifts in
391 relative abundance of OTUs.

392 **R: Agreed and changed.**

393

394 p. 10 line 342: community composition, overlap between the core microbiome (i.e. free-living
395 and particle-attached portions) of samples were not evident.

396 **R: Agreed and changed.**

397

398 p. 11 line 346: Change “further supported” with “demonstrated”

399 **R: Agreed and changed.**

400

401 p. 11 line 351: The short residence time in the Rajang River likely reflected a similar scenario
402 to San Francisco Bay (Reference).

403 **R: Agreed and changed.**

404

405 p. 11 line 372-378: Delete these sentences. Beta-proteobacteria was already mentioned in the
406 previous paragraph.

407 **R: Agreed and removed.**

408

409 p. 11 line 380: Were there really Cyanobacterial blooms?

410 **R: Thank you for pointing this out. Cyanobacterial bloom was changed to “the higher
411 abundance of *Cyanobacteria*”, which more accurately describes the composition as shown
412 by the abundance in taxa.**

413

414 p. 12 line 385: *Sphingomonas*, a purple-sulfur bacteria,

415 **R: Agreed and changed.**

416

417 p. 12 line 391: indicating its preference for this environment. It’s interesting to note that most
418 studies on

419 **R: Agreed and changed to recommended sentence.**

420

421 p. 12 line 394: In most of these studies, *Deinococcus-Thermus* was found in low abundance
422 (e.g. 1% in Antarctic marine environments, 1.5% in hypersaline soils; Giudice and Azzaro,
423 2019; Vera-Gargallo et al., 2019) when compared to the Rajang River.

424 **R: Agreed and changed.**

425

426 p. 12 line 397: Start new paragraph with: “There was a fundamental shift in bacterial
427 community composition when taking the major taxa into consideration. There was a clear
428 distinction between dry and wet seasons with an overall higher species richness and diversity
429 for the dry season” For the wet season, focus on both the March and September cruises to make
430 a conclusion

431 **R: Agreed and changed.**

432

433 p. 13 line 421-427: Delete these sentences, was already mentioned in Methodology

434 **R: Agreed, the sentences were removed.**

435

436 p. 13 line 427: There was a continual shift

437 **R: Agreed, changed “is” with “was”.**

438

439

440 p. 13 line 432: similar to findings by

441 **R: Agreed, changed “akin” to “similar”**

442

443 p. 13 line 434: likely explaining the reduced relative abundances of some taxa. For example,
444 *Chloroflexi* has a higher relative abundance upstream while *Deinococcus-Thermus* shows
445 lower relative abundance downstream.

446 **R: Agreed and changed to recommended sentence.**

447

448 p. 13 line 438: Delete “salinity gradients”

449 **R: Agreed and removed.**

450

451 p. 13 line 451: Salinity, DIP () and dissolved oxygen are major environmental drivers of species
452 distribution (References). In this study, marine and brackish peat samples correlated well with
453 salinity.

454 **R: Agreed and changed.**

455

456 p. 14 line 459-469: Not sure what the authors want to say here. Do they assume there was high
457 or low bacterial productivity?

458 **R: We deduced that even though there was high abundance of associated phyla that may**
459 **contribute to the production of O₂ (via primary production), the high CO₂ emissions and**
460 **higher respiratory rate show that there was higher bacterial productivity versus primary**
461 **production.**

462

463 p. 14 line 478-480: While the development of unique community structures was strongly
464 influenced by spatial factors, seasonality also played a role. Seasonal variability was also
465 observed between the

466 **R: Agreed and changed to recommended sentence.**

467

468 p. 14 line 485-490: Again, can the term Cyanobacterial bloom be accurately used?

469 **R: Thank you once again for pointing this out. The sentence was changed to “The greater**
470 **abundance of Bacteroidetes in March 2017 may be indicative of the community**
471 **composition adjusting due to the processing of organic material caused by the higher**
472 **cyanobacterial abundance in the September 2017 cruise. This was similar to a study by**
473 **Pinhassi et al., (2004), in which the higher abundance of Bacteroidetes follows after an**
474 **algal bloom.” This would reduce the assumption of a cyanobacterial bloom. The study**
475 **quoted (Pinhassi et al. (2004)) was used as an example for probable inference of**
476 **cyanobacterial bloom but cannot yet be confirmed.**

477

478 p.15 line 494: “were similar in terms of climate”

479 **R: Agreed, and changed.**

480

481 p. 15: Start the paragraph with line 515: “The results obtained from this study suggest that the
482 run-off from anthropogenic activities alters the microbial community composition.
483 Anthropogenic disturbances, in particular settlements and logging (secondary forest), led to
484 higher diversity indices (Fig .6). On the contrary, sites surrounded by oil

485 **R: Agreed and changed. The breaking of paragraphs provide better clarity to the overall**
486 **flow.**

487

488

489 p. 16 line 543: The authors refer to “pristine and less pristine environments”. Which sites were
490 classified as pristine, and which were less pristine?

491 **R: Thanks for pointing this out. We have changed this to “anthropogenic perturbations**
492 **(regions with oil palm plantations and human settlements) led to increased richness but**
493 **less diversity compared to those that were less affected by anthropogenic perturbations**
494 **(coastal zone and secondary forest).”**

495

496 p. 16 line 545: The PICRUST results showed differences between source types

497 **R: Agreed, changed “difference” to “differences”**

498

499 p. 16 line 550: mixing experiments. This approach will contribute towards a better
500 understanding of the response of microbial communities to anthropogenic perturbations, as
501 well as their role in degrading peat-related run-off from

502 **R: Agreed, and changed to suggested sentence.**

503

504

505

506

507

508 *We would like thank Ref #2 for the comments and suggestions which helped to improve the*
509 *manuscript significantly. Our point-by-point responses are posted below, with the reviewer's*
510 *comments being quoted first and our response (R) below each comment.*

511
512 The manuscript of Sia et al. describes a study of bacterial communities' distribution in a section
513 of the Rajang River. Overall, the quality and content of the paper is in line with similar
514 publications on lotic bacterial communities, where the community composition is linked to
515 environmental parameters. The strongest point of the study is that it covers multiple time points
516 (different seasons) and several salinity zones. The authors also made an attempt to estimate
517 potential functions of the bacterial communities. I would like to note a detailed and
518 comprehensive Discussion section. However, some revision is necessary. Certain results need
519 to be verified, methods described more in details (please see specific comments). English
520 language could be improved; the manuscript is not free of mistakes and misprints.

521
522 Some specific questions and comments:

523
524 P 5 L 146 – it is not clear for me how is classification into freshwater and brackish water
525 described in Fig. 1(B). Possibly that is due to the poor quality of the map.

526 **R: Thank you for pointing this out. We removed this sentence “as described in Fig 1. (B)”**
527 **as Fig. 1(B) is to show the areas with peat only.**

528
529 P 5 L 150, 152 – Are you sure that those were polycarbonate filters? GF are usually glass fiber
530 filters.

531 **R: Thank you for pointing this out. The correct filter used was Nuclepore™ Track-**
532 **Etched Polycarbonate Membrane Filter. We have removed the (GF/C) description.**

533
534 P 5 L 156 – Incorrect reference. Caporaso et al. 2012 describe QIIME pipeline,
535 not Illumina sequencing.

536 **R: Agreed. Changed to Bentley et al. (2008) which describes the first paper that Illumina**
537 **was based upon.**

538
539 P 5 L 156 – Could you please add more information on DNA
540 extraction and library preparation procedures, for example, which primers were used
541 for amplification?

542 **R: Thank you for pointing this out. We have included the relevant information in the**
543 **methods section. It now reads:A total of 117 filters were recovered (1 x 3.0 µm was**
544 **discarded due to contamination) and immediately stored at -20 °C and sent to the**
545 **Australian Centre for Ecogenomics (ACE), Brisbane for DNA extraction, library**
546 **preparation and processing utilizing the Illumina (Bentley et al., 2008) platform.**

547 **2.2 Illumina Sequencing and Bioinformatics Analyses**

548 **Initial upstream processes were carried out by the Australian Centre for Ecogenomics**
549 **utilizing the ACE mitag pipeline (ACE, 2016). The primers utilized were based on the V3**
550 **– V4 hypervariable regions of the 16S rRNA gene.**

551
552 P 6 L 163 – Reference for Mothur pipeline missing.

553 **R: Thank you for pointing this out. The relevant citation was added (Schloss et al., 2009)**

554
555 P 6 L 175 –Reference for the GreenGenes database missing.

556 **R: Thank you for pointing this out. The relevant citation was added (DeSantis et al., 2006)**

557 P 7 L 215 – Can you explain why the sequencing depth was so low, especially for some samples?
558 Was it on purpose?

559 **R: Thank you for this question. The minimum sequencing depth was 10,000 reads per**
560 **sample. After QC and removal of unknown sequences, some samples were left with a very**
561 **low read count. Given the general lack of data from these systems and to ‘lose’ as little**
562 **information /samples as possible, we chose a low read number for the subsampling.**

563

564 P 7 L 215 – Were the sequences deposited to a public database?

565 **R: Raw sequences have been deposited with the NCBI BioSample database under**
566 **BioProject ID PRJNA565954.**

567

568 P 7 L 232 – Are you sure it is “brackish peat” and not “freshwater peat”, which seems to me
569 from Fig.2?

570 **R: Yes, thank you for pointing this out. “Brackish peat” was changed to “freshwater peat”**

571

572 P 8 L 247-249 – This observation is not obvious to me from Fig. 3.

573 **R: Agreed, this portion was removed.**

574

575 P 8 L 258-259 – was the difference between OTU counts statistically significant?

576 **R: The results shown were plotted based on the calculations from the estimate_richness**
577 **function in the *phyloseq* package, and hence the observation was more a qualitative**
578 **observation.**

579

580 P 10 L 324 – I didn’t find any description of the results separately for free-living and particle-
581 attached bacteria, however you discuss them a bit in chapter 4.1 in relation to Supp. Fig. 3.
582 Were the results pooled together for free-living and particle-attached bacteria in Fig. 2-7?

583 **R: Thank you for pointing this out. Yes, for Figures 2 – 7 the results were pooled together**
584 **for discussion as the difference between free-living and particle-attached bacteria did not**
585 **exhibit clear distinction and hence was not further elaborated. The following sentence**
586 **was added in Section 2.2: “Apart from the results and discussion shown for free-living**
587 **and particle-attached bacteria, the remaining discussion is based on the pooled results of**
588 **both components”**

589

590 P 11 L 378-380 – How does the dominance of Proteobacteria indicate its role in nitrogen
591 cycling? Please explain how it is complementary to Cyanobacteria bloom, the message
592 is unclear.

593 **R: The sentence was rephrased as “In a study by Yang et al. (2013), the dominance of**
594 **Proteobacteria influenced the nitrogen cycle via the processes of nitrification and**
595 **denitrification, in which aeration would increase its abundance and result in higher**
596 **mortality of cyanobacteria.**

597

598 P 12 L 394- 397 – “In contrast, most extreme environments show” this sentence sounds strange
599 and needs to be rephrased.

600 **R: Agreed, this sentence was changed to “In most of these studies, Deinococcus-Thermus**
601 **was found in low abundance (e.g. 1% in Antarctic marine environments, 1.5% in**
602 **hypersaline soils; Giudice and Azzaro, 2019; Vera-Gargallo et al., 2019) when compared**
603 **to the Rajang River.”**

604

605

606 **Biogeographical distribution of Microbial Communities along the Rajang River-South China**
607 **Sea Continuum**

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619

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621

622 **Abstract**

623 The Rajang River is the main drainage system for central Sarawak in Malaysian Borneo which passes
624 through peat domes whereby peat-rich material is being fed into the system and eventually into the
625 southern South China Sea. Microbial communities found within peat-rich systems are important
626 biogeochemical cyclers in terms of methane and carbon dioxide sequestration. To address the critical
627 lack of knowledge about microbial communities in tropical (peat-draining) rivers, this study represents
628 the first seasonal assessment targeted at establishing a foundational understanding of the microbial
629 communities of the Rajang River-South China Sea continuum. This was carried out utilizing 16S rRNA
630 gene amplicon sequencing via Illumina MiSeq in size-fractionated samples (0.2 and 3.0 µm GF/C filter
631 membranes) covering different biogeographical features/sources from headwaters to coastal waters.
632 The microbial communities found along the Rajang river exhibited taxa common to rivers (i.e.
633 predominance of *β-Proteobacteria*) while estuarine and marine regions exhibited taxa that were
634 common to the aforementioned regions as well (i.e. predominance of *α-* and *γ-Proteobacteria*). This is
635 in agreement with studies from other rivers which observed similar changes along salinity gradients. In
636 terms of particulate versus free-living bacteria, nonmetric multi-dimensional scaling (NMDS) results
637 showed similarly distributed microbial communities with varying separation between seasons. Distinct
638 patterns were observed based on linear models as a result of the changes in salinity along with variation
639 of other biogeochemical parameters. Alpha diversity indices indicated that microbial communities were
640 higher in diversity upstream compared to the marine and estuarine regions whereas anthropogenic
641 perturbations led to increased richness but less diversity. Despite the observed changes in bacterial

642 community composition and diversity that occur along the Rajang River to sea continuum, the PICRUSt
643 predictions showed minor variations. The results provide essential context for future studies such as
644 further analyses on the ecosystem response to anthropogenic land-use practices and probable
645 development of biomarkers to improve the monitoring of water quality in this region.

646

647 Keywords: particle-associated microbes, free-living microbes, 16S rRNA, Rajang river, River-sea
648 continuum

649

650

651

652

653

654 **1. Introduction**

655 Biogeochemical transformations are primarily governed by microbial communities (Konopka, 2009),
656 and it is crucial to understand their dynamics in order to predict biosphere modulations in response to a
657 changing climate. Despite the importance of freshwater to society and despite hosting the highest
658 microbial diversity (Besemer et al., 2013), microbial community composition and diversity in
659 freshwater habitats, especially in lotic environments, are much less studied compared to marine and soil
660 communities (Kan, 2018).

661

662 Lotic environments are the interface between soil and aquatic environments and aquatic environments
663 as terrestrial environments seed microbes into the adjacent water column due to surface runoff (Crump
664 et al., 2012). Until recently, rivers were thought to be passive channels in the carbon (C) cycling and
665 weathering products until it became clear that rivers regulate for example the transfer of nutrients from
666 land to coastal areas (Smith and Hollibaugh, 1993). Several studies have shown that bacteria are key
667 players in nutrient processing in freshwater systems (Cotner and Biddanda, 2002; Findlay, 2010;
668 Madsen, 2011). Zhang et al. (2018a) stated that the organic matter composition is strongly modified by
669 bacteria as well as its resistance to degradation. Bacteria strongly influence the fluvial organic matter,
670 hence playing a role in carbon cycle (Dittmar et al., 2001) and recent studies in the Rajang river have
671 demonstrated that as indicated by high concentrations of D-form amino acids (Zhu et al., 2019).
672 Moreover, it was demonstrated by Jiang et al. (2019) that dissolved organic nitrogen was mineralized
673 to NH_4^+ , again highlighting the biogeochemical activity and the importance of microbes in the Rajang
674 River. Until now, there has, however, been no study on their diversity yet; a gap that this study aims to
675 fill. Thus, it is essential to understand the dynamics and structure of microbial communities in them to
676 assess their contribution towards biogeochemical fluxes such as carbon and nitrogen (Battin et al., 2008;
677 Raymond et al., 2013), as well as phosphorus cycling (Hall et al., 2013). In addition, the fluxes as well
678 as transformations of organic matter as well as nutrients in aquatic systems are environmentally driven
679 by parameters such as temperature or the availability of nutrients in these ecosystems (Welti et al.,
680 2017). In turn, various gradients (i.e physical, chemical, hydrological or even biological) contribute to
681 the changes in the microbial diversity and distribution living within the lotic environments (Zeglin,
682 2015).

683

684 Next-generation sequencing technologies have enabled a better understanding of the rare or
685 unculturable biosphere which traditional culture methods would not have been able to elucidate
686 (Boughner and Singh, 2016; Cao et al., 2017). Only few studies assessing bacterial community
687 composition have been undertaken in lotic/riverine environments (Fortunato et al., 2012; Ladau et al.,
688 2013; Zwart et al., 2002), with even less focusing on the diversity of surface-attached biofilms in lotic
689 environments, particularly in comparison to biofilm studies in benthic habitats (Zeglin, 2015).
690 Furthermore, bacterial assemblages on suspended particles were shown to differ from free-living

691 bacterioplankton in a number of studies (Bidle and Fletcher, 1995; Crump et al., 1999) in which the
692 ratios between both fractions are often influenced by the quality of suspended particulate matter
693 (Doxaran et al., 2012). Even less studies attempt to map bacterial community composition in a river-to-
694 sea continuum across multiple seasons and habitats (Fortunato et al., 2012) and it was only recently
695 reported that the most abundant riverine bacterioplankton resemble lake bacteria and can be regarded
696 as ‘typical’ freshwater bacteria (Lozupone and Knight, 2007; Zwart et al., 2002). Metagenomics studies
697 substantiated the dominance of *Proteobacteria* and *Actinobacteria* whereby *Bacteroidetes*,
698 *Cyanobacteria*, and *Verrucomicrobia* were found also found to be abundant in rivers (Cottrell et al.,
699 2005; Kolmakova et al., 2014; Lemke et al., 2009; Newton et al., 2011; Read et al., 2015; Staley et al.,
700 2013). While there are studies related to the freshwater-marine gradients of rivers such as studies by
701 Crump and Hobbie (2005) and Fortunato et al. (2013) and tropical peatlands (Kanokratana et al., 2011;
702 Mishra et al., 2014; Yule et al., 2016; Too et al., 2018), to the author’s knowledge, this is the first study
703 which links both freshwater-marine gradients as well as tropical peatlands as a cohesive component (i.e.
704 tropical peat-draining river to coastal ecosystem). Due to their high diversity and fast generation time,
705 microbial communities (Hunt and Ward, 2015) are the first responders to environmental changes (both
706 natural and anthropogenic events such as storms, upwelling and pollutants). Liao et al. (2019) showed
707 that extensive agricultural land-use in the inter-tidal region of a watershed resulted in the prevalence of
708 bacteria pathogen-like sequences whereas Bruland et al. (2008) stated that the assemblages of microbes
709 also vary temporally as a function of oceanographic conditions, river discharge, tidal phase and season.

710

711 This study focuses on the Rajang River, which is the longest river in Malaysia and one of the most
712 socio-economically important peat-draining rivers in South East Asia. It transports large amounts of
713 terrestrial material (Müller-Dum et al., 2019), experiences two monsoonal seasons (Sa’adi et al., 2017),
714 and is subjected to anthropogenic disturbances (Gaveau et al., 2016; Miettinen et al., 2016). Thus, it is
715 fundamental to take into consideration both seasonal and anthropogenic influences on the microbial
716 communities of the Rajang River. Given the rapid development in Sarawak and the importance of
717 microbes in several biogeochemical processes in the Rajang river (Jiang et al., 2019; Martin et al., 2018;
718 Müller-Dum et al., 2019; Zhu et al., 2019), it is imperative to study the microbial communities to enable
719 future predictions and management responses. The Rajang river offers the opportunity to study the
720 microbial diversity along a river to sea continuum and at the same time assess influence of natural
721 conditions such as seasons (dry vs. wet), different soil types (peat vs. mineral soil), as well as
722 anthropogenic disturbances (e.g human settlements and plantations) on microbial succession. This study
723 aims to investigate (1) the microbial community structure, diversity and probable function across wet
724 and dry seasons in order to (2) understand the underlying factors that may influence the spatial and
725 seasonal distribution of the prokaryotic communities and the nutrient dynamics involved in the Rajang
726 River.

727

728 **2. Methodology**

729

730 **2.1 Study area and sampling strategy**

731 This study was conducted along ~300 km of the Rajang river in Sarawak, Malaysia (**Fig. 1A**). The
732 region has an equatorial climate characterized by constant temperatures, high extensive rainfall and
733 high humidity (Wang et al., 2009, 2005; see also **Supp. Fig. 1**). The Rajang delta system consists of an
734 alluvial valley, an associated coastal plain and a delta plain (Staub and Esterle 1993). The coastal plain
735 is dissected into several small tributaries, namely Igan, Lassa, Paloh and Rajang (**Fig. 1A**). The
736 shoreline experiences tides and seasonally strong waves ranging from 3 – 6 m with intensity increasing
737 from the east to the west. According to Wetlands International (2015), the land surrounding the study
738 sites is characterised by land use change (**Fig 1B**) and a range of anthropogenic activities, such as oil
739 palm and sago plantations (**Fig 1C**), human settlements as well as transportation and sand dredge.

740

741 A total of 59 water samples were collected along salinity-gradients during three (3) cruises (**Fig. 1A**),
742 covering both wet and dry seasons as well as different source types (i.e. mineral or peat soils). Source
743 types sampled were grouped as follows: 1) marine 2) brackish peat 3) freshwater peat and 4) mineral
744 soils. From Sibul towards Kapit (upriver), the riparian zone is mineral soil whereas from Sibul
745 downwards to the coast it consists of peat which was then further divided into freshwater (salinity 0 to
746 ~ 1 PSU) and brackish (salinity 2- 28 PSU). The August 2016 cruise (coloured red) is classified as the
747 dry season based on the lower mean rainfall value as compared to the other two (March 2017 and
748 September 2017), in which the both are classified as the wet season (refer to **Supp. Fig. 1**). The cruise
749 in August 2016 represented the highest sampling frequency in order to obtain complete coverage of
750 representative regions, while the cruises in March and September 2017 were aimed to obtain seasonal
751 representatives for each region. Approximately 250 – 500 mL of water were filtered through 3.0 µm
752 pore size track-etched membranes (Nucleopore™, Whatman, Germany) via vacuum filtration. This was
753 referred to as the ‘Particulate-attached’ fraction. The filtrate from the 3.0 µm portion was collected in a
754 sterile glass bottle and subsequently filtered through 0.2 µm pore size track-etched membranes
755 (Nucleopore™, Whatman, Germany). The smaller fraction was referred to as ‘free-living’ fraction. A
756 total of 117 filters were recovered (1 x 3.0 µm was discarded due to contamination) and immediately
757 stored at -20 °C and sent to the Australian Centre for Ecogenomics (ACE), Brisbane for DNA extraction,
758 library preparation and processing utilizing the Illumina platform (Bentley et al., 2008) .

759

760

761 2.2 Illumina Sequencing and Bioinformatics Analyses

762 Initial upstream processes were carried out by the Australian Centre for Ecogenomics utilizing the ACE
763 mitag pipeline (ACE, 2016). The primers utilized were based on the V3 – V4 hypervariable regions of
764 the 16S rRNA gene. Briefly, fastq files generated from the Illumina platform were quality trimmed with
765 fastqc, primer sequences trimmed with Trimmomatic, and poor quality sequences removed using a
766 sliding window of 4 bases with an average base quality of more than 15. High quality sequences were
767 subsequently processed using the mothur (Schloss et al., 2009) pipeline. Sequences were aligned against
768 the SILVA database (Quast et al., 2013; Yilmaz et al., 2014), ‘pre.cluster’ command executed for
769 denoising, and chimeric sequences removed using the ‘chimera.vsearch’ function. Chimera-free 16s
770 rRNA bacterial gene sequences were taxonomically assigned against the EzTaxon database (Kim et al.,
771 2012) using the Naïve Bayesian classifier with a threshold of 80%. The quality-filtered sequences were
772 then clustered into operational taxonomic units (OTUs) at 97% similarity cutoff with singleton OTUs
773 being omitted. In order to reduce bias caused by variations in sample size, high-quality reads were
774 randomly subsampled to 923 reads per sample. Apart from the results and discussion shown for free-
775 living and particle-attached bacteria, the remaining discussion is based on the pooled results of both
776 components. The alpha diversity was calculated using the *estimate_richness* function embedded within
777 the *plot_richness* function found within the *phyloseq* package utilizing R (v.3.5.3). For the analyses of
778 potential functional genes, Phylogenetic Investigation of Communities by Reconstruction of
779 Unobserved States (PICRUSt, Langille et al., 2013) was utilized. The metagenomics prediction table
780 produced from PICRUSt was utilized to produce pathway abundance profiles using HUMAnN2
781 (Franzosa et al., 2018). It should be noted that the reconstructed functional genes were based on the
782 GreenGenes (DeSantis et al., 2006) database and not the EzTaxon database used for the phylogeny.
783 Raw sequences have been deposited with the NCBI BioSample database under BioProject ID
784 PRJNA565954.

785

786 2.3 Physico-chemical Data and Geochemical Analyses

787 Monthly precipitation for the period in-between the cruises (August 2016 to September 2017) were
788 obtained from the Tropical Rainfall Measuring Mission website (NASA, 2019) in order to gauge the
789 seasonality (wet or dry; see **Supp. Fig. 1**). In the laboratory, nutrients (Nitrate, NO₃⁻, Nitrite, NO₂⁻,
790 Ammonium, NH₄⁺, Phosphate, PO₄³⁻ and Silicate, SiO₄⁴⁻) were photometrically determined utilizing a
791 SKALAR Sanplus continuous flow analyser in the State Key Laboratory of Estuarine and Coastal
792 Research (SKLEC), Shanghai (details described in (Sia et al., 2019). NH₄⁺ and PO₄³⁻ were determined
793 manually following Grasshoff et al., (1999). The total dissolved nitrogen, TDN, and total dissolved
794 phosphate, TDP, were determined indirectly by obtaining the values for NO₃⁻ and PO₄³⁻ via oxidation
795 with alkaline-persulfate solution (Ebina et al., 1983). The concentrations of dissolved organic nitrogen
796 (DON) and dissolved organic phosphorus (DOP) are estimated by subtraction of DIN from TDN and
797 PO₄³⁻ from TDP, respectively. Belawai samples (2°13'47.16"N, 111°12'19.04"E) were used in an

798 incubation experiment to study the net primary productivity and respiration rate of the Rajang River.

799 Technical triplicates were incubated in both light and dark set-ups (refer to **Supp. Table 1** for details).

800

801

802 **2.4 Statistical Analyses and distLM model**

803 Ordination visualization, non-metric multidimensional scaling (NMDS, Kruskal-Wallis: Kruskal stress
804 formula: 1; minimum stress: 0.01), similarity analyses (ANOSIM) and coherence plots were executed
805 using PRIMER 7 (Clarke and Gorley, 2015) to determine if the various terrestrial source types or
806 different land use impacted the bacterial community. Permutational multivariate analysis of variance
807 (PERMANOVA) was used based on the Bray-Curtis dissimilarity of Hellinger Transformed
808 resemblance matrix to infer the impact of anthropogenic activities (land use) on the microbial
809 communities. By partitioning the community variation (using a Bray-Curtis dissimilarity matrix
810 resemblance), distance-based linear models (DistLM) were used to determine the extent of which the
811 bacterial community structure can be explained by environmental variables (Legendre and Anderson,
812 1999). Normalizing transformations of the environmental variables were carried out prior to execution
813 of DistLM analyses using the “Normalise Variables” function in the PRIMER 7 software. Hellinger
814 Transformed OTU abundance table was used as the response variable for the variation partition
815 analysis. The authors would like to note that the distLM models are based on only the August 2016 and
816 March 2017 cruise as there was a lack of physico-chemical data from the September 2017 cruise due to
817 malfunctioning equipment. Multi-collinearity between variables was tested utilizing the ‘Draftsman
818 Plot’ function in Primer 7 (Clarke and Gorley, 2006; Supp. Fig. 1). However, it is sufficient to draw
819 linkages between the major drivers of microbial communities between seasons as Mar 2017 and
820 September 2017 were considered wet seasons based on the average precipitation (see **Supp. Fig. 1**).

821

822 **3. Results**

823 **3.1 Clustering of Samples according to ANOSIM Global Test Scores**

824 A total of 74,690 high quality bacterial sequences were obtained from a total of 117 samples, with 200
825 to 2,615 sequence reads per sample. The sequences were clustered into 2,087 OTUs at the 97%
826 confidence interval. Instead of displaying bacterial diversity by station, bacterial communities were
827 grouped together according to the R scores obtained from the ANOSIM Global test, with the parameters
828 'cruise', 'source type' and 'land use' showing the highest scores (ANOSIM Global $R = 0.737$, $P < 0.001$,
829 **Table 1**). Furthermore, multi-variate analysis showed that the microbial community composition
830 differed among the different land use as well as site nested with land use and source type (**Table 2**).

831

832 **3.2 Shifts in bacterial community structure**

833 The NMDS graph (2D stress score: 0.18, **Fig. 2**) supported ANOSIM results by clustering samples
834 according to (i) source type and land use as well as (ii) cruises. The X axis (MDS1 scores) clearly
835 reflects changes in terms of salinity (river-sea continuum) while the Y axis (MDS2 scores) emulates the
836 different cruises. It is apparent that there were seasonal variations as shown from the lighter shade points,
837 representing the August 2016 (dry season) samples, compared to those with darker shades representing
838 both March 2017 and September 2017 (wet season) samples (**Fig. 2**). There were clear overlaps of
839 samples from mineral soil and freshwater peat origin. We also observed a gradual shift of samples from
840 mineral soils and freshwater peat towards brackish and then marine samples.

841

842 **3.3 Bacterial Distribution according to source type and cruise**

843 To further support that the four different source types support distinct bacterial communities, the
844 relative abundance was mapped into a percentage plot (**Fig. 3**).

845 The core microbial communities along the Rajang River-South China Sea continuum consist of
846 *Proteobacteria*, *Firmicutes*, *Actinobacteria*, *Bacteroidetes*, *Deinococcus-Thermus* and *Cyanobacteria*
847 in varying abundances (**Fig. 3, Supp. Fig. 4**), indicating high variation within the system. The phylum
848 *Deinococcus-Thermus* was abundant in freshwater peat and in mineral soils, albeit at a lesser extent
849 compared to freshwater peat (**Fig. 3**). Taking into consideration seasonality, the relative abundance (%)
850 of *Deinococcus-Thermus* drastically decreased in September 2017. Contrary, the abundance of
851 *Cyanobacteria* was greater within marine as well as brackish peat for the cruises of March 2017 and
852 September 2017 but not for August 2016. For the August 2016 cruise, *Cyanobacteria* were found
853 throughout all source types albeit at lower counts compared to the other cruises. Similar changes in
854 bacterial community were observed during different cruises but at different sections of the river. For
855 the freshwater peat and mineral soils, the cruises of August 2016 and March 2017 had greater
856 resemblance towards each other. Furthermore, there was a distinct split in terms of the bacterial
857 community composition for the four source types across all sampling cruises i.e. marine and brackish
858 peat had similar composition and freshwater peat and mineral soils had similar composition. In terms

859 of a river-sea continuum, the most apparent changes in the community composition were observed
860 during March 2017 which presented an almost step-wise change in bacterial community composition.

861

862 **3.4 Alpha Diversity Indices**

863 Based on the observed indices (**Fig. 4**), mineral soils generally had the highest counts of unique OTUs.
864 However, during the September 2017 cruise, the freshwater region had the highest values. Based on the
865 Chao1 indices, there was a significant effect of the source type on the observed richness ($p < 0.001$), with
866 increasing values from marine to mineral soils. In the March 2017 and September 2017 cruise, the
867 Chao1 indices were found to have greater variability as compared to the August 2017 cruise. For the
868 September 17 cruise, we observed increased values of Chao1 across the brackish peat, freshwater peat
869 as well as mineral soils. According to the Shannon indices, the diversity of the microbial communities
870 varied significantly along the different source types ($p < 0.001$). In the dry season the Shannon indices
871 were found to be higher than that of March 17 and September 2017 samples, except for the Brackish
872 peat September 2017 samples. In terms of the Simpson diversity indices, the August 2016 season was
873 found to have the higher values as compared to the March 2017 and September 2017 season.

874

875 Based on the effects of land use on the diversity indices (**Fig. 5**), the sites which are surrounded by
876 human settlements had higher observed indices (regardless of the cruise), with the exception of the
877 Shannon indices in August 2016. Samples surrounded by secondary forest had the second-highest
878 values with samples from August 2016 repeatedly higher than the other two cruises. There were
879 significant differences ($p < 0.001$) between samples from the coastal region with generally lower indices
880 compared to upstream samples (i.e. Human Settlement, Oil Palm and Sago Plantation, Oil Palm
881 Plantation and Secondary Forest).

882

883 **3.5 Functional Profile of Bacterial Communities**

884 Based on the potential KEGG pathways (**Fig. 6**), the functional profiles of the microbial communities
885 were predicted for the Aug 2016 and Mar 2017 samples. The main functions found were oxidative
886 phosphorylation (20.09%), carbon fixation pathways in prokaryotes (19.00%) and methane metabolism
887 (18.36%), respectively. This was then followed by nitrogen metabolism (11.50%), carbon fixation in
888 photosynthetic organisms (7.67%), inorganic ion transport and metabolism (5.68%). The remaining
889 functional groups were photosynthesis, sulphur metabolism, inositol phosphate metabolism,
890 phosphotransferase system (PTS), carbohydrate metabolism, phosphonate and phosphinate metabolism
891 and lastly mineral absorption (4.92%, 4.31%, 2.96%, 2.34%, 1.83%, 1.11% and 0.23%, respectively).
892 Clear differences were observed between source types and seasons and potential KEGG pathways
893 displayed similar composition among samples originating from either (i) marine and brackish peat, or
894 (ii) freshwater peat and mineral soil. In terms of gene abundances, the March 2017 samples (wet season)
895 were found to have higher gene abundances with the highest counts in brackish peat followed by marine

896 samples. However, marine samples in August 2016 displayed slightly higher gene counts compared to
897 the brackish peat.

898

899 **3.6 Distance-based Linear Model of bacterial communities and environmental parameters**

900 Marginal DistLM was performed in order to gauge the extent of physicochemical parameters or
901 environmental variables accounting for a compelling proportion of variation in the bacterial
902 communities. Significant vectors of environmental variables ($R_2 > 0.3892$, $P < 0.001$) were calculated
903 based on a linear model (DistLM) and plotted against the bacterial community composition (**Fig 7**).
904 Salinity was the single best predictor variable explaining bacterial community variation (15.27%),
905 followed by DIP (10.57%). The remaining physico-chemical parameters were dissolved oxygen (DO,
906 9.64%) and suspended particulate matter (SPM, 6.55%) whereas for the biogeochemical parameters,
907 Silicate (9.27%), DOP (8.04%), DON (6.37%), dissolved organic carbon (DOC, 5.27%) and dissolved
908 inorganic nitrogen (DIN, 4.29%) respectively made up the remaining variables (all variables $P = 0.001$,
909 except for DIN, $P = 0.002$).

910

911 The distLM model clustered samples from the August 2016 cruise separately from the March 2017
912 samples. Brackish peat, as well as marine samples from August 2016, correlated more strongly with
913 salinity, irrespective of land use. On the contrary, the March 2017 samples were found to cluster
914 separately with DO. In addition, the August 2016 mineral soil samples correlated with silicate.

915

916 **4. Discussion**

917 This study presents seasonal and spatial distribution of particulate-attached and free-living bacteria in
918 the longest river in Malaysia in an attempt to map the bacterial community composition of the water
919 column across several habitats with relation to the riparian zones and anthropogenic activities in a river-
920 to-sea continuum. Our dataset develops a comparison of the microbial community across two
921 dimensions: spatial biogeography from headwaters to the coastal zone as well as through time
922 (seasonally). The rich supporting dataset also allows us to assess underlying nutrient dynamics
923 influencing the microbial communities.

924

925 **4.1 General diversity of core bacterial communities along the Rajang river-South China Sea**
926 **continuum in comparison with global systems**

927 The majority of bacterial taxa were restricted to a relatively small number of assemblages. Dominant
928 phyla typically found in Malaysian peat swamps such as *Proteobacteria* (Kanokratana et al., 2011; Too
929 et al., 2018; Tripathi et al., 2016) are found throughout the Rajang river whereas *Acidobacteria* is not a
930 major phylum in the Rajang river. However, due to the heterogeneity of the Rajang River, substantial
931 shifts in OTU diversity were shown, while exhibiting successional changes in community composition
932 downstream. We observed abrupt shifts in terms of richness and diversity as well as bacterial
933 distribution, which were structured according to macro-scale source types. Staley et al. (2015) proposed
934 that variability in microbial communities were less due to the presence/absence but likely due to shifts
935 in relative abundance of OTUs. While there were shifts in the community composition, overlap between
936 the core microbiome (i.e. free-living and particle-attached portions) of samples were evident (**Supp.**
937 **Fig. 2, Supp. Fig. 8**). The similar bacterial community structure in terms of particle association was in
938 line with studies by Noble et al. (1997) and Hollibough et al. (2000) in the Chesapeake Bay (winter
939 season) and San Francisco Bay, respectively. Hollibough et al. (2000) demonstrated that the difference
940 or similarity of the particle association of bacterial community was due to the origin as well as
941 composition of the particles, particularly in marine snow or estuarine particles. In the aforementioned
942 study, there was limited metabolic divergence and similar communities between the estuarine turbidity
943 maxima and the river samples. Due to the short residence time, the rapid exchange of organisms likely
944 reduced the divergence of phylogenetic composition. The short residence time in the Rajang River
945 likely reflected a similar scenario to San Francisco Bay (Müller-Dum et al. 2019). When comparing
946 with other rivers, the predominance of the *Proteobacteria* phylum, especially within the brackish peat
947 region (**Fig. 3, Supp. Fig. 4**) was similar to a recent study on the Pearl River Delta (Chen et al., 2019).
948 In another study by Doherty et al. (2017) on the mainstem of the Amazon River (a blackwater influenced
949 river, similar to the Rajang River), *Actinobacteria* were much more abundant (25.8%) compared to the
950 Rajang River (11.95%).

951

952

953 4.2 Factors determining bacterial community composition

954 4.2.1 Spatial and environmental drivers

955 As shown in **Fig. 2**, it can be observed that there was a continual shift in microbial communities,
956 suggesting mixing of the microbial communities from the headwaters to the coast (Fortunato et al.,
957 2012) which has also been observed along the Upper Mississippi River (Staley et al., 2015) and along
958 the Danube River (Savio et al., 2015). The decrease in richness and evenness was similar to a study
959 conducted by Savio et al. (2015) in which the bacterial evenness and richness declined downriver, which
960 was in line with the River Continuum Concept (Vannote et al., 1980). The presence of peat did not
961 affect the alpha-diversity indices which was reflected in the shift in taxa occurring from freshwater
962 (which includes freshwater peat) towards the saline region (which includes brackish peat).

963

964 Salinity, DIP and DO are major environmental drivers of species distribution (Peter et al. 2011; Wilhelm
965 et al., 2015). In this study, marine and brackish peat samples correlated well with salinity. This was
966 neatly supported by the distribution of samples on the distLM fitted dbRDA graph (**Fig. 7**) whereby the
967 affinity for each of the samples correlates to the physical environment (e.g. the samples which group
968 along the salinity vector were the samples which correlate with the marine as well as brackish peat
969 region. The predominance of β -*Proteobacteria* in the freshwater region and the predominance of α - and
970 γ -*Proteobacteria* (**Supp. Fig. 3**) in the estuarine region is typical as the main group in seawaters
971 (Nogales et al., 2011) and similar to findings by Silveira et al. (2011) on the bacterioplankton
972 community along the river-to-ocean continuum from the Parnaioica River towards the Atlantic Ocean.
973 This shows that salinity exhibited a strong influence on the abundances of *Proteobacteria* and
974 *Firmicutes*. Furthermore, based on the linear model (**Fig. 7**), salinity was an important factor in driving
975 the shift in microbial communities (**Table 3**), similar to findings by Herlemann et al. (2011) along a
976 200 km salinity gradient in the Baltic Sea. The dispersal of taxa of microbial communities from fresh
977 to marine waters faces a strong barrier due to salinity (Fortunato and Crump, 2015), likely explaining
978 the reduced relative abundances of some taxa (**Fig. 3**). For example, *Chloroflexi* has a higher relative
979 abundance upstream while *Deinococcus-Thermus* shows lower relative abundance downstream. Such
980 dispersals are further influenced by transitional waters such as estuaries and plumes whereby the
981 microbial communities are exposed to rapidly changing physico-chemical conditions such as nutrients,
982 temperature as well as sporadic anthropogenic inputs (Crump et al., 2004).

983

984 While the distribution of the core microbial communities are indicative of the river-sea continuum, it is
985 noteworthy that several phyla were distinctly associated with specific source types. The distinct shift in
986 bacterial taxa for example from Freshwater to Brackish waters (and lack thereof between freshwater
987 peat and brackish peat; **Fig. 3**) indicates that peat did not have a significant effect on the distribution of
988 bacterial taxa. This was further supported by the fact that DOC (as a proxy for organic matter of peat
989 origin) only accounts for 5.27% of the community variation (**Table 3**). A study on blackwater rivers in

990 the Orinoco Basin, Venezuela (Castillo et al., 2004) showed that increased DOC resulted in higher
991 bacterial production, however, the change in bacterial production was not a reflection of its influence
992 on the community composition. This was supported based on a simple respiration experiment conducted
993 in Aug 2016 (**Supp. Table 1**) whereby the respiration rate (0.44 ± 0.16 g DO L⁻¹ d⁻¹) was higher than
994 that of the primary production rate (0.39 ± 0.08 g DO L⁻¹ d⁻¹).

995
996 Samples influenced by DO (**Fig. 7**) are from the estuarine region which showed an almost anoxic zone
997 (refer to **Supp. Fig. 6**). The low availability of oxygen was mirrored in higher counts (samples
998 belonging to the brackish peat category showed highest counts regardless of phyla as well as season;
999 **Supp. Fig. 4**). However, higher counts (particularly the phylum *Chloroflexi* and *Cyanobacteria* which
1000 are normally associated with production of oxygen via primary productivity) do not reflect higher
1001 primary production within this zone. Zones of coastal estuaries are usually deemed to have higher
1002 primary productivity, however, it can be inferred that the depletion in oxygen and higher pCO₂
1003 emissions (Müller-Dum et al., 2019) within the brackish peat region of the August 2016 campaign was
1004 a result of high bacterial productivity. This can be further supported by the high SPM as a proxy of
1005 turbidity of the brackish peat (**Supp. Fig. 6**) which may have resulted in the reduced primary
1006 productivity, which in turn can explain the lower DO values. As aforementioned earlier, the respiration
1007 rate (0.44 ± 0.16 g DO L⁻¹ d⁻¹) was higher than that of the primary production rate (0.39 ± 0.08 DO L⁻¹
1008 d⁻¹). This was similar to a study in the Scheldt River whereby the higher bacterial production occurred
1009 in the turbidity maxima together with the depletion of oxygen (Goosen et al., 1995).

1010

1011 **4.2.1.1 Functional potential of major taxa according to source types**

1012 In the Rajang River, the relative abundance of bacterial OTUs were higher in the estuary as well as
1013 marine region, reflecting that while the microbial communities are structured by salinity, the abundance
1014 was more a reflection of the nutrients available, especially in estuaries which exhibit circulation patterns
1015 which can result in localised nutrient-rich conditions (They et al., 2019). This was further supported by
1016 the higher relative abundance of oxidative phosphorylation genes as well as nitrogen metabolism within
1017 the brackish peat and further supported by Jiang et al. (2019) demonstrated through incubations studies
1018 whereby N transformations in the Rajang River estuary mixing zone was higher than in the Rajang
1019 River and coastal region. In a study done by Yang et al., (2013), the dominance of *Proteobacteria*
1020 influenced the nitrogen cycle via the processes of nitrification and denitrification, in which aeration
1021 would increase its abundance and result in higher mortality of *Cyanobacteria*. Hence, lower
1022 *Proteobacteria* abundance resulted in the higher abundance of *Cyanobacteria* which occur as evidently
1023 shown in **Fig. 3**. Furthermore, the higher presence of *Chloroflexi* (Ward et al., 2018) and *Cyanobacteria*
1024 (Guida et al., 2017) within the marine and brackish peat region indicated its probable role in carbon
1025 fixation as reflected by the higher gene counts (carbon fixation pathways in prokaryotes) in the marine
1026 and brackish peat regions as compared to the freshwater peat and mineral soil (**Fig. 6**). Furthermore,

1027 the presence of the genus *Sphingomonas*, a purple-sulphur bacteria which were able to utilize carbon
1028 dioxide (carbon fixation pathways in prokaryotes) and oxidation of hydrogen sulphide (sulphur
1029 metabolism) (Pfennig, 1975) (**Fig. 6**). In the case of *Firmicutes*, the higher abundance of *Firmicutes* in
1030 the brackish region was reflective of the overall production as opposed to selective growth of the
1031 particular source type, as *Firmicutes* were found throughout all four source types. The highest level of
1032 *Deinococcus-Thermus* (**Fig. 3**) was found in freshwater peat environments, indicating its preference for
1033 this environment. This was interesting to note that most studies on bacterial community composition
1034 show that the phylum *Deinococcus-Thermus* occurs in a higher abundance in extreme environments
1035 such as in hot springs (Zhang et al., 2018b) or in studies that are analogous for Mars (Joseph et al.,
1036 2019). In most of these studies, *Deinococcus-Thermus* was found in low abundance (e.g. 1% in
1037 Antarctic marine environments, 1.5% in hypersaline soils; Giudice and Azzaro, 2019; Vera-Gargallo et
1038 al., 2019) when compared to the Rajang River.

1039

1040 **4.2.2 Seasonality as a driver of microbial community composition**

1041 While the development of unique community structures was strongly influenced by spatial factors,
1042 seasonality also played a role. When taking into consideration the major genera, there was a
1043 fundamental shift in bacterial community composition along the continuum (**Fig 3, Fig. 4**). The second-
1044 most abundant taxon, *Proteobacteria* (β -*Proteobacteria*) peaked during seasons of high discharge. The
1045 same pattern of peaking during high discharge can be observed in the Rajang River with considerably
1046 higher relative abundance in the wet season (Fig. 3) which could be a result of the intense rainfall that
1047 led to the large input of freshwater (Silveira et al., 2011), and ultimately resulting in a “trickling” over
1048 microbial pattern from the freshwater to the brackish region. Furthermore, there was a distinct
1049 difference in terms of bacterial richness and diversity indices between the dry season (August 2016)
1050 and both wet seasons, with September 2017 having higher observed indices while the March 2017,
1051 while being a wet season as well had lower or variable observed indices. This difference in the two wet
1052 seasons could be the due to the different stages of phytoplankton bloom as mentioned earlier whereby
1053 the September 2017 was during an algal bloom while the March 2017 was after an algal bloom event.
1054 This was reflected in the Simpson index as well as the indices for September 2017 being lower than
1055 those of the August 2016 or March 2017 samples. Similarly, Zhou et al. (2018) demonstrated that the
1056 Simpson Indices for bacteria increased after the onset of an algal bloom (Brackish peat, September
1057 2017) whereas the Shannon indices was at the lowest (Brackish peat, March 2017) (when assuming that
1058 the region in which phytoplankton blooms occur was the brackish peat region). Overall, there was
1059 greater diversity (based on Shannon Indices) in the dry season (August 2016) than the wet seasons
1060 (March and September 2017) whereas there were greater OTUs in the wet season (Observed index).

1061

1062 Seasonal variability was also observed between the source types, particle association and down to the
1063 genus level (**Fig. 2, Supp. Fig. 2 and Supp. Fig. 5**). Based on the precipitation as an indicator of the

1064 seasonality, a probable “transitioning” phase was observed in the dry season (August 2016) with the
1065 microbial communities being more alike with the March 2017 samples (**Fig. 8**) when comparing both
1066 wet seasons (March 2017 and September 2017). Within the phylum rank (**Fig. 3**), the presence of
1067 *Cyanobacteria* during the March and September 2017 cruises indicates the influence of seasonality.
1068 However, while March 2017 and September 2017 were both considered to be wet seasons based on the
1069 precipitation, in terms of the relative abundance, there are considerable differences between the two
1070 cruises. The greater abundance of *Bacteroidetes* in March 2017 may be indicative of the community
1071 composition adjusting due to the processing of organic material caused by the higher cyanobacterial
1072 abundance in the September 2017 cruise. This was similar to a study by Pinhassi et al., (2004), in which
1073 the higher abundance of *Bacteroidetes* follows after an algal bloom. Moreover, the shifts in community
1074 composition from August 2016 to March 2017 and from March 2017 to September 2017 are indicative
1075 of the influence of seasonality. While March 2017 and September 2017 were similar in terms of climate,
1076 September 2017 had higher precipitation during that month, which led to higher run-off from the
1077 riparian region as compared with the March 2017 wet season. This could have led to the increase in
1078 cyanobacteria, which was also reflected increase of picoplankton size class during the wet season where
1079 it was hypothesized that the September 2017 might be more optimal for picoplankton proliferation
1080 (**Supp. Fig. 7**). Furthermore, in comparison, August 2016 and March 2017 were similar in terms of the
1081 proportion of the relative abundance of the community composition (**Fig. 3**).

1082

1083 **4.2.3 Land-use change and anthropogenic drivers**

1084 There has been little to no literature regarding the changes in microbial community composition as a
1085 result of land-use changes that occur within this region, particularly throughout the catchment area of
1086 the Rajang River. The results obtained from this study suggest that the run-off from anthropogenic
1087 activities alters the microbial community composition. The *Cytophaga-Flavobacterium-Bacteroidetes*
1088 group, or rather known as the CFB group, are commonly associated with humans (Weller et al., 2000),
1089 reflecting anthropogenic influences on the samples, especially within the brackish areas which has
1090 several human settlements and plantations. This was shown in the coherence plots in **Supp. Fig. 10** and
1091 **Supp. Fig. 11** whereby the CFB group in the *Bacteroidetes* phylum were more pronounced in areas
1092 with influence of oil palm plantations. Lee-Cruz et al. (2013) demonstrated that conversions of oil palm
1093 plantations from tropical forests are much more severe as compared to logged over forests in terms of
1094 bacterial community composition whereby logged over forests was shown to exhibit some resilience
1095 and resistance (to a certain extent). This was shown in the clustering of bacterial taxa adjacent to the oil
1096 palm plantation regardless of the source type (**Supp. Fig. 12**) in which the vector of *Flavobacteriia* fall
1097 under the samples of oil palm plantation in the brackish peat and to a certain extent, the vector of
1098 *Bacteroidia* along the oil palm plantation samples in the freshwater peat. Furthermore, among the
1099 identified possible pathogenic bacteria, according to Reza et al. (2018), the taxa *Flavobacterium* is a
1100 potential fish pathogen and is commonly found in freshwater habitats (Lee and Eom, 2017) as well as

1101 coastal pelagic zones (Eilers et al., 2001). In the Rajang river, it was the sixth most abundant class
1102 (**Supp. Fig. 4**). This is cause for concern as it was found to be high in the coastal regions as well as
1103 brackish regions where fisheries and fishing activities are concentrated.

1104

1105 Anthropogenic disturbances, in particular, settlements and logging (secondary forest) led to higher
1106 diversity indices (**Fig. 6**). On the contrary, sites surrounded by oil palm plantations displayed the lowest
1107 diversity indices, supporting results by Mishra et al. (2014) who found similar results in peatlands.
1108 Furthermore, the OTU overlapping of major anthropogenic activities (i.e. settlements and oil palm
1109 plantations) in **Supp. Fig. 9** reflected the possibility of higher abundance of generalists as compared to
1110 sensitive species (Jordaan et al., 2019) as microbial communities generally adapt to permanent stress
1111 events such as increased concentrations of inorganic or organic nutrients. In another study conducted
1112 by Fernandes et al. (2014), anthropogenically-influenced mangroves had 2x higher the amount of γ -
1113 *Proteobacteria* compared to pristine mangroves. This was similar to the March 2017 cruise along the
1114 Rajang River, whereby γ -*Proteobacteria* was the predominant class in the marine and brackish peat
1115 region along with the significant increase in *Bacteroidetes* as aforementioned, which can be associated
1116 to anthropogenic activities. On the other hand, during the dry season, the diversity of the “less-
1117 disturbed” region was higher than the disturbed regions. However, it should be noted that the coastal
1118 zone generally has the lowest richness and diversity amongst the other regions regardless of the presence
1119 or absence of anthropogenic activities. Hence, the extent of salinity intrusion may also result in the loss
1120 of diversity and richness of the microbial communities (Shen et al., 2018) in the Rajang River.

1121 **5. Conclusion**

1122 This study represents the first assessment of the microbial communities of the Rajang River, the longest
1123 river in Malaysia, expanding our knowledge of microbial ecology in tropical regions. The predominant
1124 taxa are *Proteobacteria* (50.29%), followed by *Firmicutes* (22.35%) and *Actinobacteria* (11.95%). The
1125 microbial communities were found to change according to the source type whereby distinct patterns
1126 were observed as a result of the changes in salinity along with variation of other biogeochemical
1127 parameters. Alpha diversity indices indicate that the microbial diversity was higher upstream as
1128 compared to the marine and estuarine regions whereas anthropogenic perturbations (regions with oil
1129 palm plantations and human settlements) led to increased richness but less diversity compared to those
1130 that were less affected by anthropogenic perturbations (coastal zone and secondary forest). The
1131 PICRUSt results showed differences between source types. Areas surrounded by oil palm plantations
1132 showed the lowest diversity and other signs of anthropogenic impacts included the presence of CFB-
1133 groups as well as probable algal blooms. In order to further gauge and substantiate the functional and
1134 metabolic capacity of the microbial communities within each specific source type, metaproteomics as
1135 well as metabolomics should be carried out along with mixing experiments. This approach will
1136 contribute towards a better understanding of the response of microbial communities to anthropogenic
1137 perturbations, as well as their role in degrading peat-related run-off from the surrounding riparian
1138 regions.

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1152

1153

1154 **References**

- 1155 ACE: 2016, 'ACE pipeline', [online] Available from:
1156 https://wiki.ecogenomic.org/doku.php?id=amplicon_pipeline_readme, 2016.
- 1157 Battin, T. J., Kaplan, L. A., Findlay, S., Hopkinson, C. S., Marti, E., Packman, A. I., Newbold, J. D.
1158 and Sabater, F.: Biophysical controls on organic carbon fluxes in fluvial networks, *Nat. Geosci.*, 1, 95,
1159 2008.
- 1160 Bentley, D. R., Balasubramanian, S., Swerdlow, H. P., Smith, G. P., Milton, J., Brown, C. G., Hall, K.
1161 P., Evers, D. J., Barnes, C. L., Bignell, H. R., Boutell, J. M., Bryant, J., Carter, R. J., Keira Cheetham,
1162 R., Cox, A. J., Ellis, D. J., Flatbush, M. R., Gormley, N. A., Humphray, S. J., Irving, L. J.,
1163 Karbelashvili, M. S., Kirk, S. M., Li, H., Liu, X., Maisinger, K. S., Murray, L. J., Obradovic, B., Ost,
1164 T., Parkinson, M. L., Pratt, M. R., Rasolonjatovo, I. M. J., Reed, M. T., Rigatti, R., Rodighiero, C.,
1165 Ross, M. T., Sabot, A., Sankar, S. V., Scally, A., Schroth, G. P., Smith, M. E., Smith, V. P., Spiridou,
1166 A., Torrance, P. E., Tzonev, S. S., Vermaas, E. H., Walter, K., Wu, X., Zhang, L., Alam, M. D.,
1167 Anastasi, C., Aniebo, I. C., Bailey, D. M. D., Bancarz, I. R., Banerjee, S., Barbour, S. G., Baybayan,
1168 P. A., Benoit, V. A., Benson, K. F., Bevis, C., Black, P. J., Boodhun, A., Brennan, J. S., Bridgham, J.
1169 A., Brown, R. C., Brown, A. A., Buermann, D. H., Bundu, A. A., Burrows, J. C., Carter, N. P.,
1170 Castillo, N., Chiara E. Catenazzi, M., Chang, S., Neil Cooley, R., Crake, N. R., Dada, O. O.,
1171 Diakoumakos, K. D., Dominguez-Fernandez, B., Earnshaw, D. J., Egbujor, U. C., Elmore, D. W.,
1172 Etchin, S. S., Ewan, M. R., Fedurco, M., Fraser, L. J., Fuentes Fajardo, K. V., Scott Furey, W.,
1173 George, D., Gietzen, K. J., Goddard, C. P., Golda, G. S., Granieri, P. A., Green, D. E., Gustafson, D.
1174 L., Hansen, N. F., Harnish, K., Haudenschild, C. D., Heyer, N. I., Hims, M. M., Ho, J. T., et al.:
1175 Accurate whole human genome sequencing using reversible terminator chemistry, *Nature*, 456, 53,
1176 2008.
- 1177 Besemer, K., Singer, G., Quince, C., Bertuzzo, E., Sloan, W. and Battin, T. J.: Headwaters are critical
1178 reservoirs of microbial diversity for fluvial networks, *Proc. R. Soc. B Biol. Sci.*, 280(1771),
1179 20131760, doi:10.1098/rspb.2013.1760, 2013.
- 1180 Bidle, K. D. and Fletcher, M.: Comparison of free-living and particle-associated bacterial
1181 communities in the chesapeake bay by stable low-molecular-weight RNA analysis., *Appl. Environ.*
1182 *Microbiol.*, 61(3), 944 LP – 952, 1995.
- 1183 Boughner, L. A. and Singh, P.: Microbial Ecology: Where are we now?, *Postdoc J.*, 4(11), 3–17,
1184 doi:10.14304/surya.jpr.v4n11.2, 2016.
- 1185 Bruland, K. W., Lohan, M. C., Aguilar-Islas, A. M., Smith, G. J., Sohst, B. and Baptista, A.: Factors
1186 influencing the chemistry of the near-field Columbia River plume: Nitrate, silicic acid, dissolved Fe,
1187 and dissolved Mn, *J. Geophys. Res. Ocean.*, 113(C2), doi:10.1029/2007JC004702, 2008.
- 1188 Cao, Y., Fanning, S., Proos, S., Jordan, K. and Srikumar, S.: A review on the applications of next
1189 generation sequencing technologies as applied to food-related microbiome studies, *Front. Microbiol.*,
1190 8(SEP), 1–16, doi:10.3389/fmicb.2017.01829, 2017.
- 1191 Castillo, M. M., Allan, J. D., Sinsabaugh, R. L. and Kling, G. W.: Seasonal and interannual variation
1192 of bacterial production in lowland rivers of the Orinoco basin, *Freshw. Biol.*, 49(11), 1400–1414,
1193 doi:10.1111/j.1365-2427.2004.01277.x, 2004.
- 1194 Chen, L., Tsui, M. M. P., Lam, J. C. W., Hu, C., Wang, Q., Zhou, B. and Lam, P. K. S.: Variation in
1195 microbial community structure in surface seawater from Pearl River Delta: Discerning the influencing
1196 factors, *Sci. Total Environ.*, 660, 136–144, doi:10.1016/j.scitotenv.2018.12.480, 2019.
- 1197 Clarke, K. and Gorley, R.: PRIMER version 7: User manual/tutorial., 2015.
- 1198 Clarke, K. R. and Gorley, R.: PRIMER v6: User Manual/Tutorial (Plymouth Routines in Multivariate

- 1199 Ecological Research), PRIMER-E, Plymouth., 2006.
- 1200 Cotner, J. B. and Biddanda, B. A.: Small Players, Large Role: Microbial Influence on Biogeochemical
1201 Processes in Pelagic Aquatic Ecosystems, *Ecosystems*, 5(2), 105–121, doi:10.1007/s10021-001-0059-
1202 3, 2002.
- 1203 Cottrell, M. T., Waidner, L. A., Yu, L. and Kirchman, D. L.: Bacterial diversity of metagenomic and
1204 PCR libraries from the Delaware River, *Environ. Microbiol.*, 7(12), 1883–1895, doi:10.1111/j.1462-
1205 2920.2005.00762.x, 2005.
- 1206 Crump, B. C. and Hobbie, J. E.: Synchrony and seasonality in bacterioplankton communities of two
1207 temperate rivers, *Limnol. Oceanogr.*, 50(6), 1718–1729, doi:10.4319/lo.2005.50.6.1718, 2005.
- 1208 Crump, B. C., Armbrust, E. V. and Baross, J. A.: Phylogenetic Analysis of Particle-Attached and
1209 Free-Living Bacterial Communities in the Columbia River, Its Estuary, and the Adjacent Coastal
1210 Ocean, *Appl. Environ. Microbiol.*, 65(7), 3192 LP – 3204, 1999.
- 1211 Crump, B. C., Hopkinson, C. S., Sogin, M. L. and Hobbie, J. E.: Microbial Biogeography along an
1212 Estuarine Salinity Gradient: Combined Influences of Bacterial Growth and Residence Time, *Appl.*
1213 *Environ. Microbiol.*, 70(3), 1494 LP – 1505, doi:10.1128/AEM.70.3.1494-1505.2004, 2004.
- 1214 Crump, B. C., Amaral-Zettler, L. A. and Kling, G. W.: Microbial diversity in arctic freshwaters is
1215 structured by inoculation of microbes from soils, *ISME J.*, 6(9), 1629–1639,
1216 doi:10.1038/ismej.2012.9, 2012.
- 1217 DeSantis, T. Z., Hugenholtz, P., Larsen, N., Rojas, M., Brodie, E. L., Keller, K., Huber, T., Dalevi,
1218 D., Hu, P. and Andersen, G. L.: Greengenes, a Chimera-Checked 16S rRNA Gene Database and
1219 Workbench Compatible with ARB, *Appl. Environ. Microbiol.*, 72(7), 5069 LP – 5072,
1220 doi:10.1128/AEM.03006-05, 2006.
- 1221 Dittmar, T., Fitznar, H. P. and Kattner, G.: Origin and biogeochemical cycling of organic nitrogen in
1222 the eastern Arctic Ocean as evident from D- and L-amino acids, *Geochim. Cosmochim. Acta*, 65(22),
1223 4103–4114, doi:https://doi.org/10.1016/S0016-7037(01)00688-3, 2001.
- 1224 Doherty, M., Yager, P. L., Moran, M. A., Coles, V. J., Fortunato, C. S., Krusche, A. V., Medeiros, P.
1225 M., Payet, J. P., Richey, J. E., Satinsky, B. M., Sawakuchi, H. O., Ward, N. D. and Crump, B. C.:
1226 Bacterial biogeography across the Amazon River-ocean continuum, *Front. Microbiol.*, 8(MAY), 1–
1227 17, doi:10.3389/fmicb.2017.00882, 2017.
- 1228 Ebina, J., Tsutsui, T. and Shirai, T.: Simultaneous determination of total nitrogen and total phosphorus
1229 in water using peroxodisulfate oxidation, *Water Res.*, 17(12), 1721–1726,
1230 doi:https://doi.org/10.1016/0043-1354(83)90192-6, 1983.
- 1231 Eilers, H., Pernthaler, J., Peplies, J., Glöckner, F. O., Gerdt, G. and Amann, R.: Isolation of Novel
1232 Pelagic Bacteria from the German Bight and Their Seasonal Contributions to Surface Picoplankton,
1233 *Appl. Environ. Microbiol.*, 67(11), 5134 LP – 5142, doi:10.1128/AEM.67.11.5134-5142.2001, 2001.
- 1234 Fernandes, S. O., Kirchman, D. L., Michotey, V. D., Bonin, P. C. and Lokabharathi, P. A.: Bacterial
1235 diversity in relatively pristine and anthropogenically-influenced mangrove ecosystems (Goa, India),
1236 *Brazilian J. Microbiol.*, 45(4), 1161–1171, doi:10.1590/S1517-83822014000400006, 2014.
- 1237 Findlay, S.: Stream microbial ecology, *J. North Am. Benthol. Soc.*, 29(1), 170–181, 2010.
- 1238 Fortunato, C. S. and Crump, B. C.: Microbial gene abundance and expression patterns across a river
1239 to ocean salinity gradient, *PLoS One*, 10(11), 1–22, doi:10.1371/journal.pone.0140578, 2015.
- 1240 Fortunato, C. S., Herfort, L., Zuber, P., Baptista, A. M. and Crump, B. C.: Spatial variability
1241 overwhelms seasonal patterns in bacterioplankton communities across a river to ocean gradient, *Isme*

- 1242 J., 6, 554, 2012.
- 1243 Fortunato, C. S., Eiler, A., Herfort, L., Needoba, J. A., Peterson, T. D. and Crump, B. C.: Determining
1244 indicator taxa across spatial and seasonal gradients in the Columbia River coastal margin, *Isme J.*, 7,
1245 1899, 2013.
- 1246 Franzosa, E. A., McIver, L. J., Rahnavard, G., Thompson, L. R., Schirmer, M., Weingart, G., Lipson,
1247 K. S., Knight, R., Caporaso, J. G., Segata, N. and Huttenhower, C.: Species-level functional profiling
1248 of metagenomes and metatranscriptomes., *Nat. Methods*, 15(11), 962–968, doi:10.1038/s41592-018-
1249 0176-y, 2018.
- 1250 Gaveau, D. L. A., Salim, M. and Arjasakusuma, S.: Deforestation and industrial plantations
1251 development in Borneo, , doi:doi/10.17528/CIFOR/DATA.00049, 2016.
- 1252 Giudice, A. and Azzaro, M.: *The Ecological Role of Micro-organisms in the Antarctic Environment*,
1253 edited by S. Castro-Sowinski, Springer Polar., 2019.
- 1254 Goosen, N. K., van Rijswijk, P. and Brockmann, U.: Comparison of heterotrophic bacterial
1255 production rates in early spring in the turbid estuaries of the Scheldt and the Elbe, *Hydrobiologia*,
1256 311(1–3), 31–42, doi:10.1007/BF00008569, 1995.
- 1257 Grasshoff, K., Kremling, K. and Ehrhardt, M.: *Methods of Seawater Analysis*, third ed., Wiley-VCH,
1258 Weinheim., 1999.
- 1259 Guida, B. S., Bose, M. and Garcia-Pichel, F.: Carbon fixation from mineral carbonates, *Nat.*
1260 *Commun.*, 8(1), 1–6, doi:10.1038/s41467-017-00703-4, 2017.
- 1261 Hall, R. O., Baker, M. A., Rosi-Marshall, E. J., Tank, J. L. and Newbold, J. D.: Solute-specific scaling
1262 of inorganic nitrogen and phosphorus uptake in streams, *Biogeosciences*, 10(11), 7323–7331,
1263 doi:10.5194/bg-10-7323-2013, 2013.
- 1264 Herlemann, D. P. R., Labrenz, M., Jürgens, K., Bertilsson, S., Waniek, J. J. and Andersson, A. F.:
1265 Transitions in bacterial communities along the 2000 km salinity gradient of the Baltic Sea, *ISME J.*,
1266 5(10), 1571–1579, doi:10.1038/ismej.2011.41, 2011.
- 1267 Hollibough, J., S. Wong, P. and Murrell, M.: Similarity of particle-associated and free-living bacterial
1268 communities in northern San Francisco Bay, California., 2000.
- 1269 Hunt, D. E. and Ward, C. S.: A network-based approach to disturbance transmission through
1270 microbial interactions , *Front. Microbiol.* , 6, 1182, 2015.
- 1271 Jiang, S., Müller, M., Jin, J., Wu, Y., Zhu, K., Zhang, G., Mujahid, A., Rixen, T., Muhamad, M. F.,
1272 Sia, E. S. A., Jang, F. H. A. and Zhang, J.: Dissolved inorganic nitrogen in a tropical estuary at
1273 Malaysia: transport and transformation, *Biogeosciences Discuss.*, (February), 1–27, doi:10.5194/bg-
1274 2019-7, 2019.
- 1275 Jordaan, K., Comeau, A. M., Khasa, D. P. and Bezuidenhout, C. C.: An integrated insight into the
1276 response of bacterial communities to anthropogenic contaminants in a river: A case study of the
1277 Wonderfonteinspruit catchment area, South Africa, *PLoS One*, 14(5), e0216758, 2019.
- 1278 Joseph, R. G., Dass, R. S., Rizzo, V., Cantasano, N. and Bianciardi, G.: Evidence of Life on Mars ?, ,
1279 1, 40–81, 2019.
- 1280 Kan, J.: Storm Events Restructured Bacterial Community and Their Biogeochemical Potentials, *J.*
1281 *Geophys. Res. Biogeosciences*, 123(7), 2257–2269, doi:10.1029/2017JG004289, 2018.
- 1282 Kanokratana, P., Uengwetwanit, T., Rattanachomsri, U., Bunternngsook, B., Nimchua, T.,
1283 Tangphatsornruang, S., Plengvidhya, V., Champreda, V. and Eurwilaichitr, L.: Insights into the

- 1284 Phylogeny and Metabolic Potential of a Primary Tropical Peat Swamp Forest Microbial Community
1285 by Metagenomic Analysis, *Microb. Ecol.*, 61(3), 518–528, doi:10.1007/s00248-010-9766-7, 2011.
- 1286 Kim, O.-S., Cho, Y.-J., Lee, K., Yoon, S.-H., Kim, M., Na, H., Park, S.-C., Jeon, Y. S., Lee, J.-H., Yi,
1287 H., Won, S. and Chun, J.: Introducing EzTaxon-e: a prokaryotic 16S rRNA gene sequence database
1288 with phylotypes that represent uncultured species., *Int. J. Syst. Evol. Microbiol.*, 62(Pt 3), 716–21,
1289 doi:10.1099/ijs.0.038075-0, 2012.
- 1290 Kolmakova, O. V., Gladyshev, M. I., Rozanov, A. S., Peltek, S. E. and Trusova, M. Y.: Spatial
1291 biodiversity of bacteria along the largest Arctic river determined by next-generation sequencing,
1292 *FEMS Microbiol. Ecol.*, 89(2), 442–450, doi:10.1111/1574-6941.12355, 2014.
- 1293 Konopka, A.: What is microbial community ecology, *ISME J.*, 3(11), 1223–1230,
1294 doi:10.1038/ismej.2009.88, 2009.
- 1295 Ladau, J., Sharpton, T. J., Finucane, M. M., Jospin, G., Kembel, S. W., O’Dwyer, J., Koeppl, A. F.,
1296 Green, J. L. and Pollard, K. S.: Global marine bacterial diversity peaks at high latitudes in winter,
1297 *ISME J.*, 7(9), 1669–1677, doi:10.1038/ismej.2013.37, 2013.
- 1298 Langille, M. G. I., Zaneveld, J., Caporaso, J. G., McDonald, D., Knights, D., Reyes, J. A., Clemente,
1299 J. C., Burkepile, D. E., Vega Thurber, R. L., Knight, R., Beiko, R. G. and Huttenhower, C.: Predictive
1300 functional profiling of microbial communities using 16S rRNA marker gene sequences, *Nat.*
1301 *Biotechnol.*, 31(9), 814–821, doi:10.1038/nbt.2676, 2013.
- 1302 Lee-Cruz, L., Edwards, D. P., Tripathi, B. M. and Adams, J. M.: Impact of Logging and Forest
1303 Conversion to Oil Palm Plantations on Soil Bacterial Communities in Borneo, *Appl. Environ.*
1304 *Microbiol.*, 79(23), 7290–7297, doi:10.1128/aem.02541-13, 2013.
- 1305 Lee, S.-Y. and Eom, Y.-B.: Analysis of Microbial Composition Associated with Freshwater and
1306 Seawater, *Biomed. Sci. Lett.*, 22(4), 150–159, doi:10.15616/bsl.2016.22.4.150, 2017.
- 1307 Legendre, P. and Anderson, M. J.: DISTANCE-BASED REDUNDANCY ANALYSIS: TESTING
1308 MULTISPECIES RESPONSES IN MULTIFACTORIAL ECOLOGICAL EXPERIMENTS, *Ecol.*
1309 *Monogr.*, 69(1), 1–24, doi:10.1890/0012-9615(1999)069[0001:DBRATM]2.0.CO;2, 1999.
- 1310 Lemke, M. J., Lienau, E. K., Rothe, J., Pagioro, T. A., Rosenfeld, J. and Desalle, R.: Description of
1311 freshwater bacterial assemblages from the upper Paraná river floodpulse system, Brazil, *Microb.*
1312 *Ecol.*, 57(1), 94–103, doi:10.1007/s00248-008-9398-3, 2009.
- 1313 Liao, H., Yu, K., Duan, Y., Ning, Z., Li, B., He, L. and Liu, C.: Profiling microbial communities in a
1314 watershed undergoing intensive anthropogenic activities, *Sci. Total Environ.*, 647, 1137–1147,
1315 doi:https://doi.org/10.1016/j.scitotenv.2018.08.103, 2019.
- 1316 Lozupone, C. A. and Knight, R.: Global patterns in bacterial diversity, *Proc. Natl. Acad. Sci.*, 104(27),
1317 11436–11440, doi:10.1073/pnas.0611525104, 2007.
- 1318 Madsen, E. L.: Microorganisms and their roles in fundamental biogeochemical cycles, *Curr. Opin.*
1319 *Biotechnol.*, 22(3), 456–464, doi:10.1016/j.copbio.2011.01.008, 2011.
- 1320 Martin, P., Cherukuru, N., Tan, A. S. Y., Sanwlani, N., Mujahid, A. and Müller, M.: Distribution and
1321 cycling of terrigenous dissolved organic carbon in peatland-draining rivers and coastal waters of
1322 Sarawak, Borneo, *Biogeosciences*, 15(22), 6847–6865, doi:10.5194/bg-15-6847-2018, 2018.
- 1323 Miettinen, J., Shi, C. and Liew, S. C.: Land cover distribution in the peatlands of Peninsular Malaysia,
1324 Sumatra and Borneo in 2015 with changes since 1990, *Glob. Ecol. Conserv.*, 6, 67–78,
1325 doi:https://doi.org/10.1016/j.gecco.2016.02.004, 2016.
- 1326 Mishra, S., Lee, W. A., Hooijer, A., Reuben, S., Sudiana, I. M., Idris, A. and Swarup, S.: Microbial

- 1327 and metabolic profiling reveal strong influence of water table and land-use patterns on classification
1328 of degraded tropical peatlands, *Biogeosciences*, 11(7), 1727–1741, doi:10.5194/bg-11-1727-2014,
1329 2014.
- 1330 Müller-Dum, D., Warneke, T., Rixen, T., Müller, M., Baum, A., Christodoulou, A., Oakes, J., Eyre,
1331 B. D. and Notholt, J.: Impact of peatlands on carbon dioxide (CO₂) emissions from the Rajang River
1332 and Estuary, Malaysia, *Biogeosciences*, 16(1), 17–32, doi:10.5194/bg-16-17-2019, 2019.
- 1333 NASA: Tropical Rainfall Measuring Mission, [online] Available from: <https://pmm.nasa.gov/TRMM>,
1334 2019.
- 1335 Newton, R. J., Jones, S. E., Eiler, A., McMahon, K. D. and Bertilsson, S.: A guide to the natural
1336 history of freshwater lake bacteria, *Microbiol. Mol. Biol. Rev.*, 75(1), 14–49,
1337 doi:10.1128/MMBR.00028-10, 2011.
- 1338 Noble, P. A., Bidle, K. D. and Fletcher, M.: Natural Microbial Community Compositions Compared
1339 by a Back-Propagating Neural Network and Cluster Analysis of 5S rRNA, *Appl. Environ. Microbiol.*,
1340 63(5), 1762–1770, 1997.
- 1341 Nogales, B., Lanfranconi, M. P., Piña-Villalonga, J. M. and Bosch, R.: Anthropogenic perturbations
1342 in marine microbial communities, *FEMS Microbiol. Rev.*, 35(2), 275–298, doi:10.1111/j.1574-
1343 6976.2010.00248.x, 2011.
- 1344 Peter, H., Ylla, I., Gudasz, C., Romaní, A. M., Sabater, S. and Tranvik, L. J.: Multifunctionality and
1345 Diversity in Bacterial Biofilms, *PLoS One*, 6(8), e23225, 2011.
- 1346 Pfennig, N.: The phototrophic bacteria and their role in the sulfur cycle, *Plant Soil*, 43(1), 1–16,
1347 doi:10.1007/BF01928472, 1975.
- 1348 Pinhassi, J., Sala, M. M., Havskum, H., Peters, F., Guadayol, Ò., Malits, A. and Marrasé, C.: Changes
1349 in Bacterioplankton Composition under Different Phytoplankton Regimens, *Appl. Environ.*
1350 *Microbiol.*, 70(11), 6753 LP – 6766, doi:10.1128/AEM.70.11.6753-6766.2004, 2004.
- 1351 Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., Peplies, J. and Glöckner, F. O.:
1352 The SILVA ribosomal RNA gene database project: Improved data processing and web-based tools,
1353 *Nucleic Acids Res.*, 41(D1), 590–596, doi:10.1093/nar/gks1219, 2013.
- 1354 Raymond, P. A., Hartmann, J., Lauerwald, R., Sobek, S., McDonald, C., Hoover, M., Butman, D.,
1355 Striegl, R., Mayorga, E., Humborg, C., Kortelainen, P., Dürr, H., Meybeck, M., Ciais, P. and Guth, P.:
1356 Global carbon dioxide emissions from inland waters, *Nature*, 503(7476), 355–359,
1357 doi:10.1038/nature12760, 2013.
- 1358 Read, D. S., Gweon, H. S., Bowes, M. J., Newbold, L. K., Field, D., Bailey, M. J. and Griffiths, R. I.:
1359 Catchment-scale biogeography of riverine bacterioplankton, *ISME J.*, 9(2), 516–526,
1360 doi:10.1038/ismej.2014.166, 2015.
- 1361 Reza, M. S., Mizusawa, N., Kumano, A., Oikawa, C., Ouchi, D., Kobiyama, A., Yamada, Y., Ikeda,
1362 Y., Ikeda, D., Ikeo, K., Sato, S., Ogata, T., Kudo, T., Jimbo, M., Yasumoto, K., Yoshitake, K. and
1363 Watabe, S.: Metagenomic analysis using 16S ribosomal RNA genes of a bacterial community in an
1364 urban stream, the Tama River, Tokyo, *Fish. Sci.*, 84(3), 563–577, doi:10.1007/s12562-018-1193-6,
1365 2018.
- 1366 Sa’adi, Z., Shahid, S., Ismail, T., Chung, E. S. and Wang, X. J.: Distributional changes in rainfall and
1367 river flow in Sarawak, Malaysia, *Asia-Pacific J. Atmos. Sci.*, 53(4), 489–500, doi:10.1007/s13143-
1368 017-0051-2, 2017.
- 1369 Savio, D., Sinclair, L., Ijaz, U. Z., Parajka, J., Reischer, G. H., Stadler, P., Blaschke, A. P., Blöschl,
1370 G., Mach, R. L., Kirschner, A. K. T., Farnleitner, A. H. and Eiler, A.: Bacterial diversity along a

- 1371 2600km river continuum, *Environ. Microbiol.*, 17(12), 4994–5007, doi:10.1111/1462-2920.12886,
1372 2015.
- 1373 Schloss, P. D., Westcott, S. L., Ryabin, T., Hall, J. R., Hartmann, M., Hollister, E. B., Lesniewski, R.
1374 A., Oakley, B. B., Parks, D. H., Robinson, C. J., Sahl, J. W., Stres, B., Thallinger, G. G., Van Horn,
1375 D. J. and Weber, C. F.: Introducing mothur: Open-Source, Platform-Independent, Community-
1376 Supported Software for Describing and Comparing Microbial Communities, *Appl. Environ.*
1377 *Microbiol.*, 75(23), 7537 LP – 7541, doi:10.1128/AEM.01541-09, 2009.
- 1378 Shen, D., Langenheder, S. and Jürgens, K.: Dispersal modifies the diversity and composition of active
1379 bacterial communities in response to a salinity disturbance, *Front. Microbiol.*, 9(SEP), 1–13,
1380 doi:10.3389/fmicb.2018.02188, 2018.
- 1381 Sia, E. S. A., Zhang, J., Jiang, S., Zhu, Z., Carrasco, G., Holt Jang, F., Mujahid, A. and Müller, M.:
1382 Behaviour of Dissolved Phosphorus with the associated nutrients in relation to phytoplankton biomass
1383 of the Rajang River-South China Sea continuum, *Biogeosciences Discuss.*, 1–35, doi:10.5194/bg-
1384 2019-219, 2019.
- 1385 Silveira, C. B., Vieira, R. P., Cardoso, A. M., Paranhos, R., Albano, R. M. and Martins, O. B.:
1386 Influence of salinity on bacterioplankton communities from the Brazilian rain forest to the coastal
1387 Atlantic Ocean, *PLoS One*, 6(3), 1–9, doi:10.1371/journal.pone.0017789, 2011.
- 1388 Smith, S. V and Hollibaugh, J. T.: Coastal metabolism and the oceanic organic carbon balance, *Rev.*
1389 *Geophys.*, 31(1), 75–89, doi:10.1029/92RG02584, 1993.
- 1390 Staley, C., Unno, T., Gould, T. J., Jarvis, B., Phillips, J., Cotner, J. B. and Sadowsky, M. J.:
1391 Application of Illumina next-generation sequencing to characterize the bacterial community of the
1392 Upper Mississippi River, *J. Appl. Microbiol.*, 115(5), 1147–1158, doi:10.1111/jam.12323, 2013.
- 1393 Staley, C., Gould, T. J., Wang, P., Phillips, J., Cotner, J. B. and Sadowsky, M. J.: Species sorting and
1394 seasonal dynamics primarily shape bacterial communities in the Upper Mississippi River, *Sci. Total*
1395 *Environ.*, 505, 435–445, doi:10.1016/j.scitotenv.2014.10.012, 2015.
- 1396 They, N. H., Marins, L. F., Möller, O. O. and Abreu, P. C.: High bacterial activity in nutrient rich
1397 saltwater: Evidence from the uncoupling between salinity and nutrients in the Patos Lagoon estuary,
1398 *Estuar. Coast. Shelf Sci.*, 216(July 2017), 148–156, doi:10.1016/j.ecss.2018.09.001, 2019.
- 1399 Too, C. C., Keller, A., Sickel, W., Lee, S. M. and Yule, C. M.: Microbial Community Structure in a
1400 Malaysian Tropical Peat Swamp Forest: The Influence of Tree Species and Depth, *Front. Microbiol.*,
1401 9(December), 1–13, doi:10.3389/fmicb.2018.02859, 2018.
- 1402 Tripathi, B. M., Song, W., Slik, J. W. F., Sukri, R. S., Jaafar, S., Dong, K. and Adams, J. M.:
1403 Distinctive Tropical Forest Variants Have Unique Soil Microbial Communities, But Not Always Low
1404 Microbial Diversity, *Front. Microbiol.*, 7, 376, doi:10.3389/fmicb.2016.00376, 2016.
- 1405 Vannote, R. L., Minshall, G. W., Cummins, K. W., Sedell, J. R. and Cushing, C. E.: The River
1406 Continuum Concept, *Can. J. Fish. Aquat. Sci.*, 37(1), 130–137, doi:10.1139/f80-017, 1980.
- 1407 Vera-Gargallo, B., Chowdhury, T. R., Brown, J., Fansler, S. J., Durán-Viseras, A., Sánchez-Porro, C.,
1408 Bailey, V. L., Jansson, J. K. and Ventosa, A.: Spatial distribution of prokaryotic communities in
1409 hypersaline soils, *Sci. Rep.*, 9(1), 1–12, doi:10.1038/s41598-018-38339-z, 2019.
- 1410 Wang, B., Huang, F., Wu, Z., Yang, J., Fu, X. and Kikuchi, K.: Multi-scale climate variability of the
1411 South China Sea monsoon: A review, *Dyn. Atmos. Ocean.*, 47(1–3), 15–37,
1412 doi:10.1016/j.dynatmoce.2008.09.004, 2009.
- 1413 Wang, P., Clemens, S., Beaufort, L., Braconnot, P., Ganssen, G., Jian, Z., Kershaw, P. and Sarnthein,
1414 M.: Evolution and variability of the Asian monsoon system: State of the art and outstanding issues,

- 1415 Quat. Sci. Rev., 24(5–6), 595–629, doi:10.1016/j.quascirev.2004.10.002, 2005.
- 1416 Ward, L. M., Hemp, J., Shih, P. M., McGlynn, S. E. and Fischer, W. W.: Evolution of phototrophy in
1417 the Chloroflexi phylum driven by horizontal gene transfer, *Front. Microbiol.*, 9(FEB), 1–16,
1418 doi:10.3389/fmicb.2018.00260, 2018.
- 1419 Weller, R., Glöckner, F. and Amann, R.: 16S rRNA-Targeted Oligonucleotide Probes for the in situ
1420 Detection of Members of the Phylum Cytophaga-Flavobacterium-Bacteroides., 2000.
- 1421 Welti, N., Striebel, M., Ulseth, A. J., Cross, W. F., DeVilbiss, S., Glibert, P. M., Guo, L., Hirst, A. G.,
1422 Hood, J., Kominoski, J. S., MacNeill, K. L., Mehring, A. S., Welter, J. R. and Hillebrand, H.:
1423 Bridging food webs, ecosystem metabolism, and biogeochemistry using ecological stoichiometry
1424 theory, *Front. Microbiol.*, 8(JUL), 1–14, doi:10.3389/fmicb.2017.01298, 2017.
- 1425 Wetlands International: Flooding projections from elevation and subsidence models for oil palm
1426 plantations in Rajang Delta peatlands, Sarawak, Malaysia., *Deltares* [online] Available from:
1427 [https://www.deltares.nl/app/uploads/2015/06/Rajang-Delta-Peatland-Subsidence-Flooding-Deltares-](https://www.deltares.nl/app/uploads/2015/06/Rajang-Delta-Peatland-Subsidence-Flooding-Deltares-2015.pdf)
1428 [2015.pdf](https://www.deltares.nl/app/uploads/2015/06/Rajang-Delta-Peatland-Subsidence-Flooding-Deltares-2015.pdf) (Accessed 15 May 2019), 2015.
- 1429 Wilhelm, L., Besemer, K., Fragner, L., Peter, H., Weckwerth, W. and Battin, T. J.: Altitudinal
1430 patterns of diversity and functional traits of metabolically active microorganisms in stream biofilms,
1431 *ISME J.*, 9(11), 2454–2464, doi:10.1038/ismej.2015.56, 2015.
- 1432 Yang, X., Xie, P., Ma, Z., Wang, Q., Fan, H. and Shen, H.: Decrease of NH₄⁺-N by bacterioplankton
1433 accelerated the removal of cyanobacterial blooms in aerated aquatic ecosystem, *J. Environ. Sci.*
1434 (China), 25(11), 2223–2228, doi:10.1016/S1001-0742(12)60282-4, 2013.
- 1435 Yilmaz, P., Parfrey, L. W., Yarza, P., Gerken, J., Pruesse, E., Quast, C., Schweer, T., Peplies, J.,
1436 Ludwig, W. and Glöckner, F. O.: The SILVA and “all-species Living Tree Project (LTP)” taxonomic
1437 frameworks, *Nucleic Acids Res.*, 42(D1), 643–648, doi:10.1093/nar/gkt1209, 2014.
- 1438 Yule, C. M., Lim, Y. Y. and Lim, T. Y.: Degradation of Tropical Malaysian Peatlands Decreases
1439 Levels of Phenolics in Soil and in Leaves of *Macaranga pruinosa* , *Front. Earth Sci.* , 4, 45, 2016.
- 1440 Zeglin, L. H.: Stream microbial diversity in response to environmental changes: Review and synthesis
1441 of existing research, *Front. Microbiol.*, 6(MAY), 1–15, doi:10.3389/fmicb.2015.00454, 2015.
- 1442 Zhang, C., Dang, H., Azam, F., Benner, R., Legendre, L., Passow, U., Polimene, L., Robinson, C.,
1443 Suttle, C. A. and Jiao, N.: Evolving paradigms in biological carbon cycling in the ocean, *Natl. Sci.*
1444 *Rev.*, 5(4), 481–499, doi:10.1093/nsr/nwy074, 2018a.
- 1445 Zhang, Y., Huang, L., Jiang, H. and Wu, G.: Hyperthermophilic Anaerobic Nitrate-Dependent Fe(II)
1446 Oxidization by Tibetan Hot Spring Microbiota and the Formation of Fe Minerals, *Geomicrobiol. J.*,
1447 36(1), 30–41, doi:10.1080/01490451.2018.1492047, 2018b.
- 1448 Zhou, J., Richlen, M. L., Sehein, T. R., Kulis, D. M., Anderson, D. M. and Cai, Z.: Microbial
1449 community structure and associations during a marine dinoflagellate bloom, *Front. Microbiol.*,
1450 9(JUN), 1–21, doi:10.3389/fmicb.2018.01201, 2018.
- 1451 Zwart, G., Crump, B. C., Agterveld, M. P. K. and Hagen, F.: Typical freshwater bacteria: an analysis
1452 of available 16S rRNA gene sequences from plankton of lakes and rivers , *Aquat. Microb. Ecol.*,
1453 28(2), 141–155, 2002.
- 1454

1455 **Tables**

1456

1457 **Table 1:** ANOSIM Global Test scores based on various parameters

| Parameters tested, 999 permutations, random sampling | ANOSIM Global Test, R | P value |
|---|------------------------------|----------------|
| Cruise (Wet/Dry season) | 0.439 | 0.001 |
| Source Type | 0.422 | 0.001 |
| Land use | 0.182 | 0.001 |
| Particle Association | 0.037 | 0.001 |
| Source Type, Land use | 0.415 | 0.001 |
| Cruise, Source Type, Particle Association, | 0.708 | 0.001 |
| Cruise, Source Type, Land use | 0.737 | 0.001 |

1458

1459 **Table 2:** Results of permutational multivariate analysis of variance (PERMANOVA)

| Parameters tested, 9999 permutations, permutation of residuals under a reduced model | <i>df</i> | <i>F</i> | <i>P</i> |
|---|------------------|-----------------|-----------------|
| Land Use | 7 | 1.54 | 0.0016 |
| Site (nested with land use and particle attached) | 33 | 2.27 | 0.0001 |
| Site (nested with source type and land use) | 13 | 2.60 | 0.0001 |

1460

df represents degrees of freedom.

1461

1462 **Table 3:** Proportion of combined community variation based on marginal DistLM test that is explained
1463 by each predictor variable using two cruises (August and March 2017)

| Category | Variable | Pseudo-F | <i>P</i>-value | Proportion explained (%) |
|------------------------------------|-----------------|-----------------|-----------------------|---------------------------------|
| Physico-chemical parameters | Salinity | 9.6128 | 0.001 | 13.42 |
| | DO | 6.6151 | 0.001 | 9.64 |
| | SPM | 4.3486 | 0.001 | 6.55 |
| Biogeochemical parameters | DIP | 4.2218 | 0.001 | 10.57 |
| | Silicate | 9.269 | 0.001 | 9.27 |
| | DOP | 5.4246 | 0.001 | 8.04 |
| | DOC | 3.4495 | 0.001 | 5.27 |
| | DON | 4.2218 | 0.001 | 6.37 |

1464

1465 **Figure Captions**

1466 **Fig. 1:** Location of Rajang River within Sarawak, Malaysia (inset). (A) shows the stations sampled
1467 during three (3) different cruises; August 2016 (red triangles), March 2017 (blue circles) and September
1468 2017 (cyan diamonds). (B) GIS data from 2010 (Sarawak Geoportal, 2018) indicating various forest
1469 types. Red colour represents non-forest areas (2010), yellow represents non-forest areas (2013), light
1470 green represents primary forests, teal represents secondary forests whereas dark green represents
1471 potential peat swamp forests.(C) Digitized NREB map obtained from Wetlands International, (2015).
1472 The map shows the plantation cover as determined from Landsat showing licensed oil palm and sago
1473 plantations (licensed).

1474

1475 **Fig. 2:** Non-metric Multi-dimensional Scaling (NMDS) graph of samples according to cruise, source
1476 type as well as land use.

1477

1478 **Fig. 3:** Relative abundance (%) of dominant bacterial (at phylum level, top 10) along the various
1479 source types (Marine, Brackish Peat, Freshwater Peat, Mineral Soils) across 3 cruises/seasons

1480

1481 **Fig. 4:** The calculated α -diversity indices (Observed, Chao1, Shannon, Simpson and Inverse Simpson)
1482 of the four different source type along the salinity gradient.

1483

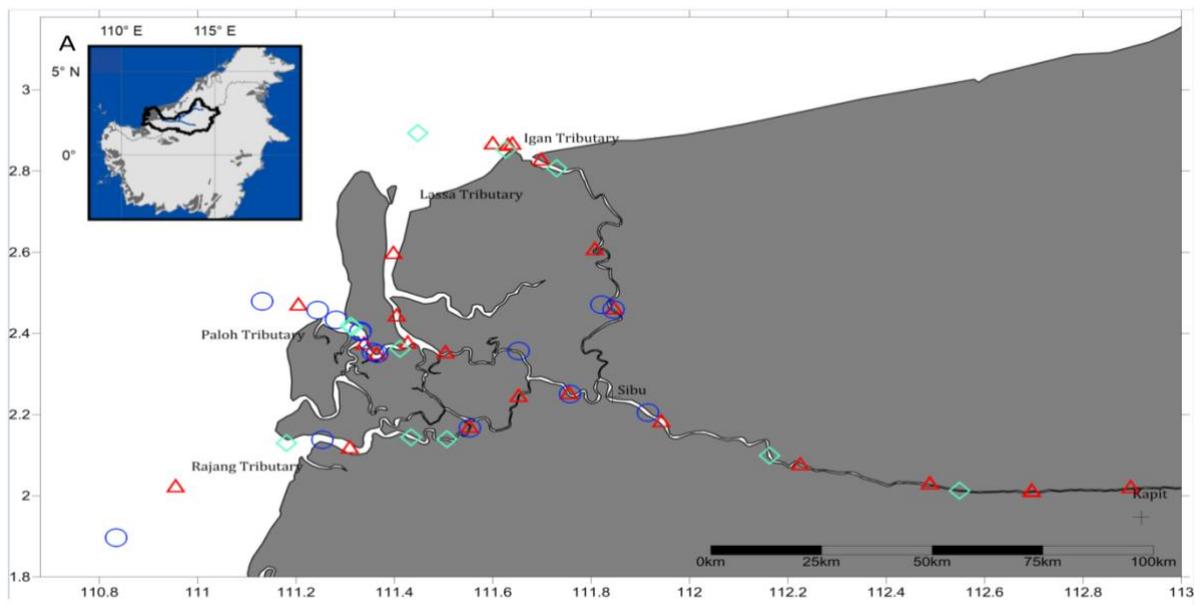
1484 **Fig. 5:** The calculated α -diversity indices (Observed, Chao1, Shannon, Simpson and Inverse Simpson)
1485 of the Land Use types (Coastal Zone, Coastal Zone with Plantation (OP) influence) Coastal Zone with
1486 Plantation (Sago and Oil Palm influence), Human Settlement, Oil Palm and Sago mixed Plantation, Oil
1487 Palm Plantation and Secondary Forest)

1488 **Fig. 6:** The relative abundance of predicted functional profiles in the four source types across two
1489 seasons based on KEGG Pathways

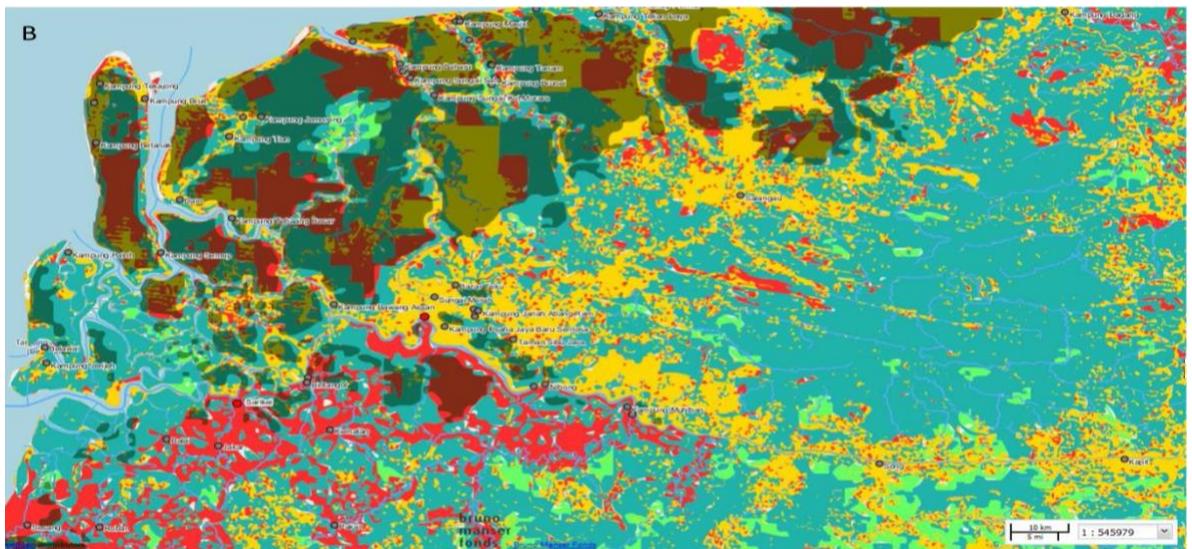
1490

1491 **Fig. 7:** Distance-based Redundancy Analysis (dbRDA) plot of cruise, source type and land use on a
1492 linear model (DistLM) of normalised predictor variables

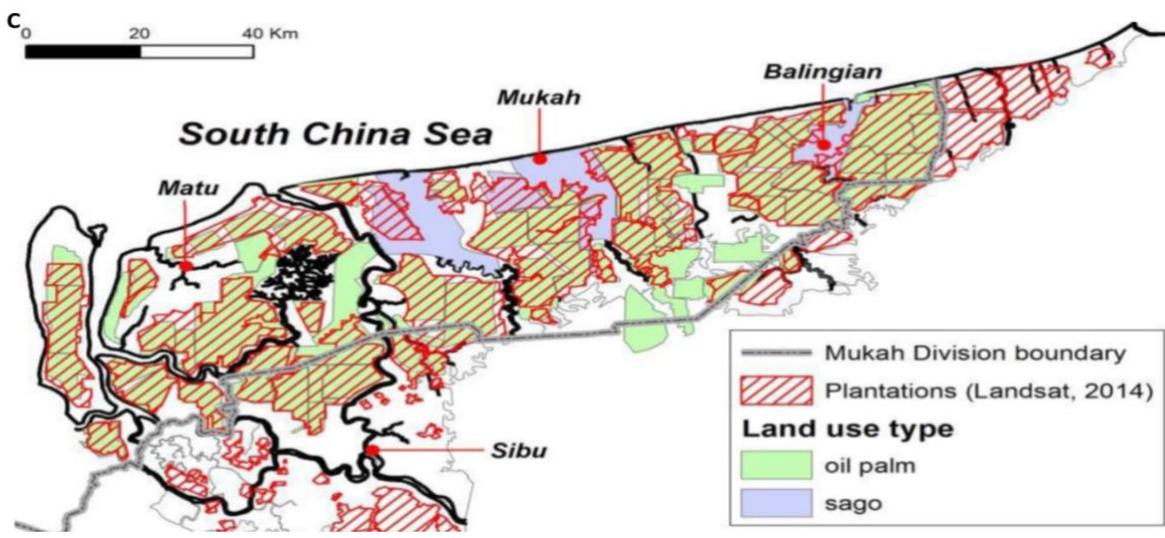
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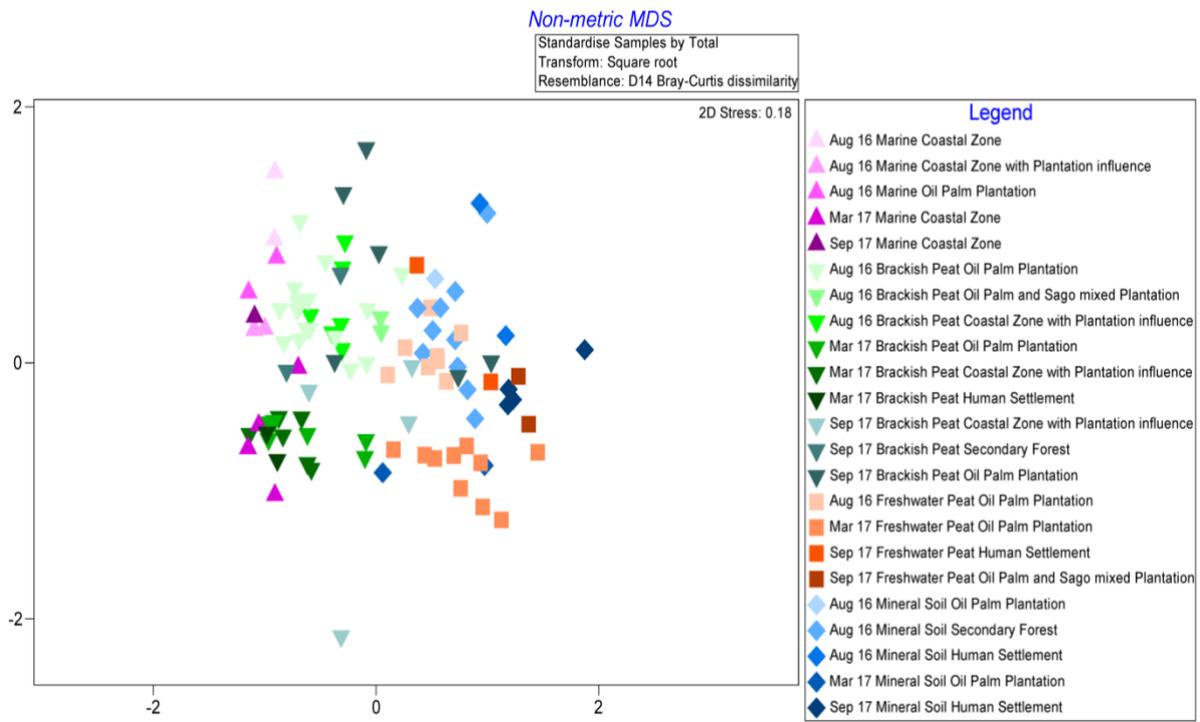


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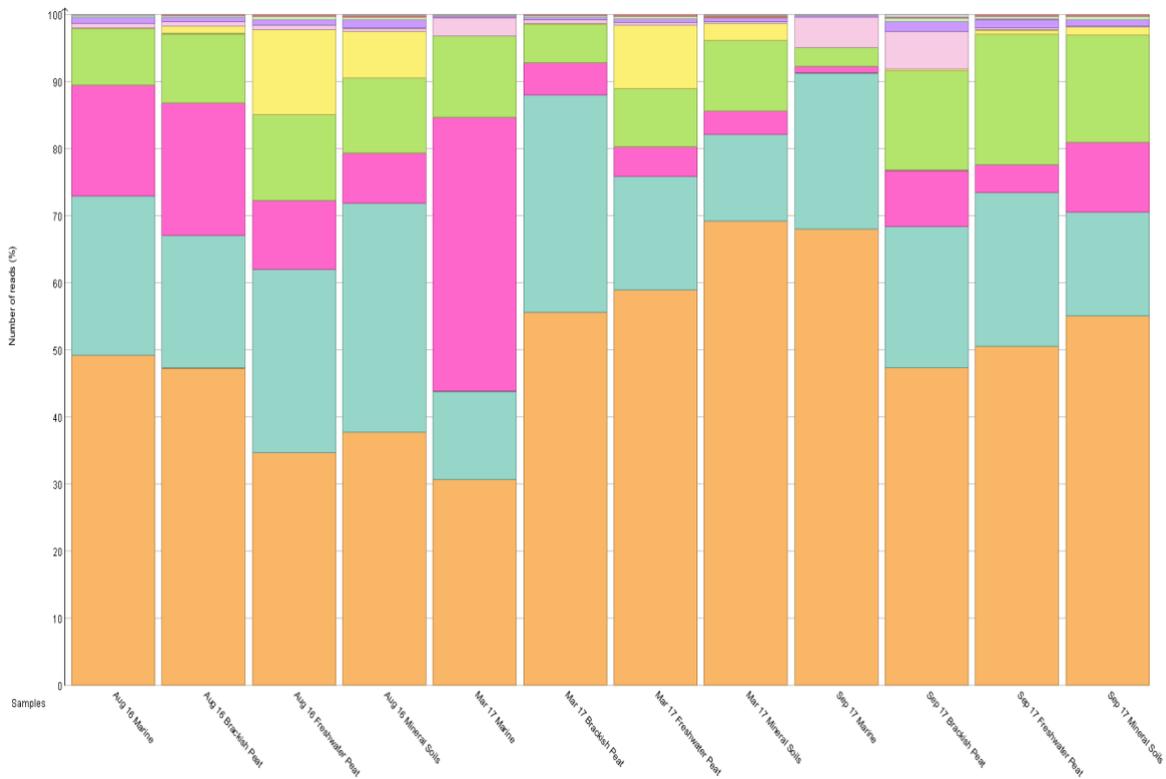
1498 **Fig. 1**



1500

1501 **Fig. 2**

1502

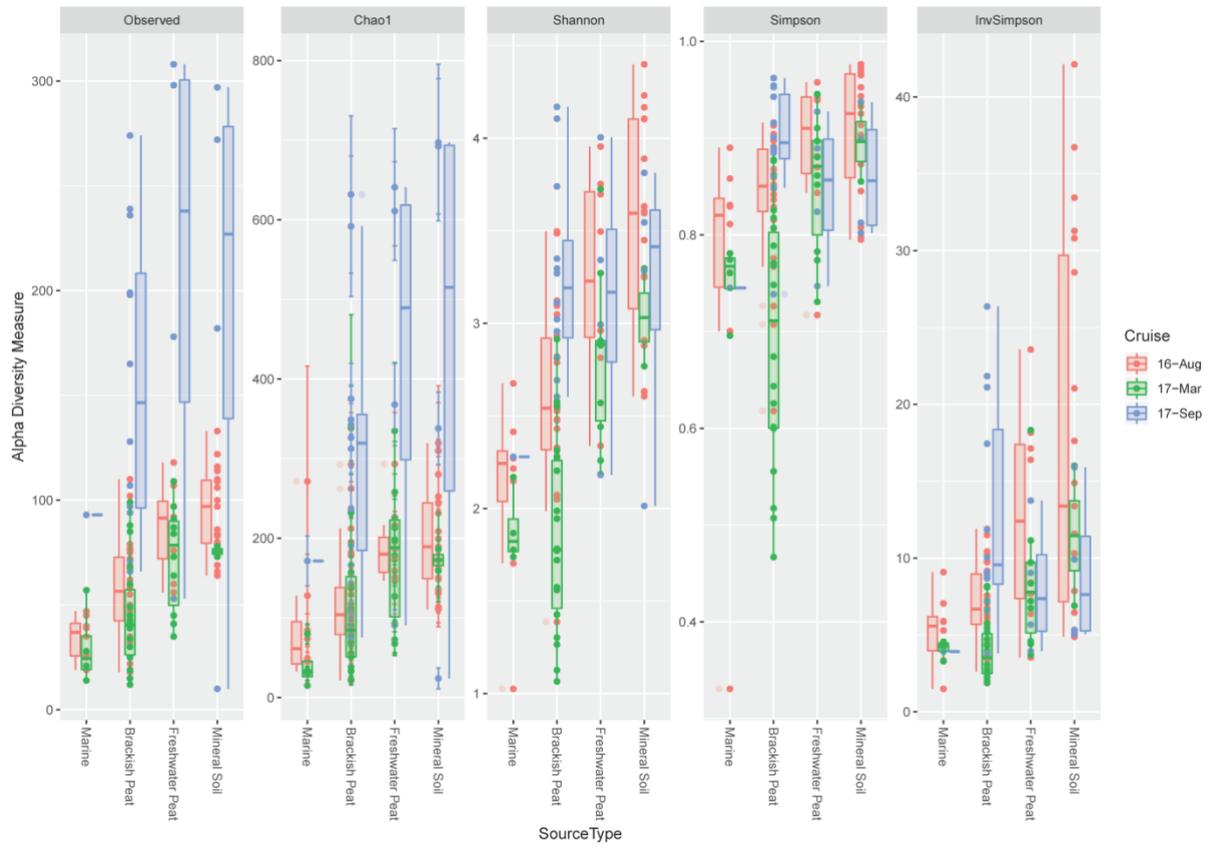


1503

1504 **Legend (Taxa):**
 Proteobacteria Firmicutes Bacteroidetes Actinobacteria <phylum> Deinococcus-Thermus Cyanobacteria Planctomycetes Chloroflexi Chlamydiae Verrucomicrobia

1504

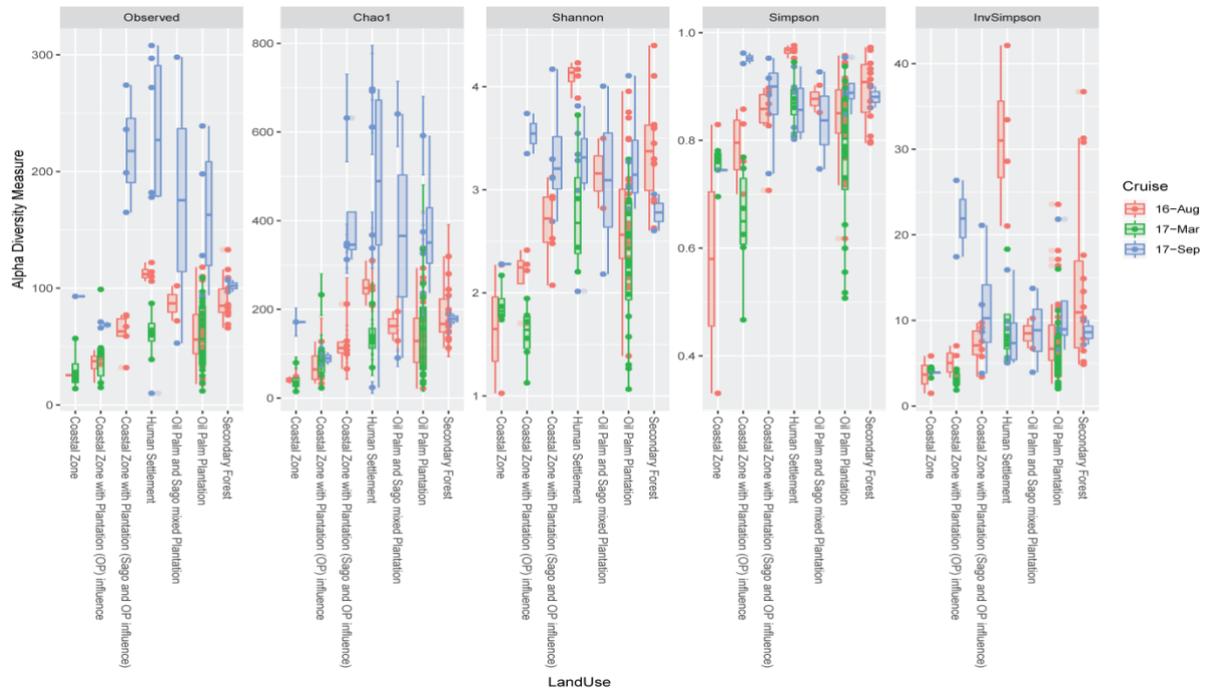
1505 **Fig. 3**



1506

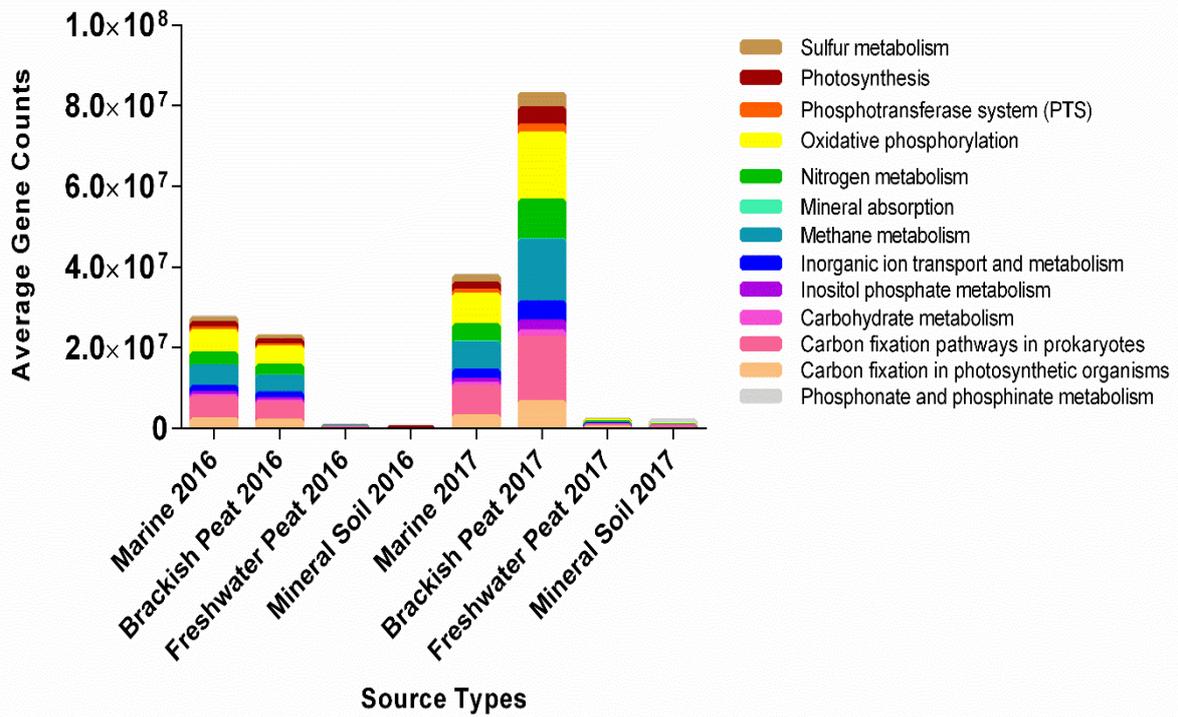
1507 **Fig. 4**

1508

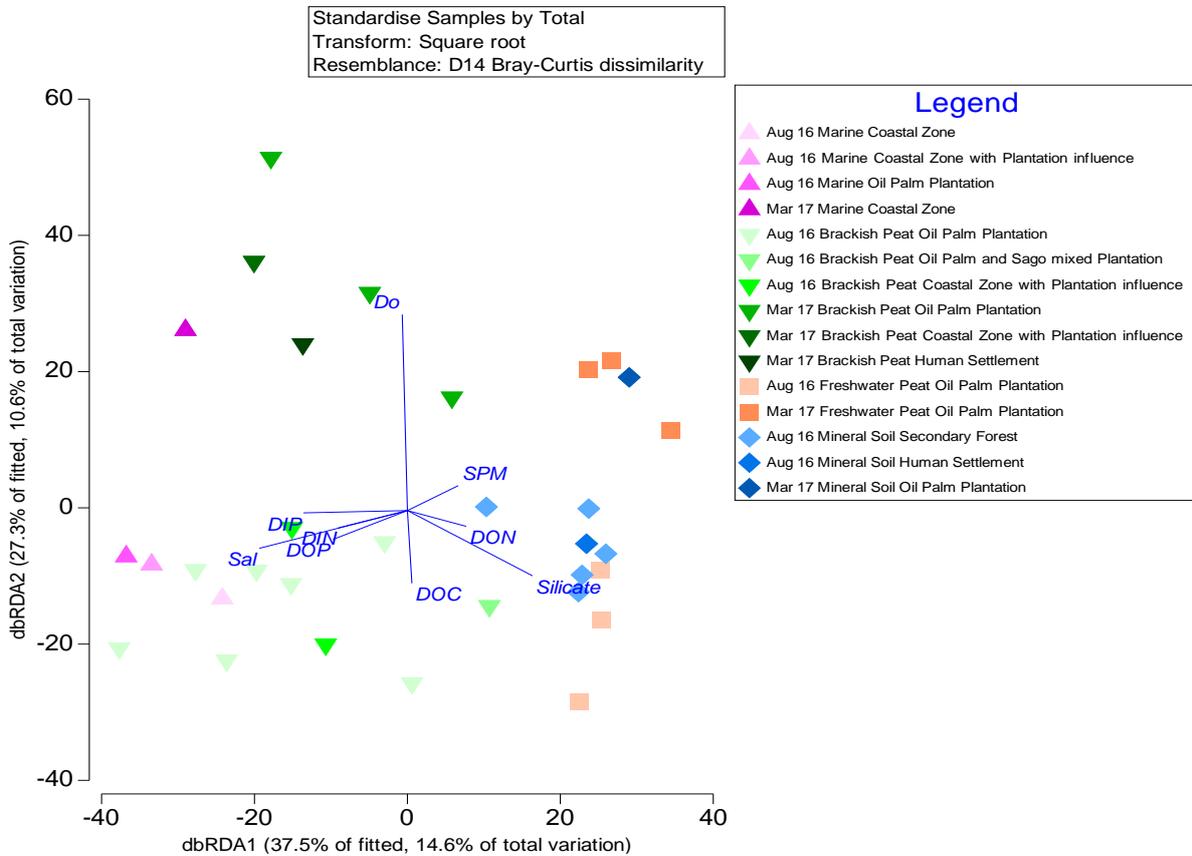


1509

1510 **Fig. 5**



1511 **Fig. 6**



1512

1513 **Fig. 7**