

- **Reply to comments by Reviewer #1** -

We thank the referee for the constructive comments. We considered all the suggestions and improved the manuscript accordingly. Answers are in red color.

**General Comments**

The paper by Legrand et al. describes an elegant mesocosm experiment testing the effects of ocean warming and acidification upon the community-scale responses of maerl bed assemblages. This work is timely, and provides an interesting insight into how the communities associated with coralline algae will respond to the impacts of anthropogenic climate change. Given the importance of coralline algae as a habitat architect, and the role of these communities in carbon fixation, I believe this work will make an important contribution to our understanding of coastal sea biogeochemistry. The statistical analysis, however, leaves a lot to be desired and as such, I cannot confidently review the authors' interpretation of their results or discussion. I am baffled as to why the authors have chosen to use permutational multivariate analysis of variance (perMANOVA) of similarity matrices (Euclidean distance) as a statistical test to test for differences in univariate response variable (e.g. respiration). Firstly, the authors make the erroneous assumption that perMANOVA does not make any assumptions about normality and homoscedacity of the data. However, Anderson (2001) point out in their papers describing these methods that the method makes the assumption of multivariate normality as measured by a homogenous dispersion of the similarity matrix data. Secondly, I simply cannot understand why perMANOVA was selected as a statistical test. There are more appropriate univariate tests such as analysis of variance (ANOVA) [with appropriate transformations applied], or if appropriate the use of Generalised Linear Models or Generalised Least Squares techniques which would allow the author to account for non-Gaussian data distributions (GLM) or heterogeneous variances between the treatments (GLS) (see Zuur et al., 2009). This represents a major issue with the handling of the experimental data, and so I cannot recommend the paper be accepted for publication in its current form. I strongly encourage the authors to revise the paper and resubmit. I think this has the potential to be an excellent paper and I will happily review a suitably revised manuscript.

Answer: As suggested, the statistical design has been modified in the m/s and is described in the section “2.5. Data analysis” (P. 9 Lines 191-197):

“Comparisons in species and assemblage physiological rates between the winter and summer seasons was performed using t-tests, after checking the normality and homogeneity of variances. The influence of temperature and pCO<sub>2</sub> was tested on metabolic rates of grazers (*P. miliaris*, *G. magus* and *J. exasperatus*), living and dead maerl, epiphytic biomass and assemblages. Normality of the data and variance homogeneity were checked for all variables. When assumptions were respected, two-way ANOVA were performed, using temperature and pCO<sub>2</sub> as fixed orthogonal factors. When assumptions were not respected, two-way non-parametric Scheirer-Ray-Hare tests were run. Statistical analyses were conducted separately for winter and summer experiments in order to keep a balanced design.”

## Specific Comments

### *Abstract*

Pg. 1 L 11: “However, little information is available on the response of marine communities...” I do not believe this is true. There has been considerable work of community scale responses to OA – see Ulf Riebesell’s work on planktonic communities and benthic-pelagic coupling as an example.

A: We have specified “benthic communities” in the abstract. (L. 11)

### *Introduction*

Pg. 2 L 34-35: Please specify examples of how species interactions are modified by climate change.

A: “Species interactions are a key element in ecosystem functioning and are likely to attenuate or amplify the direct effects of climate change on individual species (O’Connor et al., 2011; Hansson et al., 2012; Kroeker et al., 2012).” (L. 35-36)

Pg. 2 L 37: There are actually quite a number of studies examining the effects of climate change on marine communities. I recommend the authors carry out a thorough literature search.

A: We have reworded the sentence to reflect the growing interest of researches on benthic communities: (L. 36-39) “Most research on benthic ecosystems has focused on the impact of ocean acidification and

warming on the response of single species (Yang et al., 2016) and despite a growing interest, studies examining the effects of climate change at the community scale are scarce in the literature (Hale et al., 2011; Alsterberg et al., 2013).”

Pg. 3 L 61-63: “Because the responses of species...” This sentence seems rather poorly structured consider revising to clarify.

A: The sentence has been clarified: (L. 62-64) “Because the response of species and communities to climate change is likely to vary depending on seasonal changes in environmental factors, such as light intensity, photoperiod and temperature (Godbold and Solan, 2013; Martin et al., 2013; Baggini et al., 2014), it was tested in both winter and summer conditions.”

### *Materials and Methods*

Pg. 4 L 90 – 97: This should be a single paragraph.

A: Done

Pg. 5 L 100-109: This information would be better displayed as a table.

A: A new table (Table 1) shows the different pCO<sub>2</sub> and temperature conditions used for winter and summer experiments.

Table 1. Summary of the four experimental treatments. Two pCO<sub>2</sub> (ambient and high pCO<sub>2</sub>) and temperature (ambient and high temperature) conditions were tested. High pCO<sub>2</sub> (H-pCO<sub>2</sub>) corresponded to a pH decrease of -0.33 units compared to ambient conditions (A-pCO<sub>2</sub>). High temperature (T + 3°C) corresponded to a temperature increase of 3°C compared to ambient conditions (T).

	pCO <sub>2</sub>	Temperature	
1 (Control)	Ambient (A-pCO <sub>2</sub> )	Ambient (T)	A-pCO <sub>2</sub> ; T
2	High (H-pCO <sub>2</sub> )	Ambient (T)	H-pCO <sub>2</sub> ; T
3	Ambient (A-pCO <sub>2</sub> )	High (T+3°C)	A-pCO <sub>2</sub> ; T + 3°C
4	High (H-pCO <sub>2</sub> )	High (T+3°C)	H-pCO <sub>2</sub> ; T + 3°C

Pg. 9 L 190 – 201: Please revise around appropriate statistical tests.

A: The statistical design has been changed as discussed above.

### *Discussion*

Pg. 11 L 251-253: “Results show... underlying maerl.” This sentence is not clear, please specify the community responses to climate change more clearly.

A: (L. 254-256) “Results show that predicted changes may alter interactions among calcifying and fleshy macroalgae via overgrowth of epiphytic algae and an increase in competition for light and nutrients with underlying maerl.”

Pg. 16 L 358-359: The final line of the paper is vague, what specific pieces of further work would be useful?

A: The sentence has been reworded: (L. 378-381) “In order to better understand the consequences of climate change on ecosystem functioning, further work should focus on the response of marine communities and consider more specifically shifts in species interactions, including changes in trophic interactions between algae and grazers.”

### *Figures*

In the figures it would helpful to see which treatment effects are statistically significant, can you please find a way to highlight these effects in the graphs.

A: Following the suggestion of Referees #1 and #2, statistically significant results have been added on graphs: (L. 197-198) “When 2-way AVNOVAs showed significant results, post hoc tests (Tukey honest significant difference, HSD) were performed to compare the four treatments.” Results have been added on corresponding graphs. The direction of changes have also been added in tables (Tables 4, 5, 7 and 8) and interaction plots (in supplementary material) when a significant interaction between pCO<sub>2</sub> and temperature was detected.

- Reply to comments by Reviewer #2 -

## General Comments

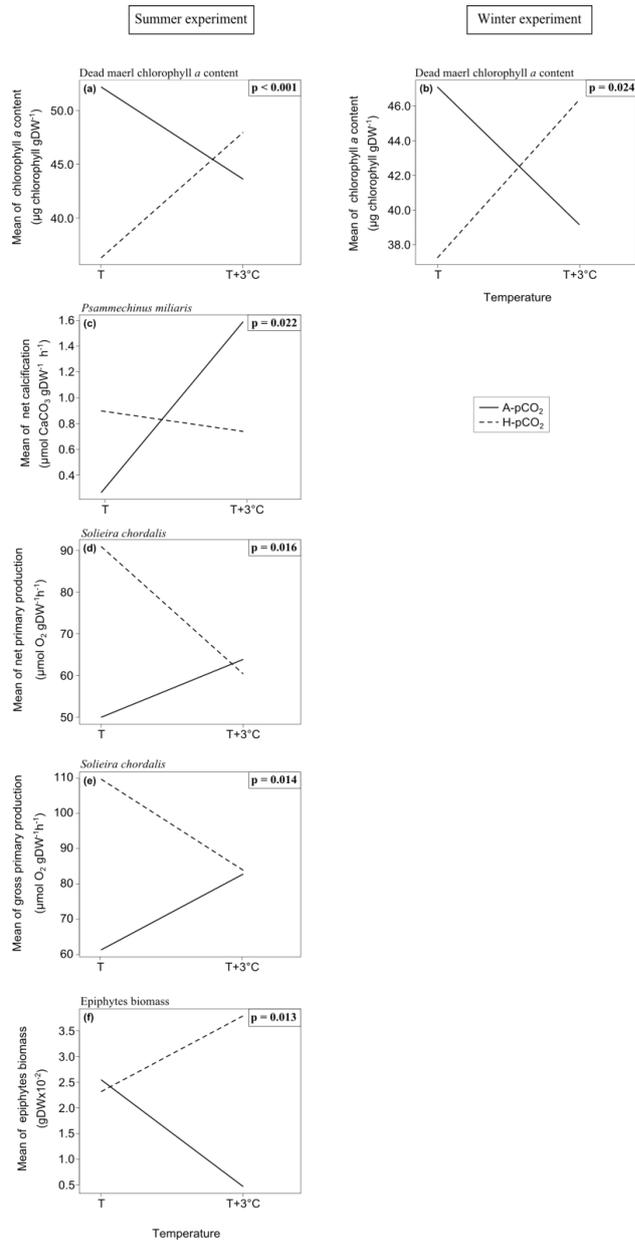
We thank the referee for the thoughtful and constructive comments that helped to improve the manuscript. We considered all the suggestions and improved the manuscript accordingly. Answers to referee's comments are in red color.

The manuscript "Species interactions can shift the response of a maerl bed community to ocean acidification and warming" describes a novel experiment with an interesting approach on community interactions under predicted global climate change that are generally lacking in the literature. Considering ocean acidification and warming are occurring simultaneously and interdependently, experiments that investigate the effects of both factors on marine organisms are important for understanding future changes in physiology and ecology. The authors were able to do this in their study, and not only did they investigate effects of ocean warming and acidification on the physiology of single organisms, but also of communities. Through their experimental design, they are able to describe changes in species interactions under future climate change conditions, which is currently rare in the literature. The experimental design is good and the manuscript is well written thorough.

My main criticism is that the results could be described more clearly and thoroughly. The interactions between the independent variables should be described more clearly. Interaction plots could help with the interpretation of the statistical analysis of the effect of season, temperature and CO<sub>2</sub> on the independent variables. The authors tested the effect of season on the dependent variables, but they often fail to describe this effect in the results section, and focus only on the CO<sub>2</sub> and temperature effect. They also fail to mention in some cases that temperature ameliorates the negative effect of pCO<sub>2</sub> on some variables, which is important considering both warming and acidification are occurring interdependently. I have made specific comments below.

A: The statistics have been changed according to the comment of Reviewer #1. The seasonal effect has been analyzed separately using t-tests. The effect of increased pCO<sub>2</sub> and temperature on the metabolism

of species and assemblage was examined in the winter and the summer using 2-way ANOVA. When an interactive effect of pCO<sub>2</sub> and temperature was evidenced, interaction plots were performed and provided in supplementary material to this paper. As suggested by the reviewer, the seasonal effect on metabolic parameters has now been discussed in the discussion section of the revised manuscript (L. 263-267). We also considered more closely the importance of season in the response of organisms and assemblages to acidification and warming.



Supplementary material. Interaction plots for the effects of temperature and pCO<sub>2</sub> on dead maerl chlorophyll *a* content in (a) the summer and (f) winter seasons, (b) *P. miliaris* net calcification in the summer, *S. chordalis* (c) net and (f) gross primary production in the summer, and (e) epiphytes biomass in the summer. Plots were done only when an interactive effect of temperature and pCO<sub>2</sub> was detected using 2-way ANOVA (p-value in bold).

## Specific comments

### *Materials and methods*

Line 141-142 "Before incubation, epiphytic algae that spontaneously grew on *L. corallioides* during the experiments were carefully removed and incubated separately." I assume this was done after the assemblage measurements were made? The authors could clarify this here.

A: Assemblage incubation was performed first. After this, epiphytic algae were removed from maerl. The sentence has been modified: (L. 137-138) "After assemblage incubations, epiphytic algae that spontaneously grew on *L. corallioides* during the experiments were carefully removed and incubated separately."

Line 156 What buffer solutions were used to calibrate the reactive spots?

A: (L. 152-154) "The 0% buffer solution was prepared by dissolving 1 g of sodium sulfite ( $\text{Na}_2\text{SO}_3$ ) in 100 mL of seawater. The 100% buffer solution was prepared by bubbling air into 100 mL of seawater using an air-pump for 20 min to obtain air-saturated seawater." This information has been added in the revised m/s.

### *Results*

In the results headings, the authors mention acidification and warming, but ignore the factor season

A: The factor season has now been taken into account for grazers (L. 203-205), living maerl (L. 214), dead maerl (L. 225-226), epiphyte biomass (L. 240-241) and assemblages (L. 245-246).

Lines 205-208 I think the results can be described more thoroughly here. There actually was not a negative effect of CO<sub>2</sub> and temperature on GI compared to the control. Temperature increased calcification rates in the summer. CO<sub>2</sub> alone did not seem to have an effect in either season. The combination of high CO<sub>2</sub> and temperature in summer negated the positive effect of temperature.

A: The main effect of season on *P. miliaris* has been mentioned (L. 203-204). The sentence on the effect of temperature and pCO<sub>2</sub> on *P. miliaris* G<sub>1</sub> has been reworded: (L. 206-207) “*P. miliaris* G<sub>1</sub> was significantly affected by the interaction between temperature and pCO<sub>2</sub> in the summer (Fig. 1b, supplementary material b), which negated the positive effect of increased temperature and pCO<sub>2</sub> alone.”. We also used interaction plots (in supplementary material) to illustrate pCO<sub>2</sub> and temperature combined effect.

The authors should mention there was a main effect of season on *P. miliaris* E. Excretion was highest under control conditions in the summer. High temperature, CO<sub>2</sub> and the combination of both decreased excretion rates in the summer.

A: We have added the effect of season on *P. miliaris* E. We have reworded this section: (L. 208-209) “*P. miliaris* E was higher under control conditions in the summer and increased temperature significantly reduced *P. miliaris* E (Table 4; Fig. 1c).”

Lines 209-210 It is confusing to say R was positively or negatively affected - please rather describe if it increased or decreased. Also, although there was no temperature or CO<sub>2</sub> effect on R, G<sub>1</sub>, or E, there was a strong effect of season.

A: We agree with the Reviewer and changed the sentence accordingly: (L. 210-211) “In *J. exasperatus*, R increased under elevated temperature but in winter conditions only (Table 4; Fig. 1g).” The effect of season has now been added.

Line 213 Please add the effect of season, e.g. "Living maerl GPP did not differ among temperature and pCO<sub>2</sub> conditions, but there was a strong effect of season, with higher rates in the summer than in the winter."

A: The effect of season has now been added: (L. 214) “The metabolism of living *L. corallioides* was higher in the summer than in the winter, except for NPP (Table 3).”

Line 215 Add the effect of season on chlorophyll a

A: We have added: (L. 217-218) “No effect of season was observed on chlorophyll *a* content (Tables 3; 6).”

Line 216 "Temperature had a positive effect on the GI of living maerl. Conversely, GI was significantly reduced under high pCO<sub>2</sub>..." The authors fail to mention that in the combined treatment, temperature alleviated the negative effect of pCO<sub>2</sub>. This is very important to the story.

A: The sentence has been revised due to the change in statistical design. (L. 218-219) "The G<sub>1</sub> of living maerl was not significantly influenced by increased temperature and pCO<sub>2</sub>, regardless of the season"

Line 220 "Net dissolution, because G<sub>d</sub> was negative, was recorded in the winter under high pCO<sub>2</sub> conditions" But dissolution was less in the combined temperature + CO<sub>2</sub> treatment in the winter, so temperature alleviated some of the negative effect of CO<sub>2</sub> in the winter, although net dissolution still occurred.

A: A sentence has been added: (L. 222-223) "This negative effect of increased pCO<sub>2</sub> was alleviated under elevated temperature."

Line 222 Again, mention the main effect of season

A: The effect of season has now been added (L. 225-226).

Line 223 I did not see an interaction between season and pCO<sub>2</sub> for GPP in Table 3 Line 225 Mention the effect of season on dead maerl

A: We apologize for this mistake, this has been withdrawn. The sentence has been modified to consider the effect of season (L. 225-226).

Line 233 "R was enhanced by the high temperature and pCO<sub>2</sub> conditions..." alone, and their combination resulted in the greatest R rates.

A: We have added this information: (L. 238-239) "R was enhanced by the high temperature and pCO<sub>2</sub> conditions and their combination resulted in a greater R"

Line 238 Add that temperature alone decreased epiphyte biomass in the summer.

A: We have added "Epiphyte biomass was not affected by increased temperature or pCO<sub>2</sub> in the winter (2-way ANOVA, p=0.95 and 0.67 respectively), while an interactive effect of temperature and pCO<sub>2</sub> was observed in the summer (p=0.013, supplementary material e)."

Line 240 "No temperature effect was observed..." But all response variables were higher in the summer than in the winter.

A: The effect of season has now been added (L. 246-247).

Line 248 "In the summer, the interactive effect of temperature and pCO<sub>2</sub> increase was more complex, with a (change to) increase in G<sub>d</sub> detected under high temperature conditions only."

A: The sentence has been changed: (L. 250-251) "In the winter, high pCO<sub>2</sub> increased net dissolutions rates, while in the summer G<sub>d</sub> increased under elevated temperature."

### *Discussion*

The authors state that "ocean acidification and warming will strongly destabilize communities through both direct effects on species physiology and changes in the interaction strengths between coralline algae, fleshy algae, and grazers." Based on the assemblage data, I do not think that the effect is so negative, at least in the summer. There is a strong difference in the effect of the combination of CO<sub>2</sub> and temperature in winter and summer. In summer, assemblages exposed to high temperature and pCO<sub>2</sub> combined actually had similar to or even slightly higher light calcification rates than the ambient treatment. In winter, there was a decrease in light calcification compared to the ambient treatment, but the positive effect of temperature and the negative effect of pCO<sub>2</sub> were weakened when the two were combined. I think it is important for the authors to point out that the combination of pCO<sub>2</sub> and temperature often subdued the effects of each single factor, because it illustrates the point that experiments investigating only the effect of pCO<sub>2</sub> or temperature may present more dramatic responses than when the two are combined, which represents a more realistic scenario.

A: The conclusion section has been reworded to consider this comment and those of Referee #1 and #3: "In conclusion, the community response to climate change does not appear to be only the result of individual species' metabolic responses, but also strongly depends on shifts in species interactions. In contrast with other studies, which evidenced larger impacts of the combination of increased pCO<sub>2</sub> and temperature than that of these factors alone (Reynaud et al., 2003; Anthony et al., 2008; Martin and Gattuso, 2009; Rodolfo-Metalpa et al., 2010), we showed here that the effects of pCO<sub>2</sub> and temperature on maerl bed communities were weakened when these factors were combined. Under the predicted business-as-usual conditions, epiphyte overgrowth may exacerbate the negative impact of climate change on underlying coralline algae. Here, we also demonstrated that climate change may affect grazer

physiology, with major consequences on their ability to regulate epiphyte biomass. Climate change may also affect other components that we did not assess in the present study, such as algal palatability and potential changes in grazer trophic behavior (Campbell et al., 2014; Duarte et al., 2015; Poore et al., 2013; Poore et al., 2016). Algal palatability to grazers may also be affected by predicted changes through shifts in the composition and the quantity of allelopathic compounds, as suggested by Del Monaco et al. (2017). In order to better understand the consequences of climate change on ecosystem functioning, further work should focus on the response of marine communities and consider more specifically shifts in species interactions, including changes in trophic interactions between algae and grazers.”

### *Technical Comments*

Line 129: insert "the" before CO2SYS

A: Done (L. 125)

It would be helpful to be able to identify statistically significant differences in the figures

A: (L. 197-198) “When 2-way ANOVAs showed significant results, post hoc tests (Tukey honest significant difference, HSD) were performed to compare the four treatments.” Results have been added on corresponding graphs. We have also added the direction of changes (2-way ANOVA) in tables 4, 5, 7 and 8 and interaction plots (in supplementary material) when a significant interaction between pCO<sub>2</sub> and temperature was detected.

- **Reply to comments by Reviewer #3** -

We thank the referee and appreciate the thoughtful and constructive comments. We have fully considered the referee's comments and improved the manuscript accordingly. Answers to referee's comments are in red color.

The study by Legrand et al. assessed the metabolic responses of a range of species associated to maerl beds (incl coralline algae, grazers and epiphytic fleshy algae), as well as the metabolic responses of the maerl assemblage to changes in seawater carbonate chemistry and temperature across two climatic seasons. The authors found complex interactions among experimental factors and seasons on the species and community metabolism. The coralline algae exhibited responses which were expected under CO<sub>2</sub> perturbation experiments, but importantly, the study documented significant changes to grazers' metabolism and enhanced epiphytic algal biomass under CO<sub>2</sub> enrichment. Although ecological interactions were not directly assessed, changes in the metabolic responses of the experimental species are assumed to influence species interactions. Based on these results the authors were able to propose that ocean acidification and warming will have considerable impacts on the functioning of maerl beds.

I read this manuscript with great interest and believe the authors have done a comprehensive and thorough study. Most studies in the field of impacts of climate change on marine systems focus on responses of one or two species, generally within the same taxonomic group, and it is refreshing to see that this study took a step forward and assessed the impacts at the community level during two climatic seasons. Individual responses focussed on a range of response variables, incl chlorophyll (for the algae), net production, respiration, net calcification (light and dark), and excretion for the grazers. In combination with the assemblage's responses, this allowed the authors to discuss some potential ecological implication such as shifts in species composition, competition, carbon storage, etc. The Methods are generally well described and provide enough detail so that other researchers can repeat the experiments. Methods are appropriate for ocean acidification research.

## Main comments

I have two main comments to the paper. First, seasonal effects on both the individual and assemblage responses were not fully explored or discussed in the m/s. One of the strengths of this m/s is that it was conducted in two different climatic seasons, but how the strength of the responses varied between seasons was not clear. I would suggest that the authors include a section where this comment can be fully addressed.

**Answer:** We agree with this comment. A new paragraph has been added in the discussion to explore the seasonal effects on species and community metabolism. In the result section, the influence of season has been detailed for each species and the community. Further information was also added throughout the m/s to understand how the strength of the response varied between the two seasons tested.

(L. 262-267): “Assemblage exhibited a strong seasonal pattern for all metabolic parameters, which is consistent with the higher metabolism in the summer for most of the species incubated at the specific scale. This higher metabolism in the summer has already been evidenced in urchins (Brockington and Peck, 2001), gastropods (Davies, 1966; Innes and Houlihan, 1985; Martin et al., 2006b) and living maerl (Potin et al., 1990; Martin et al., 2006a) and is strongly related to changes in numerous environmental and biological variables, such as light intensity and photoperiod, temperature and nutrient or food availability (Godbold and Solan, 2013; Thomsen et al., 2013).”

The statistical analyses seem to be well executed, however, I would argue that because there were significant interactions between treatments (OA, temp, and season), there is a need to conduct further statistical analyses within treatment combinations, as in several instances, the main factor was significant, but in fact it was only significant for one or the other season, or under a particular treatment combination. For example, in line 213, “R was significantly reduced by the high temperature condition in the winter, whereas an increase in R was observed in the summer.” This statement is fine, but is not actually supported by a statistical analysis as Table 3 only provides p values for the main effects. This issue is also evident 216-219. Underwood (1997; Experiments in Ecology: Their Logical Design and Interpretation Using Analysis of Variance, Cambridge University Press) provides information on this topic. These new analyses could be included as supplementary material.

A: According to the comments of Referee #1, we have modified statistical analyses. The seasonal effect has now been tested separately (using t-tests) in order to keep a balanced statistical design. The effect of temperature and pCO<sub>2</sub> was analyzed through 2-way ANOVA for each season separately. When an interactive effect of temperature and pCO<sub>2</sub> was observed, interaction plots were performed and provided in the supplementary material. (L. 197-198) “When 2-way ANOVAs showed significant results, post hoc tests (Tukey honest significant difference, HSD) were performed to compare the four treatments.” (according to suggestions of Referee #2). Results have been shown in corresponding graphs.

There are some statements that are not supported by the experiments. Although the authors demonstrated changes in algal and grazer metabolisms, species interactions among those organisms were not examined experimentally. E.g. Line 251. “Our study demonstrates that the response of maerl bed communities to increased temperature and pCO<sub>2</sub> conditions is a complex function of direct effects of climate variables on species physiology and shifts in species interactions”. Reword this statement.

A: We agree with this comment and, as suggested by the reviewer, this sentence has been reworded: (L. 253-254) “The response of communities to increased temperature and pCO<sub>2</sub> conditions is likely to be a complex function of direct effects of climate variables on species physiology and shifts in species interactions (Lord *et al.*, 2017).”

*Lord, J. P., Barry, J. P., and Graves, D.: Impact of climate change on direct and indirect species interactions, Marine Ecology Progress Series, 571, 1-11, 2017.*

## Minor comments

Unclear why chl<sub>a</sub> was measured on dead Lithothamnion. Provide a brief justification in section 3.3.

A: A sentence has been added in the “Material and Methods” section to justify chlorophyll *a* measurements: (L. 185-186) “In dead maerl, chlorophyll *a* content was measured in order to check for the presence of associated microflora and potential subsequent metabolism.”

Line 90: In general avoid single-sentence paragraphs.

A: The sentence was grouped with the following paragraph

Line 237: “.. having positive effect”. Was this effect significant?

A: More details have been provided due to the change in statistics recommended by referees: (L. 241-243) “Epiphyte biomass was not affected by increased temperature or pCO<sub>2</sub> in the winter (2-way ANOVA, p=0.95 and 0.67 respectively), while an interactive effect of temperature and pCO<sub>2</sub> was observed in the summer (p=0.013, supplementary material e).”

L260-280: This is a very long paragraph, try breaking it into two.

A: This paragraph has now been divided in 2 paragraphs in the revised manuscript.

L285-305: This is also a very long paragraph.

A: This paragraph has now been divided in 2 paragraphs in the revised manuscript.

L291: Ordonez et al. (Ordonez Alvarez et al. 2014 Effects of ocean acidification on population dynamics and community structure of crustose coralline algae. Biological Bulletin 226, 255-268.) also found a failure in recruitment of tropical CCA and importantly documented shifts in species composition.

A: The reference has now been added in the revised manuscript. (L. 304)

Line 303: “However, the present findings do not support this idea, because a decline in G<sub>1</sub> was observed under high pCO<sub>2</sub> despite high”. Short et al (2014) paper dealt with minute algal turfs which may have altered the thickness of the diffusive boundary layer on the coralline algae. The macroalgae investigated in the present study were much bigger and may interact in many different ways. It is perhaps very difficult to generalise the impacts of epiphytic algae on coralline algae given the diversity of algae in marine systems. Perhaps a line or two addressing this would be useful.

A: We agree with this comment. We have completed with the following sentences: (L. 314-325) “Conversely, other studies evidenced that the overgrowth of epiphytic fleshy algae may shade underlying coralline algae and reduce coralline net calcification rates (Garrabou and Ballesteros, 2000; Martin and Gattuso, 2009). The present findings support this idea, because a decline in assemblage G<sub>1</sub> was observed under high pCO<sub>2</sub> and high epiphyte biomass. [...] Thus, overgrown maerl would be negatively affected by the direct effect of ocean acidification on calcification rates and indirect effects due to shifts in competition dynamics with fleshy epiphytic algae (Kuffner et al., 2008). However, the response of epiphytic algae is likely to be specie-specific and it appears difficult to generalize the impacts of epiphytic algae on coralline algae.”

Pages 14-15: Grazing responses may also be altered by changes in seaweed allelopathic compounds, brought about by changes in composition, quantity, or in the magnitude/potency of the allelopathic interactions. A recent study showed that the potency of allelopathic interactions towards a tropical coral was intensified under ocean acidification conditions (Del Monaco et al. 2017 Effects of ocean acidification on the potency of macroalgal allelopathy to a common coral. Scientific Reports 7, 41053). May be worth adding this potential mechanism as drivers of changes in species interactions in response to acidification and warming.

A: As suggested by the reviewer, we have now added this information: (L. 377-378) “Algal palatability to grazers may also be affected by predicted changes through shifts in the composition and the quantity of allelopathic compounds, as suggested by Del Monaco et al. (2017).”

- **List of all relevant changes made in the manuscript –**

- The main effect of season has been mentioned and discussed for all species and the community.
- Statistical analyses have been modified. The seasonal effect has now been tested separately using t-tests. The effect of temperature and pCO<sub>2</sub> was analyzed through 2-way ANOVA for each season separately.
- Statistically significant differences have been identified in the figures (Tukey HSD test).

# Species interactions can shift the response of a maerl bed community to ocean acidification and warming

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## Abstract

10 Predicted ocean acidification and warming are likely to have major implications for marine organisms, especially marine calcifiers. However, little information is available on the response of marine benthic communities as a whole to predicted changes. Here, we experimentally examined the combined effects of temperature and partial pressure of carbon dioxide (pCO<sub>2</sub>) increases on the response of maerl bed assemblages, composed of living and dead thalli of the free-living coralline alga *Lithothamnion corallioides*, epiphytic fleshy algae, and grazer species. Two three-month experiments were performed

15 in the winter and summer seasons in mesocosms with four different combinations of pCO<sub>2</sub> (ambient and high pCO<sub>2</sub>) and temperature (ambient and + 3°C). The response of maerl assemblages was assessed using metabolic measurements at the species and assemblage scales. This study suggests that seasonal variability represent an important driver influencing the magnitude and the direction of species and community response to climate change. Gross primary production and respiration of assemblages were enhanced by high pCO<sub>2</sub> conditions in the summer. This positive effect was attributed to the increase in

20 epiphyte biomass, which benefited from higher CO<sub>2</sub> concentrations for growth and primary production. Conversely, high pCO<sub>2</sub> drastically decreased the calcification rates in assemblages. This response can be attributed to the decline in calcification rates of living *L. corallioides* due to acidification as well as increased dissolution of dead *L. corallioides*. Future changes in pCO<sub>2</sub> and temperature are likely to promote the development of non-calcifying algae to the detriment of the engineer species *L. corallioides*. The development of fleshy algae may be modulated by the ability of grazers to regulate

25 epiphyte growth. However, our results suggest that predicted changes will negatively affect the metabolism of grazers and potentially their ability to control epiphyte abundance. Here, We evidenced here that the effects of pCO<sub>2</sub> and temperature on

~~maerl bed communities were weakened when these factors were combined~~~~we demonstrate that the response of marine communities to climate change will depend on the direct effects on species physiology and the indirect effects due to shifts in species interactions.~~ This ~~double, interdependent response~~ underlines the importance of examining multi-factorial approaches and community-level processes, which integrate species interactions, to better understand the impact of global change on marine ecosystems.

## 1. Introduction

A growing body of literature predicts that ocean acidification and warming will be the main anthropogenic drivers affecting marine species by the end of the century (Kroeker et al., 2013). Due to the increase in atmospheric CO<sub>2</sub>, seawater surface temperatures have been predicted to increase by 0.71-2.73°C and pH to decline by 0.07-0.33 units in the surface ocean by the end of the 21<sup>st</sup> century (Bopp et al., 2013).

Species interactions are a key element in ecosystem functioning and are likely to attenuate or amplify the direct effects of climate change on individual species ~~play an important role in species responses to climate change~~ (O'Connor et al., 2011; Hansson et al., 2012; Kroeker et al., 2012). ~~To date,~~ Most research on benthic ecosystems has focused on the impact of ocean acidification and warming on the response of single species (Yang et al., 2016) and despite a growing interest, studies examining the effects of climate change ~~on marine communities~~ at the community scale are scarce in the literature (Hale et al., 2011; Alsterberg et al., 2013). Understanding the mechanisms and interactions that occur among marine communities that face the predicted changes is necessary for a better overview of marine ecosystem response. Climate change is likely to strongly alter interactions between macroalgae (e.g. calcifying and non-calcifying macroalgae; Olabarria et al., 2013; Short et al., 2014; Short et al., 2015), interactions between grazers and macroalgae (Poore et al., 2016; Sampaio et al., 2017) as well as prey-predator dynamics (Asnaghi et al., 2013; Jellison et al., 2016), inducing drastic consequences on the structure and functioning of marine ecosystems (Widdicombe and Spicer, 2008; Hale et al., 2011).

Maerl beds feature high structural and functional diversity arising primarily from the numerous species interactions that occur in this environment — in particular, interactions between fleshy and calcareous macroalgae and grazers and macroalgae (Hily et al., 1992; Guillou et al., 2002; Grall et al., 2006). The accumulation of living and dead thalli of free-

living coralline algae (Corallinaceae, Rhodophyta) creates a complex three-dimensional structure that provides habitat for many faunal and floral species (Foster et al., 2007; Amado-Filho et al., 2010; Peña et al., 2014), some of which have high commercial value (Grall and Hall-Spencer, 2003). In some locations, dead maerl can reach high proportions compared with living maerl (Hily et al., 1992), thereby contributing substantially to the local carbonate dynamics (Martin et al., 2007).

55 The main species inhabiting maerl beds may respond differently to ocean acidification and warming. Coralline algae are known to be among the most vulnerable species facing ocean acidification (McCoy and Kamenos, 2015; Martin and Hall-Spencer, 2016), due to their highly soluble Mg-calcite skeleton (Morse et al., 2006). The deleterious consequences of ocean acidification have also been demonstrated for other calcareous marine taxa, such as mollusks (Gazeau et al., 2013; Parker et al., 2013) and echinoderms (Dupont et al., 2010), with reductions in survival, growth, development, and abundance (Kroeker et al., 2013). Conversely, some species can benefit from the increase in CO<sub>2</sub> concentration and temperature. Positive responses, such as increases in primary production and growth, have been found mostly among non-calcifying organisms, such as fleshy algae and seagrasses (Koch et al., 2013; Pajusalu et al., 2013).

Here, we experimentally investigated the impact of ocean acidification and warming on the metabolism and the interactions of the main maerl-forming species in Brittany *Lithothamnion coralloides* and the epiphytic fleshy macroalgae and main grazer (gastropods and sea urchins) associated with it. Because the response of species and communities to climate change is ~~also~~ likely to vary depending on seasonal changes in environmental factors, such as light intensity, photoperiod and temperature (Godbold and Solan, 2013; Martin et al., 2013; Baggini et al., 2014), it was tested in both winter and summer conditions, experiments were performed in both winter and summer conditions. The response of marine communities to climate change is likely to be influenced by the direct effects of environmental stressors on individual organisms, and by the indirect effects induced by shifts in interspecific interactions (Harley et al., 2012; Auster et al., 2013). In the present study, we therefore performed metabolic measurements at the species and at the community scale. At the species scale, studying species physiology is useful for understanding how organisms cope with changing climatic conditions and for analyzing the community metabolic response. Community-scale measurements provide information on the potential shifts in species interactions induced by climate change. In particular, we tested the hypothesis that climate change will increase epiphytic fleshy algal growth, exacerbating the deleterious consequences of predicted changes on *L. coralloides* metabolism. We also

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investigated whether the predicted changes can modify interactions between grazers and macroalgae, and their ability to regulate epiphytic biomass.

## 2. Materials and methods

### 2.1. Species collection and assemblages

80 Organisms were collected from a maerl bed in the Bay of Brest, France (48°18'N 4°23'W) using a naturalist's dredge (width: 1 m, height: 0.2 m, net: 1.5 m long) deployed from the research vessel *Albert Lucas*. In the Bay of Brest, maerl beds are located at depths of between 0.7 and 6.8 m, according to the tide (Dutertre et al., 2015). We deliberately selected thalli of the maerl species *L. corallioides* Crouan and Crouan, 1867 that were devoid of any apparent epiphytes; nonetheless, they were not cleaned so as to retain any epiphyte spores that may have been present on their surface. Medium-sized individuals of the  
85 three main species of grazers living in maerl beds were also sampled: two gastropod species (sea snails) *Gibbula magus* Linnaeus, 1758 and *Jujubinus exasperatus* Pennant, 1777 and an urchin species *Psammechinus miliaris* Müller, 1771 (Grall et al., 2006). Samples were collected on 24 January 2015 (winter conditions) and 15 September 2015 (summer conditions). In each season, 1 kg of living thalli of *L. corallioides*, 500 g of dead thalli of *L. corallioides*, 40 individuals of *G. magus* (shell length range 17-29 mm; Table S1), 40 individuals of *P. miliaris* (test diameter range 11-23 mm), and 80 individuals of  
90 *J. exasperatus* (shell height range 5-11 mm) were randomly selected and transported in seawater tanks to the Roscoff Marine Station. To mitigate the stress experienced by the species during sampling and transport, they were kept in open-flow aquaria at ambient pH and *in situ* temperature conditions at the time of collection for at least one week before starting the experiments. No mortality was recorded during this period.

### 2.2. Experimental design

95 Two three-month long experiments were conducted for both winter (March to June 2015) and summer (September to December 2015) conditions. For each season, 20 artificial assemblages were created and randomly assigned to 20 15 L aquaria. Each assemblage was composed of 45 g of living *L. corallioides* thalli, 20 g dead *L. corallioides* thalli, two *G. magus* individuals, two *P. miliaris* individuals and four *J. exasperatus* individuals, according to the proportions observed on maerl beds.

100 Algae and grazers were acclimated to laboratory conditions for 7 days. Then, the pH was gradually decreased by 0.05 units per day over 7 days and temperature increased by 0.5°C per day. The pH was controlled by modifying pCO<sub>2</sub> through CO<sub>2</sub> bubbling. At each season, two pCO<sub>2</sub> conditions were tested, each with two temperature conditions to examine the interaction between pCO<sub>2</sub> and temperature. ~~There were therefore four conditions~~ These four conditions are presented in table 1.

~~1) ambient pCO<sub>2</sub> and ambient temperature (control, A-pCO<sub>2</sub>; T)~~

105 ~~2) high pCO<sub>2</sub> and ambient temperature (H-pCO<sub>2</sub>; T)~~

~~3) ambient pCO<sub>2</sub> and high temperature (A-pCO<sub>2</sub>; T + 3°C)~~

~~4) high pCO<sub>2</sub> and high temperature (H-pCO<sub>2</sub>; T + 3°C).~~

Ambient pCO<sub>2</sub> conditions (A-pCO<sub>2</sub>) were determined according to *in situ* winter (7.98) and summer (8.06) mean pH<sub>T</sub> (pH on the total scale) monitored above maerl beds in the Bay of Brest (data from Martin, unpublished data). High pCO<sub>2</sub> (H-pCO<sub>2</sub>) corresponded to the “business-as-usual” scenario predicted for the end of the century, with a pH decrease of -0.33 units (RCP8.5; Bopp et al., 2013). Ambient temperature (T) corresponded to *in situ* winter (10.0°C) and summer (17.1°C) conditions in the Bay of Brest recorded by SOMLIT (from 2003 to 2014), and high temperature (T + 3°C) was determined according to the business-as-usual scenario predicted for 2100 (Bopp et al., 2013).

The pH and the temperature were controlled in four 100 L tanks, continuously supplied with filtered (5 µm) natural seawater, with a high water flow rate of 150 L h<sup>-1</sup> per tank. They were maintained by an off-line feedback system (IKS Aquastar, 115 Karlsbad, Germany) that activated or stopped heaters and solenoid valves, controlling temperature and CO<sub>2</sub> (Air Liquide, France) bubbling in the tanks, respectively. Each 100 L tank provided seawater to five 15 L aquaria for each of the four conditions using pumps. The water flow rate was 15 L h<sup>-1</sup> in each aquarium. Temperature was maintained constant in aquaria with water baths. Seawater pH (pH<sub>T</sub>, expressed on the total hydrogen ion concentration scale, Dickson et al., 2007) and 120 temperature were monitored every two days in the 20 aquaria, at different times of the day. Seawater pH<sub>T</sub> and temperature measurements were carried out using a pH probe associated with a temperature sensor (PHC101, Hach Lange, IntelliCAL). The pH probe was calibrated using Tris/HCl and 2-aminopyridine/HCl buffers (Dickson et al., 2007). The pH values of the off-line feedback system were adjusted from measurements of pH<sub>T</sub> carried out every two days in each aquarium. Total alkalinity (A<sub>T</sub>) was also monitored during the experiment in each aquarium at different times of the day (n = 28). For A<sub>T</sub>

125 analyses, seawater samples (60 mL) were filtered through 0.7  $\mu\text{m}$  Whatman GF/F filters and immediately poisoned with a mercuric chloride solution to prevent further biological activity (Dickson et al., 2007).  $A_T$  was determined using open-cell titration on an automatic titrator (Titroline alpha, Schott SI Analytics, Mainz, Germany) according to the method developed by Dickson et al. (2007).  $A_T$  was calculated using a Gran function applied to pH values ranging from 3.5 to 3.0 (Dickson et al., 2007) and corrected using standard reference material provided by the Andrew G. Dickson laboratory (CRM Batch 111, 130 accuracy of  $\pm 6 \mu\text{mol kg}^{-1}$ ). Salinity was measured every 2 weeks with a conductivity probe (CDC401, Hach Lange, IntelliCAL, accuracy of 0.1) and remained constant during experiments ( $35.2 \pm 0.2$ ). From  $A_T$  and  $\text{pH}_T$  measurements, dissolved inorganic carbon (DIC), saturation state of seawater with respect to aragonite ( $\Omega_{Ar}$ ) and saturation state of seawater with respect to calcite ( $\Omega_{Ca}$ ) were calculated with [the CO2SYS software](#). Mean temperature and parameters of the carbonate chemistry are given in [Table 42](#).

135 Irradiance was set to the mean *in situ* daily irradiance at 5 m depth in the Bay of Brest according to Martin et al. (2006a): 30-40  $\mu\text{mol photons m}^{-2} \text{s}^{-1}$  in winter and 90-100  $\mu\text{mol photons m}^{-2} \text{s}^{-1}$  in summer. The light was provided by two or four 80 W fluorescent tubes (JBL Solar Ultra Marin Day, JBL Aquaria, Nelson, New Zealand) above the aquaria under a 10/14 h or 14/10 h light/dark photoperiod, for winter or summer conditions, respectively.

### 2.3. Metabolic measurements

140 After three months in experimental conditions, metabolic measurements were performed at the species and assemblage level using incubations in acrylic respirometry chambers (Engineering and Design Plastics Ltd, Cambridge, UK). For species-scale measurements, each species was incubated separately. Community-scale measurements were performed on assemblages, incubating all individuals from all species present in each aquarium. The chamber volume was adapted to species size. It was of 80 mL for *J. exasperatus* and epiphytes, 185 mL for *P. miliaris*, *G. magus* and living and dead *L. corallioides*, and 600 mL for the assemblages. ~~Before incubation~~ [After assemblage incubations](#), epiphytic algae that 145 spontaneously grew on *L. corallioides* during the experiments were carefully removed and incubated separately. Metabolic measurements (net photosynthetic and respiration rates) for the main epiphytic algae *Rhodymenia ardissoni* and *Solieria chordalis* were only examined in the summer, when their biomass was sufficient for measurements. Species were placed on a plastic grid above a stir bar in the chambers to ensure the seawater was well mixed. For *G. magus* and *P. miliaris*, net

150 calcification, respiration and excretion (ammonia release) rates were measured. For *J. exasperatus*, only respiration rates were measured due to its limited size and metabolic rates. For grazers, physiological rates were measured under ambient irradiance. For each grazer species, individuals present in each aquarium were incubated together. For living and dead *L. corallioides* and assemblages, net photosynthetic and light calcification rates were measured under ambient irradiance, and respiration and dark calcification rates were measured in the dark. For light incubations, chambers were placed inside aquaria  
155 to control temperature. For dark incubations, chambers were placed in a plastic crate filled with aquaria seawater in an open circuit to keep the temperature constant. Incubation duration was adjusted to keep oxygen saturation above 80%. Incubations lasted approximately from 1 h for *G. magus* to 2.5 h for dead maerl. For assemblages, the metabolism was measured from the incubations of all species together.

Oxygen concentrations were measured at the beginning and at the end of each incubation, using an optical fiber system  
160 (FIBOX 3, PreSens, Regensburg, Germany). Reactive spots were calibrated with 0% and 100% buffer solutions. The 0% buffer solution was prepared by dissolving 1 g of sodium sulfite (Na<sub>2</sub>SO<sub>3</sub>) in 100 mL of seawater. The 100% buffer solution was prepared by bubbling air into 100 mL of seawater using an air-pump for 20 min to obtain air-saturated seawater. Net primary production (NPP,  $\mu\text{mol O}_2 \text{ g DW}^{-1} \text{ h}^{-1}$ ) or respiration (R,  $\mu\text{mol O}_2 \text{ g DW}^{-1} \text{ h}^{-1}$ ) rates were calculated following Eq. (1):

$$\text{NPP or R} = \frac{\Delta\text{O}_2 \times V}{\Delta t \times \text{DW}} \quad (1)$$

165 where  $\Delta\text{O}_2$  is the difference between the initial and final oxygen concentrations ( $\mu\text{mol O}_2 \text{ L}^{-1}$ ), V the volume of the chamber (L),  $\Delta t$  the incubation time (h), and DW the dry weight of the species incubated (g). The dry weight was obtained after 48 h at 60°C. For gastropods, the body was separated from the shell to consider the dry weight of the body only.

For algae and the assemblages, gross primary production (GPP) was calculated following Eq. (2):

$$\text{GPP} = \text{NPP} - \text{R} \quad (2)$$

170 Control incubations containing only seawater were carried out to correct for oxygen fluxes due to any additional biological activity in seawater. Oxygen fluxes calculated in control chambers were subtracted from oxygen fluxes of chambers containing species.

Seawater samples were taken in the aquaria at the beginning of the incubation and in the chambers at the end of the incubations (except for fleshy algae and *J. exasperatus*) to measure ammonium ( $\text{NH}_4^+$ ) concentration and total alkalinity ( $A_T$ ). To do so, 45 mL seawater samples for  $\text{NH}_4^+$  analyses were fixed with reagent solutions and stored in the dark.  $\text{NH}_4^+$  concentrations were determined according to the Solorzano method (Solorzano, 1969). Absorbance was measured by spectrophotometry at a wavelength of 630 nm (spectrophotometer UV-1201V, Shimadzu Corp, Kyoto, Japan). For grazers, ammonia excretion rates ( $E$ ,  $\mu\text{mol NH}_4^+ \text{ g DW}^{-1} \text{ h}^{-1}$ ) were calculated following Eq. (3):

$$E = \frac{\Delta\text{NH}_4^+ \times V}{\Delta t \times \text{DW}} \quad (3)$$

where  $\Delta\text{NH}_4^+$  is the difference between the initial and final ammonium concentrations ( $\mu\text{mol NH}_4^+ \text{ g DW}^{-1} \text{ h}^{-1}$ ).

For  $A_T$  analyses, 60 mL seawater samples were filtered through 0.7  $\mu\text{m}$  Whatman GF/F filters and were immediately poisoned with a mercuric chloride solution. Total alkalinity was determined according to the method described above. Net calcification rates at light and in the dark ( $G_l$  and  $G_d$ , respectively; in  $\mu\text{mol CaCO}_3 \text{ g DW}^{-1} \text{ h}^{-1}$ ) were calculated according to the alkalinity anomaly technique (Smith and Key, 1975) and corrected for  $\text{NH}_4^+$  fluxes (Gazeau et al., 2015). This correction was applied to calcareous species and assemblage incubations following Eq. (4):

$$G_l \text{ or } G_d = \frac{(-\Delta A_T + \Delta\text{NH}_4^+) \times V}{2 \times \Delta t \times \text{DW}} \quad (4)$$

where  $G_l$  is the net calcification in the light,  $G_d$  is the net calcification in the dark,  $\Delta A_T$  is the difference between the initial and final  $A_T$  ( $\mu\text{eq L}^{-1}$ ).

After the three-month experiments, epiphytic algae that spontaneously grew on *L. corallioides* during experiments were picked off and dried at 60°C for 48 h to determine their dry weight.

#### 2.4. Chlorophyll *a* analysis

At the end of the experiments, thalli of living and dead *L. corallioides* were collected in each aquarium and immediately frozen at -20°C pending analyses. In dead maerl, chlorophyll *a* content was measured in order to check for the presence of associated microflora and potential subsequent metabolism. Then samples were freeze-dried and crushed into a powder using a mortar, in the dark. An aliquot of 0.15 g of powder was precisely weighed and suspended in 10 mL of 90% acetone and stored in the dark at 4°C for 12 h. Samples were then centrifuged at 4000 rpm. The supernatant was collected and absorbance

195 was measured at 630 ( $A_{630}$ ), 647 ( $A_{647}$ ), 664 ( $A_{664}$ ), and 691 ( $A_{691}$ ) nm. Chlorophyll *a* (Chl *a*) concentrations ( $\mu\text{g g DW}^{-1}$ ) were calculated from Ritchie (2008) following Eq. (5):

$$\text{Chl } a = \frac{(-0.3319 A_{630} - 1.7485 A_{647} + 11.9442 A_{664} - 1.4306 A_{691}) \times V}{mp} \quad (5)$$

where *V* is the volume of acetone (mL) and *mp* the mass of powder (g).

## 2.5. Data analysis

200 Comparisons in species and assemblage physiological rates between the winter and summer seasons was performed using t-  
tests, after checking the normality and homogeneity of variances. The influence ~~of season, of~~ temperature and  $\text{pCO}_2$  was  
tested on metabolic rates of grazers (*P. miliaris*, *G. magus* and *J. exasperatus*), living and dead maerl, epiphytic biomass and  
assemblages. Normality of the data and variance homogeneity were checked for all variables. When assumptions were  
respected, two-way ANOVA were performed, using temperature and  $\text{pCO}_2$  as fixed orthogonal factors. When assumptions  
were not respected, two-way non-parametric Scheirer-Ray-Hare tests were run. Statistical analyses were conducted  
205 separately for winter and summer experiments in order to keep a balanced design. When 2-way ANOVAs showed  
significant results, post hoc tests (Tukey honest significant difference, HSD) were performed to compare the four treatments.  
~~Even after transformations, the data were non normally distributed. Therefore, analyses were conducted using a three way~~  
~~permutational multivariate analysis of variance (PERMANOVA), based on Euclidian distance (Anderson, 2001).~~  
~~PERMANOVAs were run with 4999 permutations (Anderson, 2001), using season (two levels: winter and summer),~~  
210 ~~temperature (two levels: ambient and elevated temperature) and  $\text{pCO}_2$  (two levels: ambient and elevated  $\text{pCO}_2$ ) as fixed~~  
~~orthogonal factors ( $n = 5$ ). These statistical analyses were performed with the PRIMER 7 & PERMANOVA+ software~~  
~~package.~~

The effects of  $\text{pCO}_2$  and temperature on the physiological rates of the epiphytic algae *R. ardissoni* and *S. chodalis* were  
only tested in the summer. ~~Because assumptions of normality (Shapiro test) and homogeneity of variances (Bartlett test)~~  
215 ~~were not met, two-way non-parametric Scheirer-Ray-Hare tests were performed. These s~~Statistical analyses were carried out  
using the statistical package R, version 3.2.2.

### 3. Results

#### 3.1. Metabolic responses of grazers to acidification and warming

In the urchin *P. miliaris* and the gastropod *G. magus*, the metabolism was significantly higher in the summer than in the winter, except for *P. miliaris* R, for which no difference was detected (Table 3). In the gastropod *J. exasperatus*, R did not vary with the season. High temperature (+3°C) reduced *P. miliaris* R in the summer, while pCO<sub>2</sub> had no significant effect on *P. miliaris* R (Fig. 1a; Table 4). *P. miliaris* G<sub>1</sub> was significantly affected by the triple interaction between season, temperature and pCO<sub>2</sub> in the summer (Fig. 1b; Supplementary material b), which negated the positive effect of increased temperature or pCO<sub>2</sub> alone, with a negative impact due to the combined effect of temperature and pCO<sub>2</sub> increase in the summer, but no interaction effects in the winter. The combined increase of temperature and pCO<sub>2</sub> *P. miliaris* E was higher under control conditions in the summer and significantly lower under increased temperature (Table 4; Fig. 1c). *G. magus* R was lower under high pCO<sub>2</sub> in the winter only (Table 4; Fig. 1d). Neither temperature nor pCO<sub>2</sub> increases significantly affected *G. magus* R, G<sub>1</sub> and E (Table 4; Fig. 1e-f). In *J. exasperatus*, R was positively affected by the temperature increase increased under elevated temperature but in winter conditions only (Table 4; Fig. 1g). *J. exasperatus* R was negatively influenced by the pCO<sub>2</sub> increase in the winter, but positively in the summer.

#### 3.2. Metabolic responses of living *L. corallioides* to acidification and warming

The metabolism of living *L. corallioides* was higher in the summer than in the winter, except for NPP (Table 3). Living maerl GPP-NPP did not differ among temperature and pCO<sub>2</sub> conditions regardless of the season, while GPP was lower under high temperature in the winter (Table 5; Fig. 2a,b,e). R was significantly reduced by the high temperature condition in the winter, whereas an increase in R was observed in the summer (Table 5; Fig. 2c). No effect of season was observed on chlorophyll *a* content (Tables 3; 6). Chlorophyll *a* content was negatively affected reduced by the high temperature condition in the winter only (Tables 3,4). Temperature had a positive effect on tThe G<sub>1</sub> of living maerl was not significantly influenced by increased temperature and pCO<sub>2</sub>, regardless of the season. Conversely, G<sub>1</sub> was significantly reduced under high pCO<sub>2</sub> (Table 5; Fig. 2d). Conversely, G<sub>a</sub> was significantly affected by an interaction between season and temperature (Fig. 2e). G<sub>a</sub> was positively affected by temperature increased temperature enhanced G<sub>d</sub> in the winter, but no effect was detected in the summer (Fig. 2e). A significant decline in G<sub>d</sub> occurred under high pCO<sub>2</sub> regardless of the season. Net

dissolution, because  $G_d$  was negative, was recorded in the winter under high  $pCO_2$  conditions. This negative effect of increased  $pCO_2$  was alleviated under elevated temperature.

### 3.3. Metabolic responses of dead *L. corallioides* to acidification and warming

245 In dead *L. corallioides*, NPP, GPP, R and  $G_l$  were significantly higher in the summer, while no effect of season was observed on  $G_d$  (Table 3). The high temperature condition (+3°C) did not affect dead maerl NPP, GPP or R (Table 35; Fig. 2fg,h). The  $pCO_2$  increase did not affect dead maerl NPP and GPP in either season. However, ~~there was an interaction between season and  $pCO_2$ , with~~ a decrease in R was observed under high  $pCO_2$  in the summer. Chlorophyll *a* content did not differ between seasons (Tables 3;6) but was significantly affected by the temperature and  $pCO_2$  interaction (Tables 53;4; Supplementary material a;f). Dead maerl  $G_l$  significantly increased under high temperature (Fig. 2i). Conversely, a negative impact of high  $pCO_2$  was on  $G_l$  in the winter and summer. In the dark, net dissolution was observed on dead maerl regardless of the temperature and  $pCO_2$  conditions (Fig. 2j). No temperature effect was observed on dark dissolution. However, dark dissolution rates were significantly higher under high  $pCO_2$  treatments, regardless of the season.

### 3.4. Growth and metabolic responses of epiphytic algae to acidification and warming

255 Mean GPP and R for the two epiphytic algae *R. ardissoni* and *S. chordalis* measured in the summer are presented in Figure 3. *R. ardissoni* NPP and GPP was-were not affected by high temperature or  $pCO_2$  conditions, and R was reduced under high  $pCO_2$  (Table 57; Fig. 3ab,c). In *S. chordalis*, NPP and GPP was-were significantly affected by the interaction between temperature and  $pCO_2$  (Table 57; Fig. 3d,e; Supplementary material c;d). R was enhanced by the high temperature and  $pCO_2$  conditions ~~(Fig. 3f)~~ and their combination resulted in a greater R (Fig. 3f).

260 The mean biomass of epiphytic fleshy algae at the end of the experiment was significantly higher in the summer than in the winter (+81%, t-test,  $p=0.0453$  way PERMANOVA,  $df=1$ ,  $F=5.3$ ,  $p=0.027$ , Fig. 4). Epiphyte biomass was not affected by increased temperature or  $pCO_2$  in the winter (2-way ANOVA,  $p=0.95$  and  $0.67$  respectively), while an interactive effect of temperature and  $pCO_2$  was observed in the summer ( $p=0.013$ , supplementary material e). ~~significantly affected by the triple interaction between season, temperature and  $pCO_2$  (3-way PERMANOVA,  $df=1$ ,  $F=4.9$ ,  $p=0.035$ ), with high  $pCO_2$~~

265 ~~having a positive effect on epiphyte biomass in the winter. In the summer, this positive effect was only detected under high temperature.~~

### 3.5. Metabolic responses of assemblages to acidification and warming

270 ~~Assemblage metabolism was significantly higher in the summer than in the winter regardless of the metabolic parameter tested (Table 3). No temperature effect was observed on NPP, GPP and R in either season (Table 68; Fig. 5a-c). The high pCO<sub>2</sub> condition enhanced NPP in both seasons. ~~The combined effect of season and pCO<sub>2</sub> affected GPP, with a~~ High pCO<sub>2</sub> increased positive effect GPP of pCO<sub>2</sub> increase in the summer only. Similarly, R significantly increased under high pCO<sub>2</sub> in summer conditions. ~~An interactive effect of season and temperature was detected for G<sub>1</sub>, which~~ increased under high temperature in the summer only (Fig. 5d). Conversely, high pCO<sub>2</sub> reduced G<sub>1</sub> regardless of the season. In the dark, net dissolution was observed in the winter, but net precipitation occurred in summer conditions at high temperature (Fig. 5e). ~~G<sub>d</sub> was significantly affected by the triple interaction between season, temperature and pCO<sub>2</sub>.~~ In the winter, high pCO<sub>2</sub> increased net dissolutions rates, ~~and high temperature reduced them,~~ while in the summer, G<sub>d</sub> increased under elevated temperature, the interactive effect of temperature and pCO<sub>2</sub> increase was more complex, with a decrease in G<sub>d</sub> detected under high temperature conditions only.~~

## 4. Discussion

280 ~~Our study demonstrates that t~~ The response of maerl bed marine communities to increased temperature and pCO<sub>2</sub> conditions is likely to be a complex function of direct effects of climate variables on species physiology and shifts in species interactions (Lord et al., 2017). Results show that predicted changes may alter interactions among calcifying and fleshy macroalgae via overgrowth of epiphytic algae and an increase in competition for light and nutrients with underlying maerl. Interactions between grazers and macroalgae were also affected because the grazer physiology was adversely affected by acidification and warming with potential consequences on epiphyte biomass regulation. Our results underscore the importance of examining community-level processes to integrate species interactions in the study of the impact of global change on marine ecosystems. This study also evidences that seasonal variability represent an important driver influencing the magnitude and the direction of species and community response to climate change.

290 Assemblage exhibited a strong seasonal pattern for all metabolic parameters, which is consistent with the higher metabolism  
in the summer for most of the species incubated at the specific scale. This higher metabolism in the summer has already been  
evidenced in urchins (Brockington and Peck, 2001), gastropods (Davies, 1966; Innes and Houlihan, 1985; Martin et al.,  
2006b) and living maerl (Potin et al., 1990; Martin et al., 2006a) and is strongly related to changes in numerous  
environmental and biological variables, such as light intensity and photoperiod, temperature and nutrient or food availability  
(Godbold and Solan, 2013; Thomsen et al., 2013).

295 Assemblage GPP and R were not affected by the high temperature and pCO<sub>2</sub> conditions in the winter. Conversely, in the  
summer, GPP and R increased under high pCO<sub>2</sub> conditions. The response of assemblage GPP and R appeared closely related  
to changes in epiphyte biomass and productivity. For instance, the biomass of maerl epiphytic algae was significantly higher  
in summer than in winter, which is consistent with other findings in the Bay of Brest (Guillou et al., 2002) and other Atlantic  
maerl beds (Peña and Barbara, 2010). ~~For instance, the~~ The high biomass of epiphytic algae in the summer led to high  
300 contribution to oxygen fluxes. Under high pCO<sub>2</sub> conditions, the higher availability of CO<sub>2</sub> as substrate for photosynthesis  
may stimulate epiphyte productivity and growth (Koch et al., 2013). The two main epiphytic algae that grew during the  
experiments, *R. ardissoni* and *S. chordalis*, are naturally found in maerl beds in Brittany (Peña et al., 2014). The response of  
the alga *S. chordalis* to increased temperature and pCO<sub>2</sub> differed from that of *R. ardissoni*. This difference suggests that the  
response is species-specific, even among fleshy algae, as demonstrated by Kram et al. (2016). *R. ardissoni* GPP was not  
305 affected by increased temperature and pCO<sub>2</sub>, but its R was significantly lower under high pCO<sub>2</sub>. Within the same genus,  
Cook et al. (1986) showed that *Rhodymenia palmata* can potentially use HCO<sub>3</sub><sup>-</sup> as source of inorganic carbon for  
photosynthesis. The same process may occur in *R. ardissoni*, suggesting that this alga is not carbon-limited at current  
oceanic pCO<sub>2</sub> levels. In contrast to *R. ardissoni*, increased pCO<sub>2</sub> stimulated *S. chordalis* GPP under ambient conditions of  
temperature. In their study, Short et al. (2014) indicate that the overgrowth of filamentous algae occurs synergistically with  
310 high pCO<sub>2</sub> levels and decreased photosynthesis in coralline algae. Here, the stimulation of epiphyte productivity and growth  
under high pCO<sub>2</sub> is likely to increase the competition with underlying maerl, especially through reduction in incident light.  
Although assemblages were mainly composed of living and dead maerl, the response of GPP and R of *L. corallioides* to  
increased temperature and pCO<sub>2</sub> differed from that observed in assemblages. For example, the temperature increase of +3°C

reduced living *L. corallioides* R in the winter, but increased R in the summer. Under high pCO<sub>2</sub> conditions, although CO<sub>2</sub> availability for photosynthesis was higher, no difference was observed in *L. corallioides*, probably due to the ability of this species to employ inorganic carbon acquisition mechanisms (Kübler and Dudgeon, 2015). Interestingly, GPP, R and chlorophyll *a* content of dead maerl were of the same magnitude as for living maerl. Although live algae prevent bio-fouling by shedding their surface layers (Keats et al., 1997; Villas Bôas and Figueiredo, 2004), post-mortem colonization by photosynthetic endolithic assemblages may occur within dead crusts (Diaz-Pulido et al., 2012). Moreover, dead thalli may represent a substrate for the settlement of crustose coralline algae that cover small parts of some thalli. Crustose coralline algae colonization may also contribute to the observed GPP and R values. In dead maerl, only R decreased under high pCO<sub>2</sub>, while no effect was detected for GPP.

These findings also suggest the importance of dead maerl to assemblage carbonate fluxes during the experiments. For example, endolithic algae appear to play an important role in the dissolution of a crustose coralline alga (CCA) species, *Porolithon onkodes* (Reyes-Nivia et al., 2014). Through their photosynthesis, endolithic algae may elevate interstitial pH within the *P. onkodes* skeleton (Reyes-Nivia et al., 2013), increasing carbonate cement precipitation (Diaz-Pulido et al., 2014). Within dead *L. corallioides*, the presence of endolithic algae combined with the presence of small patches of CCA on the surface of thalli may explain the calcification rates observed in light and dissolution in dark. Considering the high Mg content in the skeleton of *L. corallioides*, increased pCO<sub>2</sub> likely promotes the dissolution of dead thalli. Alternatively, the increase in dissolution observed in the present study may be associated with a reduction of CCA recruitment over the surface of dead thalli under acidified conditions (Jokiel et al., 2008; [Ordoñez et al., 2014](#)). These results are consistent with the negative response to increased pCO<sub>2</sub> observed here in assemblage G<sub>l</sub> and G<sub>d</sub> values, which appeared strongly related to the response of living maerl calcification rates. The high sensitivity of coralline algae to ocean acidification has already been attributed to their high Mg-calcite content (Morse et al., 2006; Hofmann and Bischof, 2014). In the present study, the pCO<sub>2</sub> increase had adverse consequences on assemblage G<sub>d</sub>, both in the winter and summer. In the dark, assemblage R reduced seawater pH by releasing CO<sub>2</sub>, and hindered the precipitation of CaCO<sub>3</sub> (Cornwall et al., 2013). Under high pCO<sub>2</sub> conditions, the combined effect of acidification and assemblage R in the dark is likely to increase the sensitivity of living and dead *L. corallioides* to dissolution (Andersson et al., 2009).

Moreover, as discussed above, the overgrowth of epiphytic algae under high pCO<sub>2</sub> increased assemblage R in the dark.

340 Therefore, the negative effect of ocean acidification on *L. corallioides* G<sub>d</sub> would be exacerbated by the presence of epiphytic algae, which promote a decline in pH in the dark. In light, several studies have suggested that moderate growth of fleshy macroalgal communities may reduce the impact of ocean acidification on coralline calcification by reducing the CO<sub>2</sub> concentration of seawater through photosynthesis (Semese et al., 2009; Short et al., 2014). Conversely, other studies evidenced that the overgrowth of epiphytic fleshy algae may shade underlying coralline algae and reduce coralline net

345 calcification rates (Garrabou and Ballesteros, 2000; Martin and Gattuso, 2009). ~~However,~~ The present findings ~~do not~~ support this idea, because a decline in assemblage G<sub>1</sub> was observed under high pCO<sub>2</sub> ~~despite and~~ high epiphyte biomass. Under high pCO<sub>2</sub>, the overgrowth of epiphytic fleshy algae induced by ocean acidification in the summer may reduce light, oxygen and nutrient availability for underlying maerl, affecting its primary production and calcification (D'Antonio, 1985; Short et al., 2014). Thus, overgrown maerl would be negatively affected by the direct effect of ocean acidification on

350 calcification rates and indirect effects due to shifts in competition dynamics with fleshy epiphytic algae (Kuffner et al., 2008). However, the response of epiphytic algae is likely to be specie-specific and it appears difficult to generalize the impacts of epiphytic algae on coralline algae.

In regard to the present results, the regulation of epiphyte biomass by grazers appears essential to maintain the proper functioning of maerl bed communities (Guillou et al., 2002). The importance of grazers to control epiphytes growth in

355 aquaria has been evidenced by Jokiel et al. (2008). In mollusks and urchins, several studies have demonstrated a link between feeding rates and other metabolic processes, such as respiration, calcification and excretion (Carr and Bruno, 2013; Navarro et al., 2013; Noisette et al., 2016). In mollusks, a wide range of responses to ocean acidification and warming have been revealed (Gazeau et al., 2013; Parker et al., 2013). The differences in sensitivity of mollusks to ocean acidification depend on several parameters, such as the form of CaCO<sub>3</sub> they precipitate during calcification (Ries et al., 2009), as well as

360 their ability to regulate the acid-base balance (Gutowska et al., 2010). Our results corroborate these studies, given that *G. magus* and *J. exasperatus* responded differently to acidification and warming. Increased temperature and pCO<sub>2</sub> had no effect on *G. magus* with regard to the metabolic functions tested. However, despite the apparent resistance of *G. magus* to the applied changes, other physiological parameters that we did not test here may have been affected, such as feeding rates,

somatic growth, enzyme activity or immune response (Parker et al., 2013). The respiration rates of *J. exasperatus* showed a  
365 decline under high pCO<sub>2</sub> in the winter. The lower growth of epiphytes and biofilm in winter may reduce the energy available  
to maintain the metabolism under stressful conditions (Thomsen et al., 2013; Pansch et al., 2014). This reduced energy  
availability may induce changes in energy partitioning and decrease R under high pCO<sub>2</sub>. In the summer, the increased R  
under high pCO<sub>2</sub> can be attributed to higher food supply, which is likely to increase the resistance of *J. exasperatus* to  
climate change, as reported for several marine taxa (Ramajo et al., 2016).

370 Given the relatively high resistance of *G. magus* and *J. exasperatus* to predicted changes, the metabolic response of *P.*  
*miliaris* appears to have stronger implications on assemblage functioning. For example, *P. miliaris* is considered as one of  
the main macro-epiphytic grazers on maerl beds in the Bay of Brest (Guillou et al., 2002). During the experiments, *P.*  
*miliaris* likely played an important role in the regulation of epiphytic biomass. The response of G<sub>1</sub> to temperature and pCO<sub>2</sub>  
changes was complex. The interaction between temperature and pCO<sub>2</sub> observed in the summer may cause changes in energy  
375 partitioning, thereby inducing a trade-off between metabolic processes at the expense of respiration and excretion (Garilli et  
al., 2015). However, the effect of temperature and pCO<sub>2</sub> on the calcification of *P. miliaris* must be considered carefully. For  
instance, urchins defecated carbonate pellets following consumption of maerl thalli. These feces are likely to dissolve during  
incubation, introducing a bias in the measurement of calcification (Gazeau et al., 2015). In the summer, temperature increase  
by 3°C reduced *P. miliaris* respiration rates. Moreover, the decrease in excretion under high temperature and pCO<sub>2</sub>  
380 conditions was modulated by the interaction between these two factors. Temperature is a major factor affecting physiological  
processes in ectotherms such as metabolic rates and growth (Kordas et al., 2011). In *P. miliaris*, summer temperatures are  
likely to exceed the physiological thresholds of organisms, inducing a metabolic decline when maintained at 20°C. Although  
this decline has only been measured for respiration and excretion, the increase in temperature is also likely to affect sea  
urchin feeding efficiency (Thomas et al., 2000; Carr and Bruno, 2013). Therefore, the ability of *P. miliaris* to regulate  
385 epiphyte biomass may be significantly altered under predicted acidification and warming conditions.

In addition to the impact of climate change on grazer-fleshy macroalgae interactions, predicted changes may also  
considerably alter the interaction between grazers and coralline algae. Asnaghi et al. (2013) demonstrated that the grazing  
activity by urchins may exacerbate pCO<sub>2</sub> effects on coralline algae. Ocean acidification may alter the structural integrity of

390 coralline algae, increasing its sensitivity to grazing (Johnson and Carpenter, 2012; Ragazzola et al., 2012). Coralline algae  
may thus be more susceptible to grazing by urchins, which also benefit from a higher carbonate uptake from their diet to  
modulate their response to ocean acidification (Asnaghi et al., 2013). In *L. corallioides*, the decrease in calcification rates  
may alter its structural integrity and increase its susceptibility to grazing, especially by urchins, which are considered as  
important bioeroders of coralline algae in marine ecosystems (Ballesteros, 2006; O'Leary and McClanahan, 2010),  
particularly in maerl beds (Lawrence, 2013).

395 In conclusion, the community response to climate change does not appear to be only the result of individual species'  
metabolic responses, but also strongly depends on shifts in species interactions. In contrast with other studies, which  
evidenced larger impacts of the combination of increased pCO<sub>2</sub> and temperature than that of these factors alone (Reynaud et  
al., 2003; Anthony et al., 2008; Martin and Gattuso, 2009; Rodolfo-Metalpa et al., 2010), we showed here that the effects of  
pCO<sub>2</sub> and temperature on maerl bed communities were weakened when these factors were combined.~~we showed here that  
the effects of pCO<sub>2</sub> and temperature on maerl bed communities were weakened when these factors were combined. Our  
results suggest that ocean acidification and warming will strongly destabilize communities through both direct effects on  
species physiology and changes in the interaction strengths between coralline algae, fleshy algae and grazers.~~ Under the  
predicted business-as-usual conditions, epiphyte overgrowth may exacerbate the negative impact of climate change on  
underlying coralline algae. Here, we also demonstrated that climate change may affect grazer physiology, with major  
405 consequences on their ability to regulate epiphyte biomass. Climate change may also affect other components that we did not  
assess in the present study, such as algal palatability and potential changes in grazer trophic behavior (Campbell et al., 2014;  
Duarte et al., 2015; Poore et al., 2013; Poore et al., 2016). Algal palatability to grazers may also be affected by predicted  
changes through shifts in the composition and the quantity of allelopathic compounds, as suggested by Del Monaco et al.  
(2017). In order to better understand the consequences of climate change on ecosystem functioning~~line with this study,  
410 further work should focus on the ~~impact of climate change on~~ response of marine communities ~~and species interactions and  
consider more specifically shifts in species interactions, including changes in trophic interactions between algae and grazers.  
to better understand the consequences on ecosystem functioning.~~~~

## Authors' Contributions

EL SM PR JG JC designed the experiments; EL SM JC collected the data; EL ML analyzed the data; EL SM PR JG  
415 prepared the manuscript with contributions from all co-authors.

## Competing interests

The authors declare that they have no conflict of interest.

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Table 1. Summary of the four experimental treatments. Two pCO<sub>2</sub> (ambient and high pCO<sub>2</sub>) and temperature (ambient and high temperature) conditions were tested. High pCO<sub>2</sub> (H-pCO<sub>2</sub>) corresponded to a pH decrease of -0.33 units compared to ambient conditions (A-pCO<sub>2</sub>). High temperature (T + 3°C) corresponded to a temperature increase of 3°C compared to ambient conditions (T).

	<u>pCO<sub>2</sub></u>	<u>Temperature</u>	
<u>1 (Control)</u>	<u>Ambient (A-pCO<sub>2</sub>)</u>	<u>Ambient (T)</u>	<u>A-pCO<sub>2</sub>; T</u>
<u>2</u>	<u>High (H-pCO<sub>2</sub>)</u>	<u>Ambient (T)</u>	<u>H-pCO<sub>2</sub>; T</u>
<u>3</u>	<u>Ambient (A-pCO<sub>2</sub>)</u>	<u>High (T+3°C)</u>	<u>A-pCO<sub>2</sub>; T + 3°C</u>
<u>4</u>	<u>High (H-pCO<sub>2</sub>)</u>	<u>High (T+3°C)</u>	<u>H-pCO<sub>2</sub>; T + 3°C</u>

Table 2. Physicochemical parameters (mean  $\pm$  SE) of seawater in each experimental condition (A-pCO<sub>2</sub> = ambient pCO<sub>2</sub>; H-pCO<sub>2</sub> = high pCO<sub>2</sub>; T = ambient temperature; T+3°C = high temperature) in the winter and the summer. pH<sub>T</sub> and temperature were monitored every two days in each aquarium (n = 35). Total alkalinity values (A<sub>T</sub>) are means ( $\pm$  SE) of 28 samples measured in each aquarium. The CO<sub>2</sub> partial pressure (pCO<sub>2</sub>), dissolved inorganic carbon (DIC), and saturation states of seawater with respect to aragonite ( $\Omega_{Ar}$ ) and calcite ( $\Omega_{Ca}$ ) were calculated from pH<sub>T</sub>, temperature, salinity, and A<sub>T</sub> using CO2SYS.

	Experimental condition	pCO <sub>2</sub> (μatm)	pH <sub>T</sub>	Temperature (°C)	A <sub>T</sub> (μmol kg <sup>-1</sup> )	DIC (μmol kg <sup>-1</sup> )	$\Omega_{Ar}$	$\Omega_{Ca}$
<b>WINTER</b>	A-pCO <sub>2</sub> ; T	490 ( $\pm$ 5)	7.97 ( $\pm$ 0.04)	10.1 ( $\pm$ 0.3)	2348 ( $\pm$ 6)	2189 ( $\pm$ 6)	1.84 ( $\pm$ 0.02)	2.89 ( $\pm$ 0.02)
	H-pCO <sub>2</sub> ; T	1183 ( $\pm$ 10)	7.63 ( $\pm$ 0.03)	10.1 ( $\pm$ 0.3)	2342 ( $\pm$ 7)	2306 ( $\pm$ 7)	0.89 ( $\pm$ 0.01)	1.40 ( $\pm$ 0.01)
	A-pCO <sub>2</sub> ; T+3°C	513 ( $\pm$ 5)	7.97 ( $\pm$ 0.03)	13.7 ( $\pm$ 0.1)	2341 ( $\pm$ 5)	2166 ( $\pm$ 5)	2.01 ( $\pm$ 0.01)	3.14 ( $\pm$ 0.02)
	H-pCO <sub>2</sub> ; T+3°C	1087 ( $\pm$ 18)	7.64 ( $\pm$ 0.03)	13.6 ( $\pm$ 0.2)	2329 ( $\pm$ 2)	2266 ( $\pm$ 4)	1.09 ( $\pm$ 0.01)	1.70 ( $\pm$ 0.02)
<b>SUMMER</b>	A-pCO <sub>2</sub> ; T	426 ( $\pm$ 4)	8.03 ( $\pm$ 0.04)	17.1 ( $\pm$ 0.2)	2359 ( $\pm$ 3)	2127 ( $\pm$ 3)	2.60 ( $\pm$ 0.02)	4.03 ( $\pm$ 0.03)
	H-pCO <sub>2</sub> ; T	948 ( $\pm$ 9)	7.72 ( $\pm$ 0.03)	17.1 ( $\pm$ 0.2)	2382 ( $\pm$ 4)	2279 ( $\pm$ 4)	1.45 ( $\pm$ 0.01)	2.24 ( $\pm$ 0.02)
	A-pCO <sub>2</sub> ; T+3°C	432 ( $\pm$ 4)	8.01 ( $\pm$ 0.04)	20.0 ( $\pm$ 0.5)	2364 ( $\pm$ 3)	2109 ( $\pm$ 3)	2.88 ( $\pm$ 0.02)	4.43 ( $\pm$ 0.03)
	H-pCO <sub>2</sub> ; T+3°C	879 ( $\pm$ 7)	7.74 ( $\pm$ 0.02)	20.2 ( $\pm$ 0.3)	2369 ( $\pm$ 2)	2238 ( $\pm$ 2)	1.71 ( $\pm$ 0.01)	2.64 ( $\pm$ 0.02)

Table 3. Results of mean comparison tests between seasons for the net and gross primary production, respiration, chlorophyll *a* content, light and dark calcification and excretion of the different species and the assemblages. Statistical analyses were performed using t-tests.

	<u>Net production</u> <u>NPP</u>			<u>Gross production</u> <u>GPP</u>			<u>Respiration</u> <u>R</u>			<u>Chlorophyll a</u>			<u>Light calcification</u> <u>G<sub>l</sub></u>			<u>Dark calcification</u> <u>G<sub>d</sub></u>			<u>Excretion</u> <u>E</u>		
	df	t	p	df	t	p	df	t	p	df	t	p	df	t	p	df	t	p	df	t	p
<i>P. miliaris</i>							24	-11.6	<0.001				24	1.5	0.16				38	3.5	0.001
<i>G. magus</i>							27	-20.6	<0.001				19	5.3	<0.001				29	14.1	<0.001
<i>J. exasperatus</i>							38	0.7	0.46												
Living <i>L. corallioides</i>	38	1.4	0.16	38	5.7	<0.001	38	-12.7	<0.001	38	0.4	0.66	25	8.3	<0.001	38	6.6	<0.001			
Dead <i>L. corallioides</i>	26	7.5	<0.001	24	8.4	<0.001	22	-9.8	<0.001	37	0.9	0.35	38	4.2	<0.001	28	0.3	0.80			
Assemblage	31	-4.5	<0.001	31	6.1	<0.001	37	-13.1	<0.001				38	9.0	<0.001	26	3.3	0.003			

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Table 24 Results the analysis of variance testing the effects of temperature (T) and pCO<sub>2</sub> on respiration, net calcification and excretion rates in the urchin *Psammechinus miliaris* and the two gastropods *Gibbula magus* and *Jujubinus exasperatus* for winter and summer experiments (n = 5).

Statistical analyses were performed using 2-way crossed ANOVAs and Scheirer-Ray-Hare tests when test assumptions were not respected (in italic). Significant p-values are shown in bold ( $\alpha = 0.05$ ). Degrees of freedom = 1

		Respiration		Net Calcification		Excretion	
		R		G		E	
<i>Psammechinus miliaris</i>	<b>WINTER</b>	F	p	F	p	F	p
	T	1.8	0.18	0.3	0.59	0.0	0.89
	pCO <sub>2</sub>	0.0	0.87	1.0	0.34	1.2	0.27
	pCO <sub>2</sub> x T	0.2	0.67	0.2	0.64	1.6	0.22
	<b>SUMMER</b>						
	T	20.8	<b>&lt;0.001</b> ∇	4.8	<b>0.045</b>	7.6	<b>0.014</b> ∇
	pCO <sub>2</sub>	3.9	0.07	0.1	0.82	2.7	0.12
pCO <sub>2</sub> x T	1.4	0.25	6.6	<b>0.022</b>	3.1	0.10	
<i>Gibbula magus</i>	<b>WINTER</b>	F	p	F	p	F	p
	T	1.1	0.30	0.0	0.90	0.4	0.55
	pCO <sub>2</sub>	4.8	<b>0.043</b> ∇	0.1	0.79	0.6	0.44
	pCO <sub>2</sub> x T	0.0	0.88	3.9	0.07	1.3	0.28
	<b>SUMMER</b>						
	T	0.0	0.93	0.8	0.38	2.4	0.15
	pCO <sub>2</sub>	0.2	0.68	0.6	0.45	1.8	0.20
pCO <sub>2</sub> x T	0.1	0.72	0.4	0.55	0.1	0.75	
<i>Jujubinus exasperatus</i>	<b>WINTER</b>	F	p				
	T	8.6	<b>0.010</b> ↗				
	pCO <sub>2</sub>	5.6	<b>0.031</b> ∇				
	pCO <sub>2</sub> x T	0.8	0.39				
	<b>SUMMER</b>						
	T	0.1	0.75				
	pCO <sub>2</sub>	8.9	<b>0.009</b> ↗				
pCO <sub>2</sub> x T	0.8	0.83					

Table 35. Summary of Results the analysis of variance PERMANOVA for the effects of season temperature (T) and pCO<sub>2</sub> on net and gross primary production, respiration, chlorophyll *a* content and light and dark calcification rates of living and dead *Lithothamnion corallioides* (n = 5).

Statistical analyses were performed using 2-way crossed ANOVAs and Scheirer-Ray-Hare tests when test assumptions were not respected (in *italic*). Significant p-values are shown in bold ( $\alpha = 0.05$ ). Degrees of freedom = 1

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		Net production		Gross production		Respiration		Chlorophyll a		Light calcification		Dark calcification	
		<i>NPP</i>		<i>GPP</i>		<i>R</i>				<i>G<sub>l</sub></i>		<i>G<sub>d</sub></i>	
		F	p	F	p	F	p	F	p	F	p	F	p
LIVING <i>L. corallioides</i>	WINTER												
	T	4.4	0.052	8.0	<b>0.012</b> ▽	13.1	<b>0.002</b> ▽	5.9	<b>0.027</b> ▽	3.6	0.08	10.0	<b>0.006</b> ↗
	pCO <sub>2</sub>	0.6	0.44	0.6	0.44	0.1	0.73	1.6	0.23	3.2	0.09	153.3	<b>&lt;0.001</b> ▽
	pCO <sub>2</sub> x T	3.2	0.09	1.0	0.33	3.8	0.07	0.0	0.89	0.8	0.39	3.6	0.08
	SUMMER												
	T	2.0	0.18	1.9	0.17	20.9	<b>&lt;0.001</b> ↗	1.5	0.23	3.3	0.07	0.0	0.98
pCO <sub>2</sub>	0.8	0.40	0.4	0.55	0.2	0.70	0.7	0.41	3.6	0.06	50.0	<b>&lt;0.001</b> ▽	
pCO <sub>2</sub> x T	1.5	0.24	0.7	0.41	0.4	0.52	0.6	0.45	0.2	0.65	0.0	0.96	
DEAD <i>L. corallioides</i>	WINTER												
	T	0.0	0.91	0.0	0.97	0.0	0.99	0.0	0.99	20.2	<b>&lt;0.001</b> ↗	0.1	0.72
	pCO <sub>2</sub>	0.1	0.71	0.1	0.76	0.1	0.73	0.1	0.81	61.1	<b>&lt;0.001</b> ▽	99.6	<b>&lt;0.001</b> ▽
	pCO <sub>2</sub> x T	0.3	0.59	0.0	0.93	0.9	0.35	6.3	<b>0.024</b>	0.0	0.88	0.0	0.85
	SUMMER												
	T	3.7	0.07	3.7	0.07	2.0	0.17	0.7	0.41	0.2	0.65	0.2	0.64
pCO <sub>2</sub>	1.1	0.31	1.8	0.20	4.8	<b>0.043</b> ▽	9.9	<b>0.006</b>	9.6	<b>0.002</b> ▽	17.8	<b>&lt;0.001</b> ▽	
pCO <sub>2</sub> x T	0.9	0.36	0.8	0.39	0.2	0.67	30.3	<b>&lt;0.001</b>	1.9	0.17	0.6	0.44	

690 | Table 46. Chlorophyll *a* content (mean  $\pm$  SE) of living and dead *L. corallioides* in the different pCO<sub>2</sub> (A-pCO<sub>2</sub> = ambient pCO<sub>2</sub>; H-pCO<sub>2</sub> = high pCO<sub>2</sub>) and temperature (T = ambient temperature; T+3°C = high temperature) treatments, after being maintained three months in winter and summer conditions, n = 5

	<b>Chlorophyll <i>a</i></b>			
	<b>μg chlorophyll g DW<sup>-1</sup></b>			
	A-pCO <sub>2</sub> /T	H-pCO <sub>2</sub> /T	A-pCO <sub>2</sub> /T+3°C	H-pCO <sub>2</sub> /T+3°C
<b>Living <i>L. corallioides</i></b>				
Winter	59.84 (± 1.97)	61.66 (± 3.83)	52.93 (± 3.44)	56.85 (± 2.52)
Summer	55.03 (± 2.95)	57.63 (± 3.99)	60.35 (± 0.70)	62.19 (± 3.75)
<b>Dead <i>L. corallioides</i></b>				
Winter	47.09 (± 2.72)	39.39 (± 5.65)	39.15 (± 2.20)	46.36 (± 2.19)
Summer	52.21 (± 1.92)	36.30 (± 1.83)	43.63 (± 0.90)	47.96 (± 2.54)

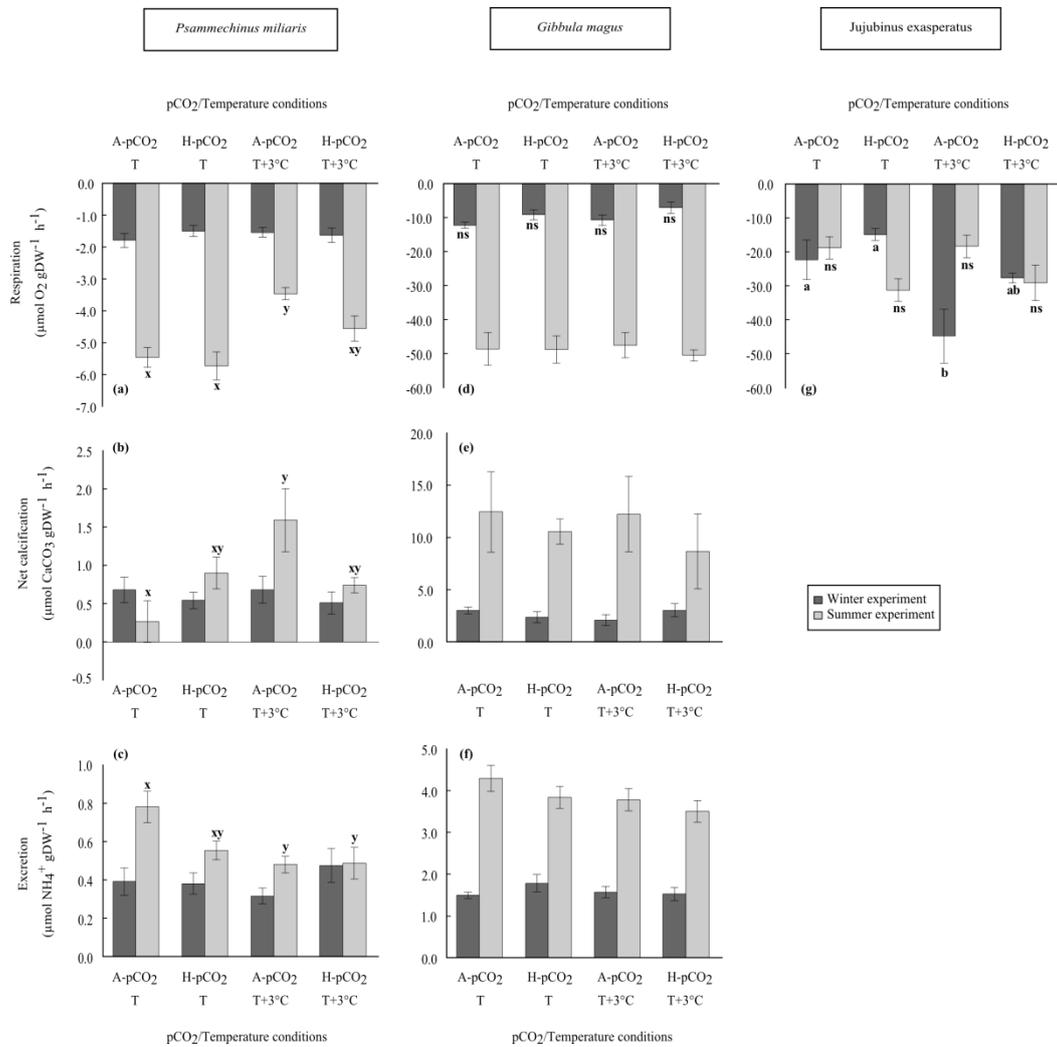
Table 57. Summary of the effects of pCO<sub>2</sub> and temperature (T) and their combined effect on gross production and respiration of the two epiphytic algae *R. ardissonaei* and *S. chordalis* in the summer (n = 5). Statistical analyses were performed using a two-way crossed Scheirer-Ray-Hare test.

695 Significant p-values are presented in bold ( $\alpha = 0.05$ ). Degrees of freedom = 1

		Net production <b>NPP</b>		Gross production <b>GPP</b>		Respiration <b>R</b>	
		F	p-value	F	p-value	F	p-value
<i>Rhodymenia ardissonaei</i>	T	0.8	0.37	0.2	0.68	1.3	0.25
	pCO <sub>2</sub>	0.0	0.96	0.8	0.38	8.6	<b>0.003</b> ↘
	pCO <sub>2</sub> x T	1.0	0.31	1.0	0.31	0.7	0.42
<i>Solieria chordalis</i>		F	p-value	F	p-value	F	p-value
	T	0.1	0.76	0.1	0.60	5.5	<b>0.019</b> ↗
	pCO <sub>2</sub>	3.0	0.082	5.5	<b>0.019</b>	3.9	<b>0.049</b> ↗
	pCO <sub>2</sub> x T	5.8	<b>0.016</b>	5.5	<b>0.019</b>	0.0	0.88

Table 68. Results the analysis of variance testing the effects of temperature (T) and pCO<sub>2</sub> on net and gross primary production, respiration and light and dark calcification rates, measured on assemblages (n = 5). Significant p-values are presented in bold ( $\alpha = 0.05$ ). Degrees of freedom = 1

		Net production		Gross production		Respiration		Light calcification		Dark calcification	
		<b>NPP</b>		<b>GPP</b>		<b>R</b>		<b>G<sub>l</sub></b>		<b>G<sub>d</sub></b>	
Assemblages	<b>WINTER</b>	F	p	F	p	F	p	F	p	F	p
	T	0.2	0.70	0.7	0.43	1.7	0.21	0.2	0.71	0.7	0.43
	pCO <sub>2</sub>	8.6	<b>0.011 ↗</b>	0.1	0.72	1.6	0.23	27.3	<b>&lt;0.001 ↘</b>	65.4	<b>&lt;0.001 ↘</b>
	pCO <sub>2</sub> x T	0.9	0.35	1.1	0.31	1.1	0.31	0.8	0.37	0.2	0.66
	<b>SUMMER</b>										
	T	2.1	0.17	1.6	0.23	0.5	0.51	40.2	<b>&lt;0.001 ↗</b>	6.8	<b>0.020 ↗</b>
	pCO <sub>2</sub>	8.2	<b>0.011 ↗</b>	14.2	<b>0.002 ↗</b>	11.1	<b>0.004 ↗</b>	16.6	<b>&lt;0.001 ↘</b>	3.0	0.10
	pCO <sub>2</sub> x T	1.9	0.19	1.3	0.27	0.3	0.60	0.8	0.38	3.6	0.08



700 | Fig. 1. Respiration, net calcification and excretion rates (mean ± SE, n = 5) of the grazers *P. miliaris* (a to c), *G. magus* (d to f) and respiration of *J. exasperatus* (g) in the different pCO<sub>2</sub> (A-pCO<sub>2</sub> = Ambient pCO<sub>2</sub>; H-pCO<sub>2</sub> = High-pCO<sub>2</sub>) and temperature (T = Ambient temperature; T+3°C = High temperature) conditions. The species were maintained in assemblages for three months in winter (dark gray) and summer conditions (light gray). Different letters show significant differences (Tukey HSD test) between the four treatments in the winter (letters a and b) and summer (letters x and y). ns = not significant. Tukey tests were performed when a significant effect of temperature or pCO<sub>2</sub> was detected using 2-way ANOVAs.

705 |

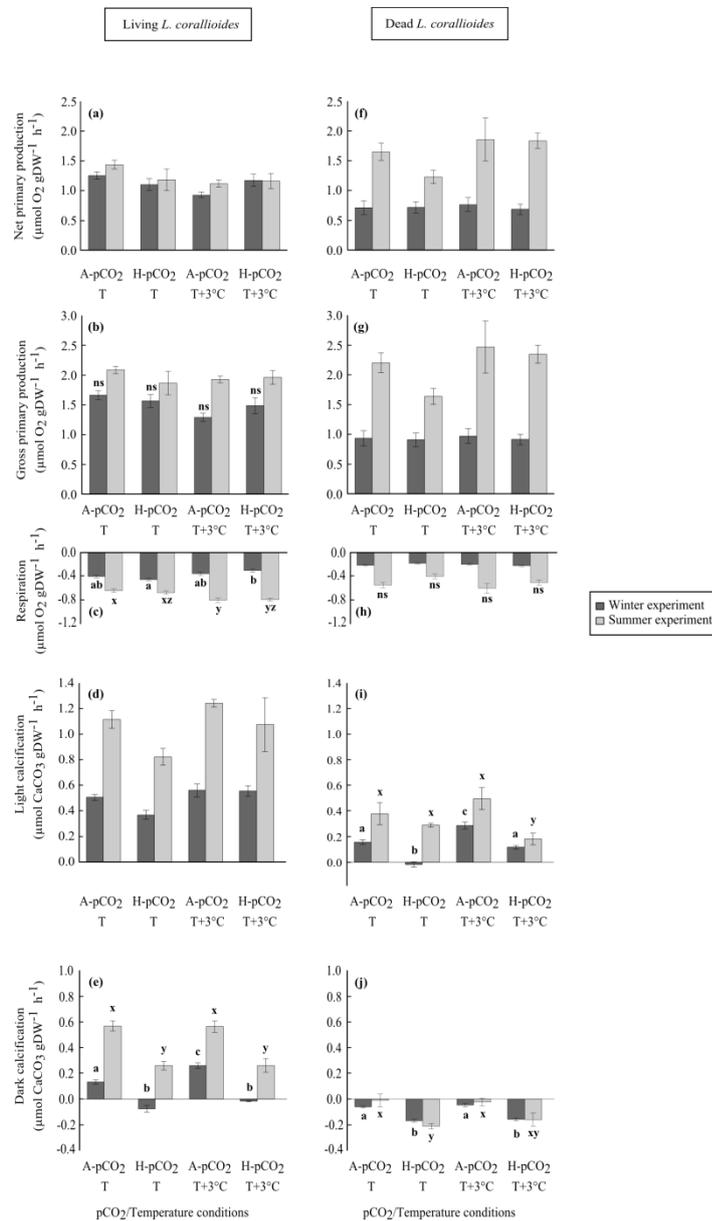


Fig. 2. Net and gross primary production, respiration, light and dark calcification rates (mean  $\pm$  SE,  $n = 5$ ) of living (a to e) and dead thalli (f to j) of *L. corallioides* in the different pCO<sub>2</sub> (A-pCO<sub>2</sub> = Ambient pCO<sub>2</sub>; H-pCO<sub>2</sub> = High-pCO<sub>2</sub>) and temperature (T = Ambient temperature; T+3°C = High temperature) treatments, after three months in winter (dark gray) and summer conditions (light gray). Letters indicate significant differences between the four treatments in winter (a, b, c) and summer (x, y, z) conditions (Tukey HSD test). ns = not significant. Tukey tests were performed when a significant effect of temperature or pCO<sub>2</sub> was detected using 2-way ANOVAs.

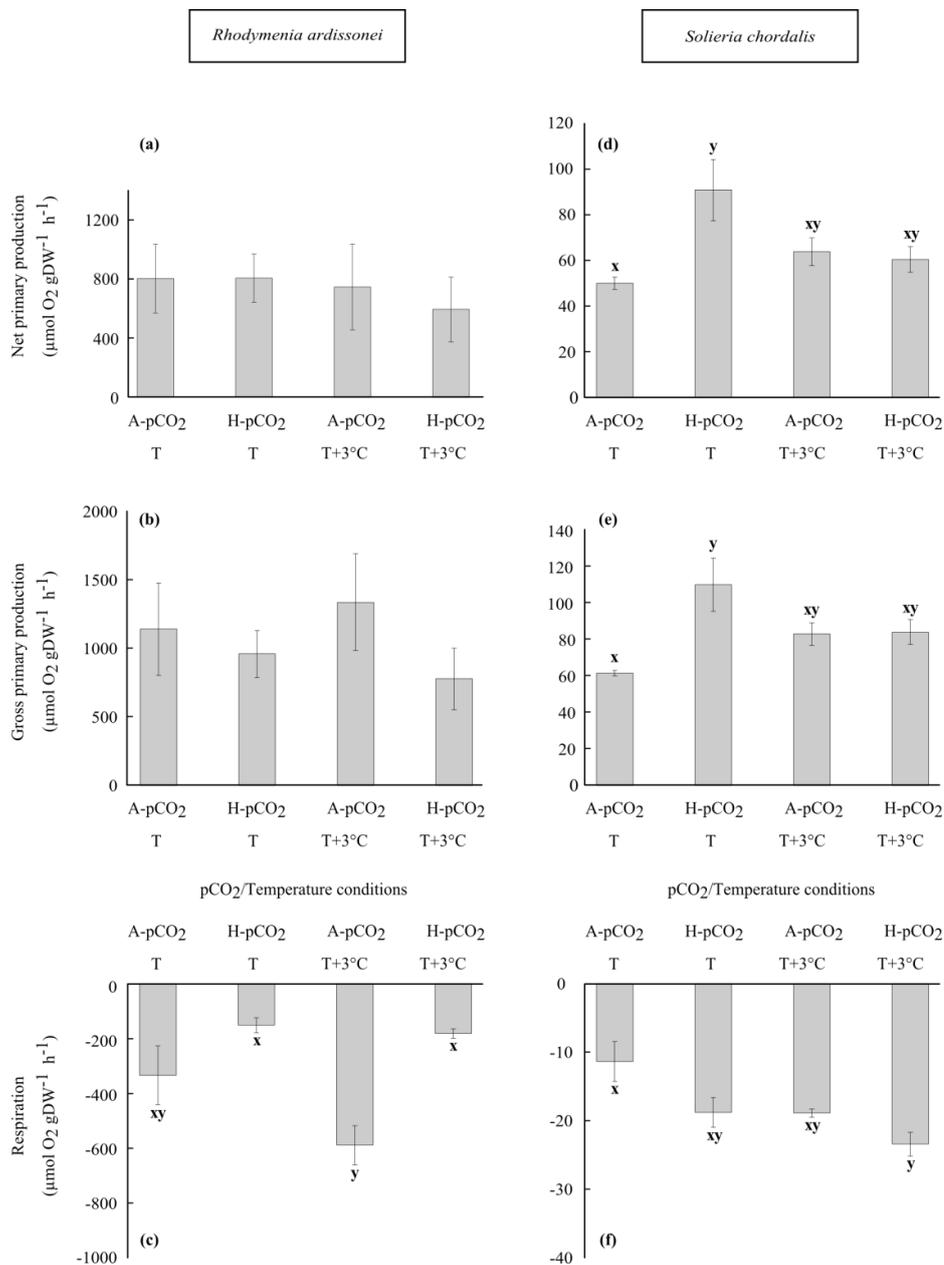
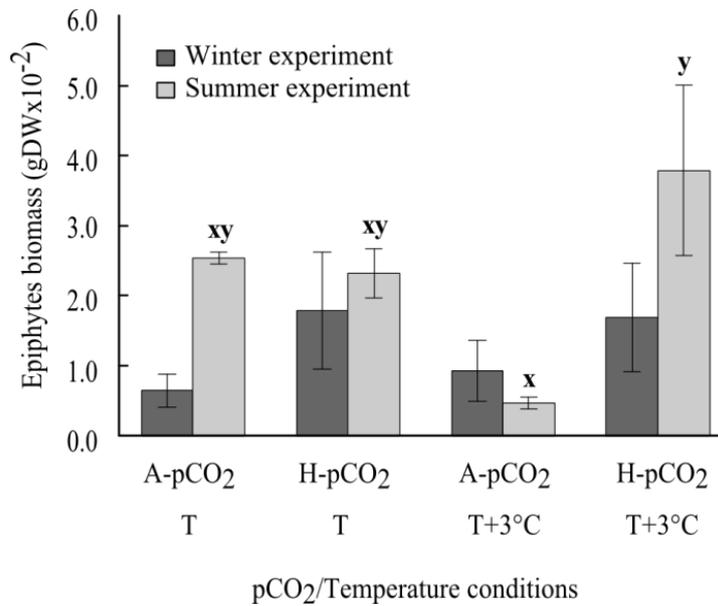


Fig. 3. Summer net and gross primary production and respiration rates (mean  $\pm$  SE,  $n = 5$ ) of the two main epiphytic fleshy algae *Rhodymenia ardissonae* (a to c) and *Solieria chordalis* (d to f), in the different pCO<sub>2</sub> (A-pCO<sub>2</sub> = Ambient pCO<sub>2</sub>; H-pCO<sub>2</sub> = High-pCO<sub>2</sub>) and temperature (T = Ambient temperature; T+3°C = High temperature) treatments. Letters indicate significant differences between the four treatments in summer (x, y) conditions (Tukey HSD test). Tukey tests were performed when a significant effect of temperature or pCO<sub>2</sub> was detected.

715



720 Fig. 4. Biomass of epiphytic fleshy algae (mean ± SE, n = 5) obtained in the different pCO<sub>2</sub> (A-pCO<sub>2</sub> = Ambient pCO<sub>2</sub>; H-pCO<sub>2</sub> = High-pCO<sub>2</sub>) and temperature (T = Ambient temperature; T+3°C = High temperature) treatments, after the three-month experiments in winter (dark gray) and summer (light gray) experiments. Letters indicate significant differences between the four treatments (x, y in the summer; Tukey HSD test). Tukey tests were performed when a significant effect of temperature or pCO<sub>2</sub> was detected using 2-way ANOVAs.

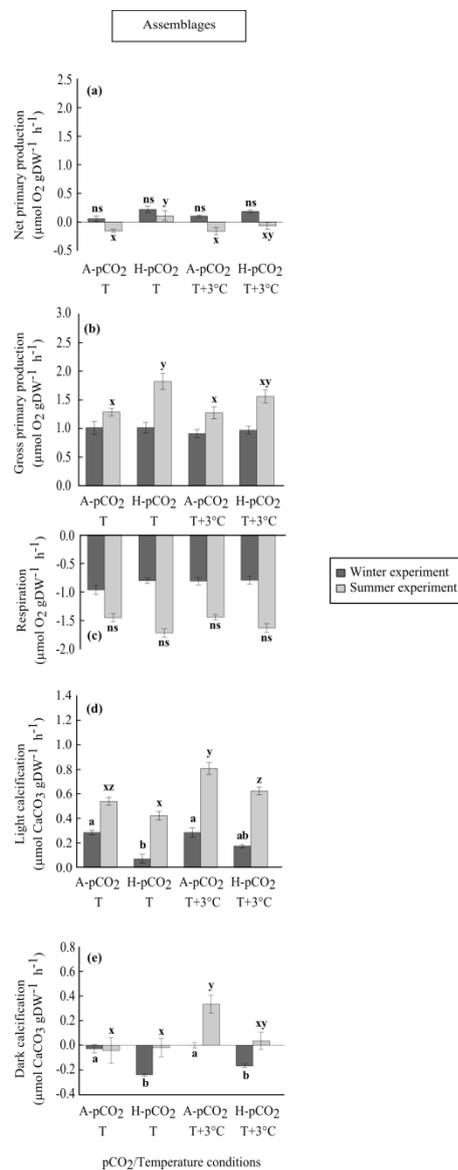


Fig. 5. Net and gross primary production (a and b, respectively), respiration (c) and light and dark calcification rates (d and e, respectively) rates (mean  $\pm$  SE,  $n = 5$ ) of assemblages in the different pCO<sub>2</sub> (A-pCO<sub>2</sub> = Ambient pCO<sub>2</sub>; H-pCO<sub>2</sub> = High-pCO<sub>2</sub>) and temperature (T = Ambient temperature; T+3°C = High temperature) treatments. The assemblages were maintained during three months in winter (dark gray) and summer conditions (light gray). Letters indicate significant differences between the four treatments in winter (a, b, c) and summer (x, y, z) conditions (Tukey HSD test). ns = not significant. Tukey tests were performed when a significant effect of temperature or pCO<sub>2</sub> was detected using 2-way ANOVAs.