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2 MONODEUTERATED METHANE:
3 AN ISOTOPIC PROBE TO MEASURE BIOLOGICAL METHANE METABOLISM
4 RATES AND TRACK CATABOLIC EXCHANGE REACTIONS

Jeffrey J. Marlow^{1*#}, Joshua A. Steele^{1^}, Wiebke Ziebis², Silvan Scheller¹,
David Case¹, Victoria J. Orphan¹

¹ Division of Geological and Planetary Sciences, California Institute of Technology, Pasadena, CA, 91125 USA

² Department of Biological Science, University of Southern California, Los Angeles, CA, 90089 USA

* Current address: Dept. of Organismic and Evolutionary Biology, Harvard University, Cambridge, MA 02138 USA

[^] Current address: Southern California Coastal Water Research Project, Costa Mesa, CA 92626 USA

Correspondence email: marlow@fas.harvard.edu

5

6 **Abstract**

7 Biological methane oxidation is a globally relevant process that mediates the flux of an
8 important greenhouse gas through both aerobic and anaerobic metabolic pathways. However,
9 measuring the rates of these metabolisms presents many obstacles, from logistical barriers to
10 regulatory hurdles and poor precision. Here we present a new approach for measuring rates of
11 microbial methane metabolism, using monodeuterated methane (CH_3D) as a metabolic substrate
12 and quantifying the change in the aqueous D/H ratio over time using off-axis integrated cavity
13 output spectroscopy. This method represents a non-toxic, comparatively rapid and straightforward
14 approach that is complementary to existing radio- (^{14}C) and stable (^{13}C) carbon isotopic methods;
15 by probing hydrogen atom dynamics, it offers an additional dimension through which to examine
16 the rates and pathways of methane metabolism. We provide direct comparisons between the CH_3D
17 procedure and the well-established $^{14}\text{CH}_4$ radiotracer method for several methanotrophic systems,
18 including type I and type II aerobic methanotroph cultures – for which the new approach is five
19 times more precise – and methane seep sediment and carbonate rocks under anoxic and oxic
20 incubation conditions. We also employ this method in a non-traditional experimental set-up,
21 investigating the role of pressure on methane oxidation rates in anoxic seep sediment. Results
22 revealed an 80% increase in methanotrophic rates at the equivalent of ~900 m water depth (40
23 MPa), revealing an important environmental parameter for methane metabolism and exhibiting the
24 flexibility of the newly described method.

25 The monodeuterated methane approach offers a procedurally straightforward, reliable
26 method that advances three specific aims. First, it allows users to directly compare methanotrophic
27 rates between different experimental treatments of the same inoculum. Second, by empirically
28 linking the CH_3D procedure with the well-established ^{14}C -radiocarbon approach, an absolute

29 scaling factor can be determined for new systems of interest. This “ground-truthing” strategy
30 enables “CH₃D only” experiments to yield rates of full methane oxidation; we demonstrate this
31 principle in the context of several methanotrophic systems. Finally, monodeuterated methane
32 facilitates a continued evaluation of C- and H-atom tracking through methanotrophic metabolisms,
33 with specific foci on enzyme reversibility and anabolic/catabolic branch points. The procedural
34 advantages, consistency, and novel research questions enabled by the monodeuterated methane
35 method should prove useful in a wide range of culture-based and environmental microbial systems
36 to further elucidate methane metabolism dynamics.

37

38 **1. Introduction**

39 Methane-consuming microbial processes represent an important component of
40 biogeochemical cycles in natural freshwater and marine environments, as well as in human-
41 impacted systems. In terrestrial soils, methane production in rice fields, anoxic wetlands, and
42 thawing permafrost supports methanotrophic communities (Holzapfel-Pschorn et al., 1985;
43 Mackelprang et al., 2011). In marine settings, an estimated 85 Tg of methane per year, derived
44 from biogenic and thermogenic sources, enters the subseafloor, the vast majority of which is
45 anaerobically consumed in anoxic sediments (Reeburgh, 2007). Much of what remains is taken up
46 in microoxic or oxic zones of the sediment or water column by aerobic methanotrophic
47 microorganisms (Valentine et al., 2001). In freshwater wetlands, approximately 200 Tg of methane
48 is generated per year, most of which is oxidized by hydroxyl radicals in the troposphere (Kirschke
49 et al., 2013). Methanotrophy is also of interest in a range of human-impacted contexts, including
50 wastewater treatment plants (Ho et al., 2013), landfills (Scheutz et al., 2009), and oil spills
51 (Crespo-Medina et al., 2014).

52 In addition to the climatic and economic implications of the methanotrophic process, its
53 biochemical details have stimulated many investigations. The anaerobic oxidation of methane
54 (AOM) has proven particularly enigmatic, often involving a mutualistic relationship between
55 anaerobic methanotrophic (ANME) archaea and sulfate reducing bacteria (SRB; Boetius et al.,
56 2000; McGlynn et al., 2015; Scheller et al., 2016; Wegener et al., 2015). A consensus on the
57 precise nature of the mutualism remains outstanding, but the net result of the process is typically
58 the stoichiometric oxidation of methane coupled with sulfate reduction (Knittel and Boetius, 2009).
59 Alternative electron acceptors including nitrate (Haroon et al., 2013), and nitrite (Ettwig et al.,
60 2010) have been demonstrated, while several studies have presented equivocal evidence for
61 methane oxidation coupled directly to iron or manganese reduction (Beal et al., 2009; Nauhaus et
62 al., 2005; Sivan et al., 2014).

63 Methane is oxidized aerobically by members of the classes *Gammaproteobacteria* (Type I)
64 and *Alphaproteobacteria* (Type II); verrucomicrobial representatives were more recently found to
65 perform aerobic methanotrophy under extremely acidic conditions (Dunfield et al., 2007; Op den
66 Camp et al., 2009). Methane monooxygenase converts methane to methanol, which is further
67 oxidized to formaldehyde; assimilatory pathways branching at this point can incorporate carbon
68 into central metabolism through the ribulose monophosphate (RuMP) cycle (Type I
69 methanotrophs) or the serine cycle (Type II). Remaining formaldehyde can undergo two additional
70 oxidation reactions, being converted first to formate and ultimately to carbon dioxide (Hakemian
71 and Rosenzweig, 2007).

72 Methanotrophy is both a biogeochemically relevant force that modulates global climate and
73 a poorly understood biochemical process; given this dual role, there is substantial interest in
74 measuring the rate of the process and understanding elemental flows through metabolic pathways.

75 The oxidation of methane in environmental samples has traditionally been studied using a handful
76 of techniques. Numerical models incorporating environmental sediment profiles of sulfate and
77 methane concentrations can be used to back-calculate methane consumption rates (Jørgensen et al.,
78 2001). Methane labeled with ^{13}C can be used to probe longer-term rates in controlled conditions
79 (Moran et al., 2008), but the presence of natural ^{13}C in marine dissolved inorganic carbon pools
80 requires long incubations as well as accurate measurements of concentrations and isotopic
81 compositions of reactants and products (Pack et al., 2011). Gas chromatography quantification of
82 dissolved (Girguis et al., 2003) or headspace (Carini et al., 2003) methane concentrations has also
83 been demonstrated as a rate measurement tool, though low concentrations can hamper
84 reproducibility and exacerbate background contamination issues, particularly in field-based
85 settings (Magen et al., 2014). Perhaps the most sensitive approach uses radiolabeled $^{14}\text{CH}_4$ to track
86 the oxidation of methane-associated carbon to inorganic carbon species (Alperin and Reeburgh,
87 1985; Treude et al., 2003). Labeling with tritiated methane was introduced for water column
88 aerobic methane oxidation measurements due to its higher specific activity and the procedural
89 advantages of working with a water-phase product rather than gaseous products (Bussmann et al.,
90 2015; Valentine et al., 2001). Logistical challenges and health and safety regulations led Pack et al.
91 (2011) to develop an accelerator mass spectrometry detection method that requires $10^3\text{-}10^5$ less
92 radiolabel than previous ^{14}C and ^3H approaches, though the analytical procedure remains labor
93 intensive.

94 Despite the range of methods available, measurement of microbial methane utilization rates
95 remains cumbersome, and the demonstration of a precise, safe, and easily enacted approach would
96 be a welcome contribution for a diverse field of researchers. Nearly all of the aforementioned
97 approaches are carbon-based; a hydrogen-based tracer offers an additional dimension to

98 investigations of methane biochemical dynamics. Here we introduce a novel method for
99 biologically mediated methanotrophy rate measurement that utilizes monodeuterated methane
100 (CH_3D) as a substrate and measures the D/H ratio of the aqueous solution.

101 We demonstrate, through methanotrophic cell cultures and microcosm incubations of
102 seafloor sediment and carbonate rock fragments, that methane activation rates derived from
103 aqueous D/H values are consistently proportional to ^{14}C -based methane oxidation rate
104 measurements for the laboratory treatments tested in this study. The resulting ratios, when viewed
105 in the context of partial (methane-associated hydrogen exchange) versus complete methane
106 oxidation (methane oxidation to CO_2), represent a new tool with which to examine the reversibility
107 and catabolic / anabolic partitioning of methanotrophic metabolisms. As a proof of concept, we
108 apply the monodeuterated methane approach to pressurized methane seep sediment incubations in
109 order to test the role of an understudied environmental variable in methanotrophic rates under non-
110 traditional empirical conditions. As a rate measurement protocol, this approach offers several
111 advantages over current techniques: it does not require the logistical, safety, and administrative
112 hurdles associated with radiotracers such as $^{14}\text{CH}_4$ and $^3\text{H}-\text{CH}_4$, it is less susceptible to analyte loss
113 than methane headspace measurements, and compares favorably in terms of equipment cost and
114 portability.

115

116 **2. Methods**

117 2.1. Experimental Set-Up

118 To demonstrate the precision and reproducibility of the monodeuterated methane approach,
119 it was tested alongside the well-established $^{14}\text{CH}_4$ radiotracer protocol. The use of $^{14}\text{CH}_4$ is an
120 accepted standard procedure in studies of methane consumption quantification (e.g., Knittel and

121 Boetius, 2009; Ruff et al., 2016; Segarra et al., 2013) and has been experimentally cross-referenced
122 with methane concentration measurements (Treude et al., 2003) and other approaches including
123 tritiated methane techniques (Mau et al., 2013; Pack et al., 2011). Both techniques were applied to
124 a) aerobic methanotrophic cultures of *Methylosinus trichosporium* OB3b (kindly supplied by
125 Marina Kalyuzhnaya and Mary Lidstrom) and *Methyloprofundus sedimenti* (isolated from a deep
126 sea whale fall; Tavormina et al., 2015) b) oxic incubations of methane seep sediment and
127 carbonate rocks, and c) anoxic incubations of methane seep sediment and carbonate rocks. In
128 addition, the monodeuterated methane protocol was employed in a pressure-based experiment to
129 demonstrate the technique's adaptability to distinct empirical set-ups and examine the relative
130 effect of heightened, environmentally relevant pressure on methane consumption rates in anoxic
131 seep sediment samples. Monodeuterated methane for all samples was 98% pure CH₃D obtained
132 from Sigma-Aldrich (\$247 / L). For a representation of all experiments conducted in this study, see
133 Table 1.

134 *2.1.1. Experiments with Aerobic Methanotroph Cultures*

135 Cultures of *Methylosinus trichosporium* strain OB3b (Whittenbury et al., 1970) were
136 grown using Nitrate Mineral Salts (NMS) medium at 30 °C. The newly characterized
137 *Methyloprofundus sedimenti* strain WF1 was grown in a modified NMS medium at 25 °C
138 (Tavormina et al., 2015). In both cases, shaking cultures were grown up from stock in sealed 25
139 mL test tubes that contained 5 mL media and 50:50 air:methane by volume. After several
140 successful transfers (as determined by an increase in optical density, data not shown), experiments
141 were initiated by passaging 0.94 mL of exponential phase inoculum into 8.5 mL media, for each of
142 ten different experimental conditions, each prepared in triplicate (see Table S1). Due to the

143 destructive nature of the $^{14}\text{CH}_4$ method, three of these triplicate sets were used to measure methane
144 oxidation at three distinct time points.

145 Samples for D/H analysis were taken at seven time points throughout 140-hour (*M.*
146 *trichosporium*) and 476-hour (*M. sedimenti*) experiments. Sampling points were most concentrated
147 around anticipated exponential growth phases as determined by optical density (600 nm) profiles
148 of earlier rounds of culture transfers. Samples for radiolabel processing were taken at 46, 102, and
149 166.5 hours for *M. trichosporium* cultures and 102, 166.5, and 432 hours for the slower-growing
150 *M. sedimenti* cultures. Killed, cell-free, oxygen-free, and CH_3D -free controls were all assessed
151 (Table S1).

152 *2.1.2. Experiments with Environmental Samples: Methane Seep Sediments and Carbonates*

153 Samples recovered from the Hydrate Ridge methane seep system were used to
154 comparatively examine the novel monodeuterated methane (CH_3D) approach alongside the $^{14}\text{CH}_4$
155 protocol with environmental samples. Hydrate Ridge, Oregon, is located along a convergent
156 tectonic margin and is well established as a site of methane seepage and sediment-based AOM
157 (e.g., Suess et al., 1999; Treude et al., 2003; Tryon et al., 2002). Methane concentrations within the
158 most active seep sediments reach several mM, and have been measured and modeled at values up
159 to 70 mM (Boetius and Suess, 2004) and 50 mM (Tryon et al., 2002) respectively.

160 Samples were collected with the Deep Submergence Vehicle (DSV) *Alvin* during *Atlantis*
161 leg AT-16-68 in September 2010 and the Remotely Operated Vehicle (ROV) *Jason II* during
162 *Atlantis* leg AT-18-10 in September 2011; materials used for methanotrophic rate experiments are
163 specified in Table 1. The “active” designation in our sample descriptions refers to sites where
164 methane seepage was manifested by seafloor ecosystems known to be fueled by subsurface
165 methane (e.g. clam beds and microbial mats) or methane bubble ebullition. The term “low activity”

166 references sampling sites that did not exhibit any clear signs of contemporary methane seepage or
167 chemosynthetic communities, though a small amount of methane supply and methanotrophic
168 potential cannot be ruled out as subsurface advective flow can shift with time (Gieskes et al., 2005;
169 Marlow et al., 2014; Tryon et al., 2002). Sample types are abbreviated by the A.Sed (active
170 sediment), A.Carb (active carbonate), L.Sed (low-activity sediment), and L.Carb (low-activity
171 carbonate) designations. Seven samples were analyzed to examine a range of physical substrate
172 type (sediment vs. carbonate rock) and seepage environments (active and low-activity): A.Sed-
173 5128, A.Carb-5305, A.Carb-5152, L.Sed-5043, L.Carb-5028, and sterilized control aliquots of
174 A.Sed-5128 and A.Carb-5305. Carbonate samples include both porous materials with macroscale
175 vugs and pore spaces, as well as massive lithologies with more homogenous structure.

176 Shipboard, push cores and bottom water-submerged carbonates were immediately
177 transferred to a 4 °C walk-in cold room and processed within several hours. To prepare material for
178 future experimentation, sediment and carbonate rocks were stored in anoxic, Ar-flushed, gas-tight
179 mylar bags (Impak Corp., Los Angeles, USA) at 4 °C until use several months later. In advance of
180 experimental set-up, carbonate samples and homogenized sediment from the 0-12 cm push core
181 horizon were prepared under anoxic conditions using 0.22 µm-filtered, anoxic N₂-sparged Hydrate
182 Ridge bottom water (at a 1:2 sediment/carbonate:bottom water ratio by volume). Samples were
183 maintained under a 2x10⁵ Pa CH₄ headspace for one month in order to resuscitate activity; the
184 corresponding dissolved concentration (3.7 mM) is consistent with environmental methane
185 concentrations at Hydrate Ridge (Boetius and Suess, 2004).

186 To set up the experimental incubations, 10 mL physical substrate (compressed sediment or
187 carbonate rock) and 20 mL filtered Hydrate Ridge bottom water were placed in 60-mL glass
188 bottles (SVG-50 gaschro vials, Nichiden Riku Glass Co, Kobe, Japan). In all experiments

189 involving carbonates, interior portions (> 5 cm from the rock surface) were used in order to ensure
190 that properties exhibited were representative of bulk rock material and not a reflection of surface-
191 based adherent cells or entrained material. Carbonate rock samples were fragmented in order to fit
192 through the 28-mm diameter bottle opening; pieces were kept as large as possible to minimize the
193 increase in surface area-to-volume ratio and maintain conditions as representative of the *in situ*
194 environment as possible. All bottles were sealed with butyl stoppers; following several minutes of
195 flushing with N₂ (g), the headspace was replaced with methane, and an additional 30 mL of gas,
196 whose composition varied depending on the experiment, was injected into the 30 mL headspace to
197 generate an absolute pressure of approximately 2×10^5 Pa. Anoxic incubation headspace was 100%
198 methane; oxic incubation headspace was 30 mL methane, 20 mL N₂, and 10 mL O₂. All incubation
199 set-up prior to gas flushing and headspace injection took place in an anaerobic chamber. Triplicate
200 samples, including killed controls, were prepared for all sample types. Measurements were taken
201 for both D/H and ¹⁴C analysis at 46 and 96 hours for oxic incubations, and 72 and 192 hours for
202 anoxic incubations. Anoxic active methane seep sediment (A.Sed-5128) incubations were used for
203 nuclear magnetic resonance (NMR) analysis of the remaining methane (set up in triplicate, with 60
204 mL CH₃D initial headspace) as well as empirical resolution studies sampled between days 20-22.

205 *2.1.3. Experiments with Environmental Samples in Pressure Vessels*

206 In order to probe the effect of pressure on anaerobic methanotrophic rates, a set of
207 experiments was established, using the monodeuterated methane technique to determine relative
208 rate differences. Active sediment from Hydrate Ridge (A.Sed-3450) was collected from a water
209 depth of 850 m and an ambient temperature of 4 °C, processed shipboard, and prepared for
210 experimentation as described above. To set up the incubations, eight 100 mL mylar bags were
211 prepared with the components shown in Table S2 using homogenized sediment from the 0-12 cm

212 horizon. 500 μM glycine or 500 μM ammonium were added in order to evaluate relative rate
213 differences associated with organic and inorganic sources of nitrogen. Identical sets of four
214 compositionally distinct samples – including killed controls – were established such that each
215 treatment could be subjected to low pressure (0.1 MPa, i.e., atmospheric pressure) and high
216 pressure (9.0 MPa, equivalent to ~900 m water depth). Prior to gas addition, each bag was flushed
217 for 5 min. with Ar.

218 The use of flexible mylar bags is essential for the application of external pressure, yet it
219 presents obstacles for “traditional” methanotrophic rate measurement protocols such as the $^{14}\text{CH}_4$
220 method. In particular, the processing of post-incubation headspace is optimized for stoppered
221 bottles, and accessing the gas phase from mylar bags in a quantitative fashion is challenging.
222 Measurement of radiolabeled dissolved inorganic carbon requires that all incubation material be
223 transferred to an Erlenmeyer flask equipped with a scintillation vial; sediment grains are commonly
224 trapped in the seals of mylar bags, complicating this transfer. For these reasons, monodeuterated
225 methane addition and subsequent aqueous measurement offered a useful tool for this challenging
226 experimental set-up.

227 Once the incubations were prepared, they were transported to the laboratory of Dr. William
228 Berelson at the University of Southern California and placed in a walk-in cold room (4 °C).
229 Incubations for pressurized treatment were inserted into a stainless steel, custom-built pressure
230 chamber with 3-cm thick walls and pressure valves rated to 40 MPa, and hydraulic fluid was
231 pumped into the sealed chamber using a Star Hydraulics P1A-250 hand pump. The pressure was
232 maintained at 9.0 MPa during the course of the 38-day experiment, with daily adjustments to
233 account for thermal compression effects. At the conclusion of the experiment, mylar bags were
234 removed from the chamber, checked for leaks (none were observed, as the bags were still inflated,

235 the seal was still gas-tight, and no hydraulic fluid was detected in the interior of the mylar bags)
236 and sampled for D/H ratio measurement.

237 2.2. Analytical Procedures

238 *2.2.1. Rate Measurements Derived from CH₃D Addition*

239 At designated sampling times, 1 mL of medium / water was collected from cultures or
240 incubations in an anaerobic chamber with a sterile syringe. A constant volume was maintained by
241 adding 1 mL of sterile media immediately after sampling; this media was previously equilibrated
242 with gaseous headspace specific to each experiment. Sampled liquid was then pushed through a
243 0.22 µm Durapore filter (EMD Millipore, Temecula, CA) and into a 1-mL GC vial. A LGR DLT-
244 100 liquid water isotope analyzer (LWIA, Los Gatos Research, Mountain View, CA) was used to
245 determine the D/H ratio of each sample. The LWIA uses off-axis integrated-cavity output
246 spectroscopy to measure isotopically specific absorption patterns and determine simultaneous D/H
247 and ¹⁸O/¹⁶O ratios with high precision and robust mechanics (Lis et al., 2008). Such instruments
248 have been used for a range of studies, including hydrological analysis (Robson and Webb, 2016),
249 mine waste management (Huang et al., 2015), and microbial metabolism (Dawson et al., 2015).

250 In this study, an injection volume of 700 nL at 1000 nL/s was used, with four intra-
251 injection flush strokes and a flush time of 60 s between injections. Four rounds of ten injections per
252 sample were performed in order to avoid memory effects; only the latter five injections were used
253 in subsequent calculations. Sample runs were limited to ~250 injections in order to minimize salt
254 precipitation, and each analysis included an appropriate blank (i.e., autoclaved media for the
255 cultures, or filter sterilized bottom water used during incubation set-up in the case of sediment and
256 carbonate incubations) and two standards of known isotopic ratios (Deep Blue: δD = 0.5‰, and
257 CIT: δD = -73.4‰). Data was removed if instrumental temperature or pressure parameters were

258 observed to fall outside of optimal instrument specifications (0.76% of all analyses), corresponding
259 to an internal temperature change of more than 0.3 °C per hour or rising pressure within the
260 measurement cell during the analysis.

261 To calculate methane consumption rates, D/H ratios from the LWIA were first normalized
262 to the Vienna standard mean ocean water (VSMOW) scale using a two-point calibration from the
263 water standards and a linear interpolation (e.g., Dawson et al., 2015). To minimize the effects of
264 instrumental drift, standards were re-measured between rounds of sample analysis (maximum of
265 40 injections) and new scaling factors were implemented. The number of total hydrogen atoms (H
266 and D) present at the initial time point was calculated using the experiment's overall water volume,
267 as in equation 1:

$$\frac{vol(L)}{1} * \frac{55.5 \text{ mol water}}{L} * \frac{6.022 \times 10^{23} \text{ molecules water}}{\text{mol water}} * \frac{2 \text{ hydrogen atoms}}{\text{molecule water}} = \text{hyd. atoms in inc.}_{T_1}$$

268 The number of D atoms newly present in the experiment's aqueous phase between time points T₁
269 and T₂ was determined using the normalized D/H values, averaging across the latter five injections
270 of the four distinct injection rounds; see equations 2.1-2.3:

$$\left[\left(\frac{D}{H} \right)_{T_2} * (\text{hyd. atoms in inc.})_{T_2} \right] - \left[\left(\frac{D}{H} \right)_{T_1} * (\text{hyd. atoms in inc.})_{T_1} \right] = \text{new D atoms in inc.}$$

$$(\text{hyd. atoms in inc.})_{T_2} \approx (\text{hyd. atoms in inc.})_{T_1}$$

$$\left[\left(\frac{D}{H} \right)_{T_2} - \left(\frac{D}{H} \right)_{T_1} \right] * (\text{hyd. atoms in inc.})_{T_1} = \text{new D atoms in inc.} = D_{\text{new}}$$

271 D_{new} was multiplied by four given the 1:3 D:H stoichiometry of the CH₃D substrate to derive the
272 maximum number of methane molecules consumed catabolically through initial C-X bond
273 activation (equation 3).

$$D_{\text{new}} * 4 = \text{maximum number of methane molecules consumed} = C$$

274 The scaling factor of four was used in the context of methane activation – the initial mobilization
275 through conversion to a methyl group – to calculate the maximum number of methane molecules
276 that could be consumed but not necessarily fully oxidized. This represents an end-member case
277 that may not be appropriate for all subsequent processing as hydrogen/deuterium atoms are
278 incorporated into biomass or exchanged. Caveats and potential interpretations of the absolute
279 numbers that result from these calculations are discussed below, but we stress that with consistent
280 implementation of scaling factors from comparisons between monodeuterated and radiolabel
281 methods, rates derived from C and downstream parameters are valid and useful.

282 C was corrected based on the fraction of incubation methane headspace composed of
283 CH₃D, yielding C_{corr} , as shown in equation 4:

$$\frac{C}{\text{fraction of methane headspace that is CH}_3\text{D}} = C_{corr}$$

284 By dividing C_{corr} by the incubation time and volume, a maximum rate of methane consumption is
285 determined (equation 5.1-5.2):

$$C_{corr} * \frac{\text{mol}}{6.022 \times 10^{23} \text{ molecules}} * \frac{10^9 \text{ nmol}}{\text{mol}} * \frac{1}{\text{incubation time (d)}} * \frac{1}{\text{incubation vol (cm}^3\text{)}} = R_{CH_3D}$$

$$R_{CH_3D} = \text{Maximum rate of methane consumption in } \frac{\text{nmol}}{\text{cm}^3 \text{ d}}$$

286 2.2.2. Rate Measurements Derived from $^{14}\text{CH}_4$ Addition

287 Methane oxidation rates using radiolabeled methane substrate were measured as described
288 in detail by Treude et al. (2005) and Treude and Ziebis (2010). Radiolabeled methane ($^{14}\text{CH}_4$
289 dissolved in seawater, corresponding to an activity of 13 kBq for culture experiments and 52 kBq
290 in sediment and carbonate samples) was injected into each sample container, and samples were
291 incubated at the appropriate temperatures for the designated amount of time (see above). To stop
292 microbial activity and begin analysis, 2.5 ml of 2.5% NaOH was injected. Sample headspace was

293 flowed through a Cu²⁺ oxide-filled 850 °C quartz tube furnace, combusting unreacted ¹⁴CH₄ to
294 ¹⁴CO₂. This ¹⁴CO₂ was collected in two scintillation vials (23 ml volume) pre-filled with 7 ml
295 phenylethylamine and 1 ml 2-methoxyethanol, to which 10 ml of scintillation cocktail (Ultima
296 Gold XR, PerkinElmer) was added. After a 24-hour incubation period, radioactivity from ¹⁴CO₂
297 was measured by scintillation counting (Beckman Coulter LS 6500 Multi-Purpose Scintillation
298 Counter, 10 minute analysis per sample).

299 Labeled ¹⁴C-inorganic carbon produced during the incubation was quantified as follows.
300 The entire volume of each incubation sample was transferred into a 250-ml Erlenmeyer flask along
301 with 1 drop of antifoam and 5 ml of 6M HCl. The flask was immediately stoppered and sealed
302 with two clamps and parafilm wrapping to prevent gas escape, and placed on a shaking table (60
303 rpm, room temperature, 24 hours). To collect ¹⁴CO₂ generated by the acidification process, a 7-ml
304 scintillation vial was pre-filled with 1 ml of 2.5% NaOH and 1 ml of phenylethylamine and
305 suspended from the rubber stopper inside the flask. After the shaking / acidification step, 5 ml of
306 scintillation cocktail was added, and the vial was measured by scintillation counting after 24 hours.
307 This method has been demonstrated to recover 98% of ¹⁴CO₂ on average (Treude et al., 2003).

308 Finally, sterilized control samples (#10, see Table S1) were set aside after ¹⁴CH₄ addition
309 for gas chromatography to determine the initial concentration of methane gas. 400 µl of headspace
310 was injected into a gas chromatograph (Shimadzu GC-2014), equipped with a packed stainless
311 steel Supelco Custom Column (50/50 mixture, 80/100 Porapak N support, 80/100 Porapak Q
312 column, 6 ft x 1/8 in) and a flame ionization detector. The carrier gas was helium at a flow rate of
313 30 ml min⁻¹, and the column temperature was 60 °C. Results were scaled based on comparison
314 with standards of known methane concentrations (10 and 100 ppm; Matheson Tri-Gas, Twinsburg,
315 OH). The rate of methane oxidation was determined by equation 6:

316

$$\text{Methane Oxidation} = \frac{{}^{14}\text{CO}_2 \cdot \text{CH}_4}{({}^{14}\text{CH}_4 + {}^{14}\text{CO}_2) \cdot v \cdot t}$$

317 in which ${}^{14}\text{CH}_4$ is the combusted unreacted radiolabeled methane, ${}^{14}\text{CO}_2$ represents the quantity of
318 acidified oxidation product, CH_4 signifies the initial quantity of methane in the experiment, v is the
319 volume of sediment or carbonate rock, and t is the time over which the incubation was active.

320 *2.2.3. Isotopic Analysis of Methane in the Headspace*

321 The methane headspace was analyzed via ${}^1\text{H-NMR}$ spectroscopy using a Varian 400 MHz
322 Spectrometer with a broadband auto-tune OneProbe. 300 μl of headspace was passed through
323 CDCl_3 with a fine needle to absorb the methane. ${}^1\text{H-NMR}$ spectra were acquired at 298 K without
324 spinning, using a repetition rate of 10 s to ensure reliable quantification. The spectra were
325 simulated with the iNMR 4.1.7 software for the determination of the fractional abundances of the
326 ${}^{12}\text{CH}_4$, ${}^{12}\text{CH}_3\text{D}$, ${}^{13}\text{CH}_4$ and ${}^{13}\text{CH}_3\text{D}$ isotopologs.

327

328 **3. Results and Discussion**

329 3.1. Comparison of CH_3D and ${}^{14}\text{CH}_4$ Rate Measurements in Aerobic Methanotroph Cultures

330 D/H ratios were acquired and corresponding C_{corr} values were calculated at eight points
331 during the *M. trichosporium* growth curve and seven points of the *M. sedimenti* growth curve.
332 Three measurements of ${}^{14}\text{C}$ distributions were acquired for each strain, targeting exponential and
333 stationary phases (Fig. 1). The Type II alphaproteobacterial methanotroph *M. trichosporium*
334 exhibited methane consumption rates more than an order of magnitude greater than those of *M.*
335 *sedimenti* (gammaproteobacterial Type I methanotroph), yet the scaling factor relating the CH_3D -
336 and ${}^{14}\text{CH}_4$ -derived rates was remarkably consistent in both cases. Scaling factors were calculated
337 for both exponential growth and stationary phase, using data points from both CH_3D and ${}^{14}\text{CH}_4$
338 experiments. For example, the *M. trichosporium* rate value calculated from CH_3D experimental

339 treatment point (47.5 hours, 4.16×10^4 nmol methane consumed) was compared with the rate
340 determined from $^{14}\text{CH}_4$ experimental treatment point (47.5 hours, 2.77×10^4 nmol methane
341 consumed), yielding a scaling factor of 1.5 for exponential phase growth. Similarly, data from (140
342 h, 5.27×10^4 nmol, CH_3D) and (166.5 h, 4.24×10^4 nmol, $^{14}\text{CH}_4$) were used for *M. trichosporium*'s
343 stationary phase scaling factor. Equivalent values were determined for *M. sedimenti* using the
344 following data points: (140 h, 7.07×10^3 nmol, CH_3D) and (102 h, 3.43×10^3 nmol, $^{14}\text{CH}_4$) for the
345 exponential growth phase, and (476 h, 7.53×10^3 nmol, CH_3D) and (432 h, 4.30×10^3 nmol,
346 $^{14}\text{CH}_4$) for stationary phase.

347 In this way, the ratio of methane consumption rates derived from the CH_3D method (using
348 equations 1-5) and the $^{14}\text{CH}_4$ method (using equation 6) can be compared. This value is hereafter
349 referred to as the “D: ^{14}C tracer ratio”. This ratio can be used to evaluate the consistency of the
350 monodeuterated methane method compared with the well-established $^{14}\text{CH}_4$ approach, and as an
351 investigatory tool in catabolic / anabolic processing of methane (see “Understanding the D: ^{14}C
352 Tracer Ratio”, below).

353 D: ^{14}C tracer ratio values were calculated for aerobic methanotroph cultures using the data
354 specified above and are shown in Table 2; their consistency is a promising indicator of the utility
355 of the monodeuterated methane approach for ground-truthed rate measurements. By dividing rates
356 derived from D/H values by 1.5, a reliable estimate of full-oxidation methanotrophy – that is, the
357 complete biological oxidation of methane to carbon dioxide – can be assessed.

358 3.2. Comparison of CH_3D and $^{14}\text{CH}_4$ Rate Measurements in Environmental Methane Seep Samples

359 Methane consumption rates under oxic microcosm incubation conditions, derived from
360 both CH_3D and $^{14}\text{CH}_4$ measurements, are provided for all five sample types (active sediment, low-
361 activity sediment, active porous carbonate, active massive carbonate, and low-activity massive

362 carbonate) in Fig. 2a. The corresponding values for anoxic conditions are shown in Fig. 2b. Values
363 were calculated from data collected after 4 days of incubation for oxic samples and after 8 days of
364 incubation for anoxic samples.

365 The D:¹⁴C tracer ratio for the oxic incubations was 1.66 +/- 0.02 SE and 1.99 +/- 0.04 SE
366 for anoxic conditions (Table 2). These relatively consistent values across physical substrate type
367 (sediment and carbonates of varying lithology) and collection site activity level (active and low-
368 activity) suggest an underlying metabolic basis of these D:¹⁴C tracer ratio that is unperturbed by
369 physicochemical factors or relative activity levels.

370 To determine the minimum number of activated CH₃D molecules needed for analytical
371 detection, we assessed the length of time required to measure a differentiable D/H ratio.
372 Measurements were acquired at multiple time points between days 20 and 22 of a triplicate set of
373 A.Sed-5128 incubations. A resolvable signal of an enhanced D/H ratio was defined as data points
374 with non-overlapping confidence intervals, representing a 95% statistical probability that D/H
375 ratios were increased. Such differentiation seen at the 20-hour sampling time for two replicates and
376 the 26-hour sampling time for the other one (Fig. S1). Using the rate determined by the first 20
377 days as a baseline, this translates to a resolution of 4.5-6.2 μmol of fully oxidized methane based
378 on the D:¹⁴C tracer ratio of 2.05 (Table 2).

379 3.3. Understanding the D:¹⁴C Tracer Ratio

380 The CH₃D and ¹⁴CH₄ approaches quantify distinct aspects of methanotrophy; that is,
381 methane activation or complete conversion to CO₂, respectively. The ¹⁴CH₄ technique quantifies
382 the amount of ¹⁴C – initially supplied as methane – that is fully oxidized and persists as soluble
383 species (HCO₃⁻) or acid-labile precipitation products (CaCO₃). The CH₃D protocol, on the other
384 hand, reports the extent to which methane-derived hydrogen atoms are detected in water. Because

methane is an inert molecule, abiotic exchange between methane- and water-associated hydrogen atoms is not expected. Indeed, D/H ratios in killed control experiments remained stable (e.g., exhibiting a value of $1.40 \times 10^{-4} \pm 3.1 \times 10^{-8}$ SE at T_{0d} and $1.40 \times 10^{-4} \pm 2.9 \times 10^{-8}$ SE at T_{140d} during experimentation with *M. trichosporium*, data incorporated into Fig. 1a). The activation of methane thereby indicates enzymatic functionalization, but the ultimate fate of each hydrogen atom during methane oxidation is unclear.

The flow of methane-derived hydrogen atoms through anaerobic and aerobic methanotrophic metabolisms was examined in an attempt to predictively evaluate the consequence of monodeuterated methane reactions. Previously published reports were used to compile Figure 3 (Hallam et al., 2004; Thauer, 2011; Vorholt and Thauer, 1997) and Figure 4 (Lieberman and Rosenzweig, 2004), which trace anaerobic and aerobic methane metabolisms, respectively, with a specific focus on hydrogen atoms. In this context, our observations of relatively consistent but distinct D: ^{14}C tracer ratios for anaerobic and aerobic methanotrophy (Table 2) likely reflect different aspects of the two metabolic pathways. In AOM, metabolite backflux (Holler et al., 2011) may increase the D/H ratio; in aerobic methanotrophy, biomass growth represents a substantial carbon and hydrogen shunt.

3.3.1. The D: ^{14}C Tracer Ratio in Anaerobic Methanotrophy

AOM is depicted in Fig. 3 via the reverse methanogenesis pathway, which is believed to be enacted by ANME based on genetic (Hallam et al., 2004) and proteomic (Marlow et al., 2016) data. In this metabolic process, methyl-coenzyme M reductase (Mcr) activates methane and generates methyl-CoM. A tetrahydromethanopterin molecule supplants CoM, and subsequent carbon oxidation steps release hydrogen atoms into the medium. Ultimately, the number of methane-derived hydrogen atoms that enter water-exchangeable products determines the

408 physiological interpretation of aqueous D/H ratios. For example, if just one methane-derived
409 hydrogen enters an intermediate and is freely exchangeable with water, then observed water-based
410 deuterium must be multiplied by four (to account for methane's hydrogen-carbon stoichiometry;
411 see equation 3) and a primary isotope effect as high as 2.44 (*M. marburgensis*' Mcr's C-H vs. C-D
412 bond-breaking preference, Scheller et al., 2013) to arrive at the actual quantity of activated
413 methane molecules. In this context, the experimental D:¹⁴C tracer ratio values may provide useful
414 insight. A D:¹⁴C tracer ratio of 2 for the reverse methanogenesis pathway suggests that, for every
415 methane molecule that is fully oxidized to CO₂, two hydrogen atoms enter water-exchangeable
416 intermediates.

417 However, the back reaction of enzymatic processes (Scheller et al., 2010) could lead to
418 heightened D/H ratios in the absence of full carbon oxidation. For example, upon the activation of
419 methane by Mcr, HS-CoB and CH₃-S-CoM form, with the thiol hydrogen exchanging with water-
420 bound hydrogen. If the S-bound hydrogen were deuterium, then the methane re-formed by Mcr
421 back reaction would be CH₄ and the aqueous deuterium would cause a heightened D/H ratio
422 despite a lack of net methane oxidation (Fig. 3). We analyzed the remaining headspace of seep
423 sediment incubations for the formation of CH₄ from CH₃D via ¹H-NMR spectroscopy. Over the
424 course of 58 days in triplicate A.Sed-5128 incubations prepared with exclusively CH₃D headspace,
425 CH₄ in the headspace increased from 0.33% +/- 0.02% SE to 4.48% +/- 0.27% SE. If this
426 demonstrated reversibility only reflects the back reaction of Mcr, then the CH₄ increase must be
427 multiplied by four to reflect the actual percentage of headspace methane that was re-formed by
428 Mcr; if the reversibility reflects back reaction of the entire pathway, then no scaling factor is
429 needed. Thus, the range of potential methane headspace percentage accounted for by methane
430 reformed from initial CH₃D is between 4.15 – 16.6%. (For clarity, these calculations neglect

431 isotope effects and activity by methanogens, factors that can be clarified through further
432 experimentation.) Reversibility can be evaluated in future stable isotope work by a) including a
433 ^{13}C -dissolved inorganic carbon (DIC) signal in the water and measuring $^{13}\text{CH}_4$, and/or b) utilizing
434 multiply deuterated methane as initial headspace and quantifying all possible isotopologs.
435 Nonetheless, even the upper bound of partially and reversibly oxidized CH_3D suggests that the
436 majority of the D/H signal is attributable to reactions indicative of net methane consumption, if not
437 complete oxidation.

438 *3.3.2. The D: ^{14}C Tracer Ratio in Aerobic Methanotrophy*

439 In aerobic methanotrophic cultures, a D: ^{14}C tracer ratio of ~1.5 was observed, suggesting
440 that on average, 2.67 of the four methane-derived hydrogen atoms likely enter water-exchangeable
441 products during the course of a full oxidation pathway. *M. tricosporium* is a type II methanotroph,
442 a member of the *Alphaproteobacteria* that uses the serine pathway for carbon assimilation; *M.*
443 *sedimenti* is a gammaproteobacterial type I methanotroph, using the RuMP carbon assimilation
444 pathway (Tavormina et al., 2015). The pathway data presented in Fig. 4 suggests that all methane-
445 bound hydrogens are water exchangeable during the catabolic oxidation of methane to carbon
446 dioxide. Thus, to achieve a D: ^{14}C tracer ratio less than 4, a substantial proportion of methane-
447 derived formaldehyde would need to proceed down the assimilatory pathway, a requirement that
448 was likely met given the cultures' increase in cell density (data not shown). Intriguingly, the D: ^{14}C
449 tracer ratios were similar for both cultured organisms despite their distinct metabolic pathways; a
450 similar phenomenon of consistent carbon conversion efficiency was recently observed among
451 distinct aerobic methanotroph communities in English riverbeds (Trimmer et al., 2015).

452 The oxic incubations of methane seep sediment produced a D: ^{14}C tracer ratio of 1.66 +/-
453 0.02 SE. Given that the known modes of biological methane oxidation – type I and type II aerobic

454 methanotrophy and reverse methanogenesis anaerobic methanotrophy – bound this observed value,
455 it appears likely that the oxic sediment incubations supported a mixture of both aerobic and
456 anaerobic methane oxidation processes. Aerobic methane oxidation likely dominated, based on the
457 ~ 7×10^4 Pa partial pressure of O₂ and the proximity of the D:¹⁴C tracer ratio to that of the aerobic
458 methanotrophic cultures, but anoxic niches likely remained or developed in the incubation bottles.

459 3.4. Specialized Application of the Monodeuterated Methane Approach: Anaerobic Methanotropy
460 at Pressure

461 To demonstrate the utility of the CH₃D rate measurement approach in addressing
462 experimentally relevant questions, particularly in nontraditional empirical contexts, we sought to
463 evaluate the influence of *in situ* pressure on methanotrophic rates of Hydrate Ridge seep sediment
464 microbial communities. Material collected for microbiological studies of AOM is frequently
465 obtained from marine settings of various depths that are subjected to distinct and substantial
466 pressure regimes (Ruff et al., 2015). Pressure is not always rigorously incorporated into microcosm
467 experiments, though evidence suggests it can be an important determinant of methanotrophic rates
468 (Bowles et al., 2011; Nauhaus et al., 2005; Zhang et al., 2010). In addition, some procedural
469 aspects of the ¹⁴CH₄ protocol, including headspace sampling and full-volume transfer, are not
470 established for use with mylar bags, making the monodeuterated methane approach an appealing
471 alternative in this context.

472 Parallel seep sediment incubations were subjected to 0.1 MPa (atmospheric pressure) and
473 9.0 MPa (equivalent to ~900 m depth). δD values derived from heightened D/H ratios attributable
474 to methane consumption, are shown in Fig. 5. A significant increase in methane consumption was
475 observed in both live conditions at heightened pressure, corresponding to sediment incubated with
476 glycine (samples 1a and 1b) and ammonium chloride (samples 2a and 2b). Controls lacking CH₃D

477 (samples 3a and 3b) and biological activity (samples 4a and 4b) showed no increase in δD (see
478 Table S2 for sample set-up details). The simulation of *in situ* Hydrate Ridge pressures led to a
479 79.5% (+/- 6.5 SE) increase in relative methane oxidizing rates. Incubation with 500 μM glycine
480 rather than ammonium at high and low pressures resulted in small but consistent rate increases of
481 12% +/- 4.1% SE, potentially reflecting the energetic and biosynthetic distinction between
482 exogenous amino acids and unprocessed fixed nitrogen.

483 Previous reports have found a wide range of different pressure-related effects. In a sulfate-
484 coupled AOM bioreactor, pressures were varied from 1 to 8 MPa and sulfide production
485 approximately tripled, demonstrating Michaelis-Menten style kinetics with an apparent K_m of 37
486 mM (Zhang et al., 2010). Methane partial pressures of 1.1 MPa led to a 5x increase in sulfate
487 reduction rates relative to ambient atmospheric pressure with Hydrate Ridge sediments
488 demonstrating methane-dependent sulfate reduction (Nauhaus et al., 2002). With methane seep
489 sediment from the Japan Trench, however, methane-driven sulfate reduction rates did not correlate
490 with changing pressure (Vossmeyer et al., 2012). Nauhaus et al. (2005) suggested that the pressure-
491 induced rate increases are due more to heightened methane solubility and bioavailability rather
492 than physiological effects or biomolecular re-ordering. Bowles et al. (2011) presented a very
493 different perspective by showing a six- to ten-fold AOM rate increase at 10 MPa when methane
494 concentrations were held constant. Deconvolving these two influences and how they depend on
495 community composition or physicochemical parameters is feasible with pressure chamber
496 experiments utilizing monodeuterated methane. Understanding the relative contributions of
497 environmental and physiological effects to methane oxidation will help constrain methane fluxes
498 across a larger envelope of the planet's methanotrophically active zones.

499 3.5. Using Monodeuterated Methane in Experimental Investigations

500 Based on $^{14}\text{CH}_4$ ground-truthing experiments with aerobic methanotrophic cultures, oxic
501 seep sediment, and anoxic seep sediment, as well as the proof-of-concept pressurized experiments,
502 we believe that the monodeuterated methane approach to methane oxidation rate measurement is a
503 useful addition to the biogeochemist's tool set. Compared with radiolabel approaches ($^{14}\text{CH}_4$, ^3H -
504 CH_4 , $^{35}\text{SO}_4^{2-}$), the method requires less safety-oriented planning, and is logically simpler, more
505 affordable, and may be less susceptible to hydrogen-associated isotope fractionation effects
506 (relative to ^3H). Our results also suggest that the monodeuterated methane technique appears to be
507 a more precise method based on standard error calculations (Figs. 1, 2). Direct comparisons of
508 environmental incubations are complicated by the microheterogeneity of seep settings (Barry et al.,
509 1996; Lloyd et al., 2010), as well as the fact that different aliquots of the same initial material were
510 used in our experiments. However, analysis of culture-based experiments reveals that standard
511 errors from R_{CH3D} values were 20% those derived from $^{14}\text{CH}_4$ -based values, making the
512 monodeuterated method five times more precise.

513 Because the monodeuterated methane method focuses on methane-bound hydrogen atoms,
514 it offers different, complementary information about methanotrophic systems than carbon-based
515 techniques like methane or bicarbonate quantification. While this distinction complicates the
516 interpretation of isolated D/H ratios, it can offer an additional dimension of information for
517 analysis of methane-derived intermediates in relevant metabolisms. Given these caveats, we
518 recommend three use cases for monodeuterated methane in methane oxidation rate measurement
519 applications.

520 1) First, the approach can be employed in a strictly comparative context using analogous
521 inoculum exposed to a range of different conditions, as demonstrated with the pressure-
522 based sediment incubations presented above. Evaluating the effect of different

523 conditions such as temperature ranges, chemical concentrations, or energetic landscapes
524 on seep sediment methane oxidizing rates would all be promising applications.

525 Comparative analysis of AOM rates at different seep sites would also be useful,
526 provided anaerobic or aerobic methanotrophic processes could be isolated.

527 2) Second, by performing side-by-side monodeuterated methane and radiocarbon tests, a
528 sample-specific D:¹⁴C tracer ratio can be determined, and absolute rates of full methane
529 oxidation can then be inferred in subsequent experiments based exclusively on D/H
530 ratios. Conducting such paired studies under additional environmental or lab-based
531 conditions would help clarify the universality of the ratios presented here and would
532 likely reveal additional questions of metabolic dynamics in a range of experimental
533 systems. We also encourage side-by-side comparisons with other rate measurement
534 approaches, including ³H-CH₄ radiotracer and methane concentration assessments, to
535 develop additional pairwise conversion factors and better constrain carbon and hydrogen
536 metabolism in methane-based biological reactions.

537 3) Finally, the use of monodeuterated methane as an analytical tool, alongside additional
538 methods such as carbon- or sulfur-tracking procedures, would enable a multi-
539 dimensional examination of anabolic and catabolic processes in methane-based
540 metabolisms. In particular, the D:¹⁴C tracer ratios presented here reveal intriguing and
541 seemingly systematic relationships between carbon and hydrogen anabolic and catabolic
542 partitioning across distinct physiologies, yet an underlying theoretical framework
543 regarding the fate of methane-bound hydrogen atoms remains outstanding. In anaerobic
544 methanotrophic systems, back-reaction rates and equilibrium constants could be
545 evaluated by a) including a ¹³CO₂ signal in the water and measuring ¹³CH₄, and/or b)

utilizing multiply deuterated methane as initial headspace and measuring all possible isotopologues via NMR or high resolution mass spectrometry. For aerobic methanotrophs, evaluating D:¹⁴C tracer ratios under more clearly defined growth and maintenance phases would elucidate distinct values associated with catabolic, RuMP, and serine pathways, enabling future use of that parameter as an arbiter of relative anabolic and catabolic activity. Furthermore, additional environmental variables can be tested to gain insight into distinct redox pathways and dynamics of reversibility. For example, AOM under lower sulfate concentrations might be expected to generate higher D:¹⁴C tracer ratios (Yoshinaga et al., 2014), and this parameter could be further developed as a measure of microbially mediated isotopic equilibration.

4. Conclusion

The ability to accurately measure methane consumption and oxidation rates – both comparatively and in absolute values – is an important component of methanotrophic studies. Such measurements frequently depend on radiotracers or measurements of chemical species that are related to, but not directly indicative of, methane metabolism. The monodeuterated methane technique presented here represents a novel approach to methane oxidation rate measurements, notable for its logistical and analytical ease (particularly in ship-board applications), as well as the added dimension provided by H-based, rather than C-based, information. We have demonstrated that the D/H ratio is a reliable proxy for methane oxidation activity: in several applications, the value is directly proportional to methane oxidation rates as measured in absolute terms by the well-established ¹⁴CH₄ method. The value of the proportionality constant differs based on the experimental system, likely dictated by environmental variables and the relative proportions of

569 aerobic and anaerobic methanotrophic metabolisms, though additional experiments to determine
570 the nature of the putative mixing line are needed.

571 Methane biogeochemistry is a dynamic field of study with implications for carbon cycling,
572 microbial ecology, and climate dynamics, though experimental challenges have slowed our
573 understanding of methane-based biological reactions. With the CH₃D approach as an added tool in
574 the arsenal of rate-based examinations, a broader understanding of the intricacies of methane
575 metabolism, as well as its role in environmental and anthropogenic systems, is within reach.

576

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769
770

771 **7. Tables and Figure Captions**

772 Table 1: A summary of the samples used for all experiments conducted in this study. Green boxes
773 indicate that the experiment took place (with all relevant permutations and controls, as described in
774 the text); blank boxes indicate experiments that were not conducted. CH₃D refers to
775 methanotrophic rate experiments using the novel monodeuterated methane technique, while ¹⁴CH₄
776 refers to the radiolabel-based experiments. The three-part codes for samples derived from
777 environmental material refer to active (A) or low-activity (L) sediments (Sed) or carbonates (Carb),
778 along with a sample-specific four-digit serial number.

779

	Oxic		Anoxic	
	CH ₃ D	¹⁴ CH ₄	CH ₃ D	¹⁴ CH ₄
Aerobic Methanotroph Cultures Experiment				
<i>M. trichosporium</i>				
<i>M. sedimenti</i>				
Seep Sediment Experiment				
A.Sed-5128				
L.Sed-5043				
Seep Carbonate Experiment				
A.Carb-5305				
A.Carb-5152				
L.Carb-5028				
Pressure Experiment				
A.Sed-3450				

780
781

782 Table 2: D:¹⁴C tracer ratios for the experimental treatments addressed in this study.

783

Aerobic Methanotroph Cultures		
	Exponential Phase	Stationary Phase
<i>M. trichosporium</i>	1.5	1.48
<i>M. sedimenti</i>	1.54	1.59
Methane Seep Sediments and Carbonates		
	Oxic Incubations	Anoxic Incubations
A.Sed-5128	1.62	2.05
L.Sed-5043	1.71	2.01
A.Carb-5305	1.65	1.96
A.Carb-5152	1.63	2.08
L.Carb-5028	1.69	1.86

784

785

786 Fig. 1: Amount of methane consumed over time for cultures of a) the type II methanotroph *M.*
 787 *trichosporium* and b) the type I methanotroph *M. sedimenti* using C_{corr} (values derived from the
 788 CH₃D method, shown with circles) and the ¹⁴CH₄ method (diamonds), calculated as discussed in
 789 the text. ¹⁴CH₄-derived data conveys values of methane consumption and full oxidation, while
 790 CH₃D-derived data provides a measure of methane consumption. Error bars show standard errors
 791 for three biological replicates, with the exception of the ¹⁴CH₄ -derived killed control (n=1).
 792 Obscured data points exhibited values between -60 and 110 nmol for a) and 0 and 60 nmol for b).

793

794 Fig. 2: Methanotrophy in a) oxic and b) anoxic incubations of active and inactive seep sediment
 795 and carbonate rocks (n=3 in all cases). Values compare methane consumption / full oxidation rates
 796 derived from ¹⁴CH₄ measurements (blue) and methane consumption rates derived from the CH₃D
 797 approach (green, R_{CH3D} values). Standard error bars are provided.

798

799 Fig. 3: A schematic diagram demonstrating the potential fate of methane-associated hydrogen
 800 atoms in the “reverse methanogenesis” pathway. Hydrogen atoms are distinguished by color and
 801 superscript number, and potential exchanges with inter- and intra-cellular water are shown.

802 Potentially detectable methane-derived hydrogen atoms (4, occurring throughout the oxidation
803 pathway) and carbon atoms (1, requiring full oxidation) are highlighted in orange and purple
804 boxes, respectively. Shorter “backflux” arrows reflect the observation that all enzymes (Thauer,
805 2008) and the entire pathway (Holler et al., 2011) have been shown to be reversible. For figure
806 simplicity, isotopically distinct backflux products and cofactor involvement in backflux reactions
807 are not shown. The extended dashed line represents the cell membrane.

808

809 Fig. 4: A schematic diagram demonstrating the potential fate of methane-associated hydrogen
810 atoms in the aerobic methanotrophy pathways. Hydrogen atoms are distinguished by color and
811 superscript number; asterisks represent location-specific ambiguity. Potentially detectable
812 methane-derived hydrogen atoms and carbon atoms are highlighted in orange and purple boxes,
813 respectively. Mmo enzymes are not believed to perform reversible reactions.

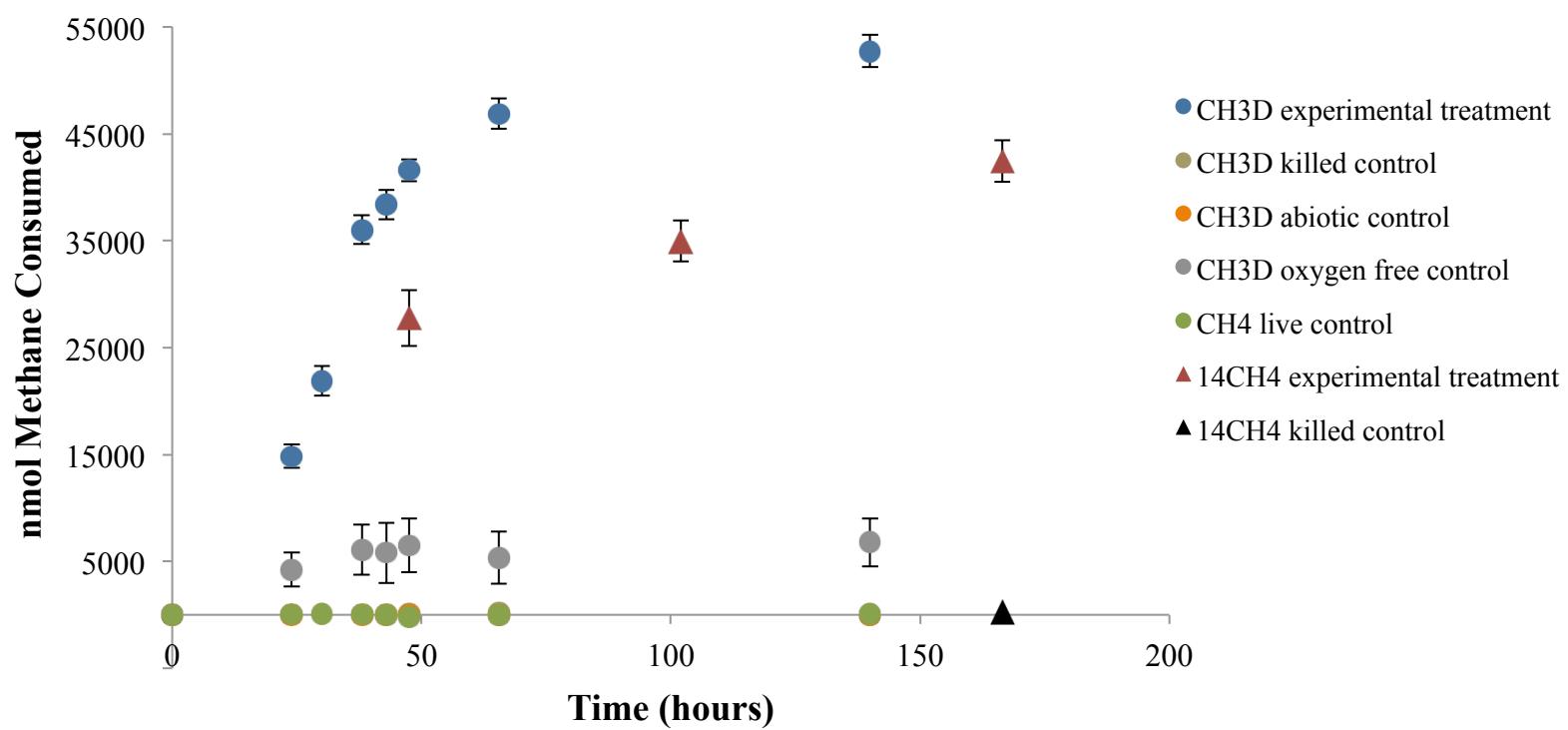
814

815 Fig. 5: Pressure experiment results showing water δD values with standard error bars of seep
816 sediment samples following 38-day incubations with CH₃D at 9.0 MPa (brown bars, “b” samples)
817 or 0.1 MPa (pink bars, “a” samples). Additional details on sample treatments can be found in Table
818 S2.

Figure 1

a)

Methanotrophy in *M. trichosporium* Cultures



b)

Methanotrophy in *M. sedimenti* Cultures

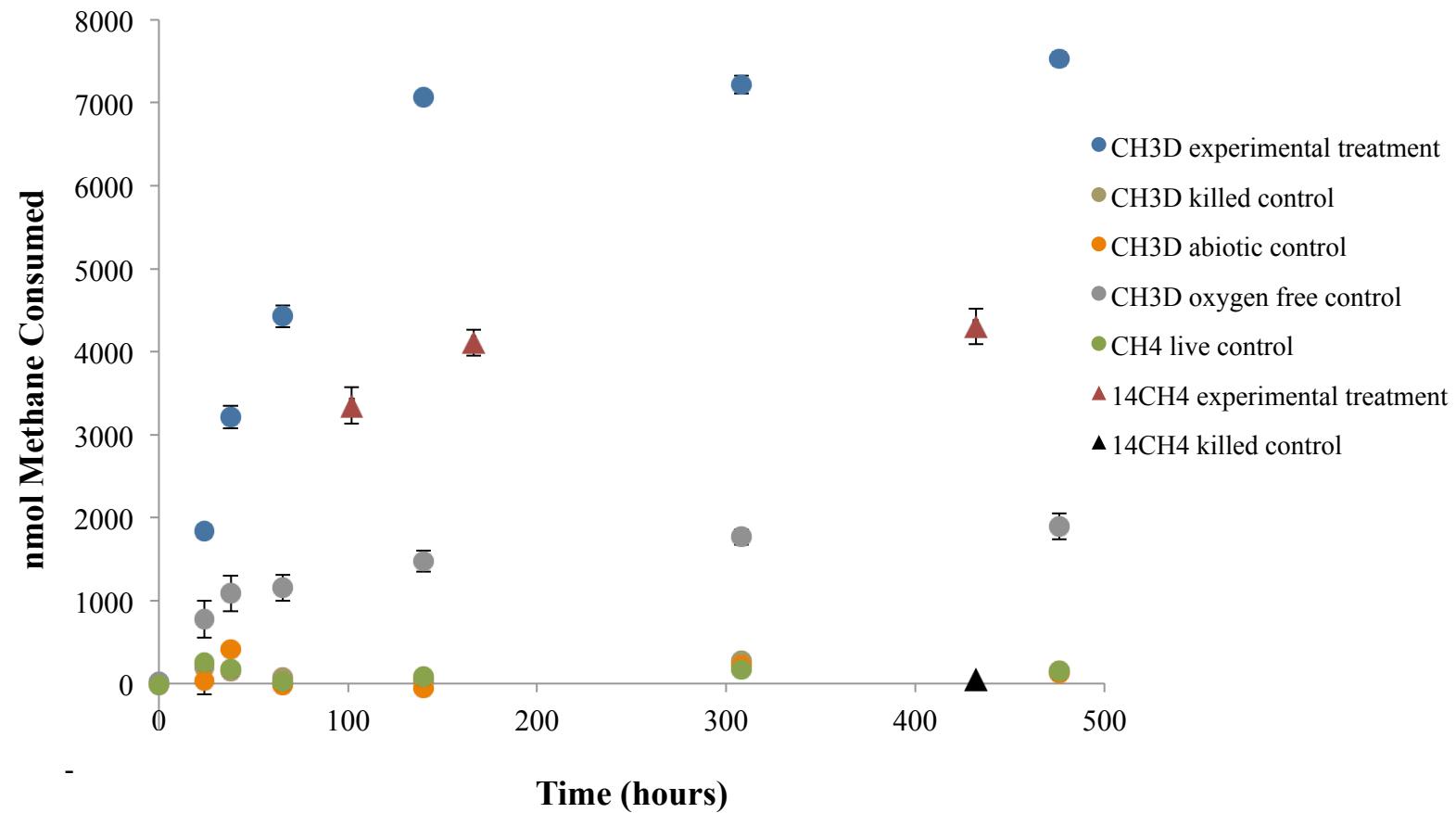


Figure 2

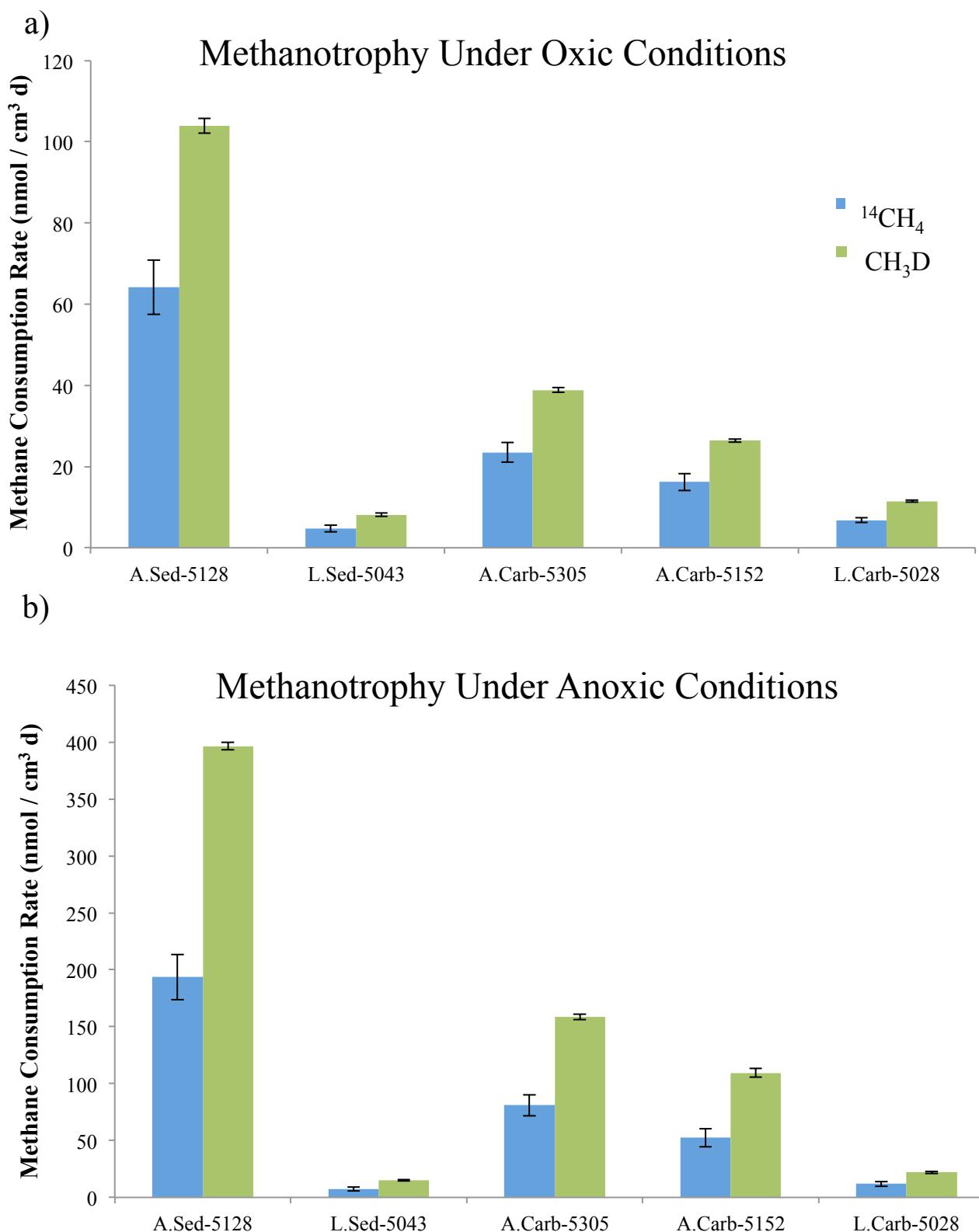


Figure 3

“Reverse Methanogenesis” Pathway

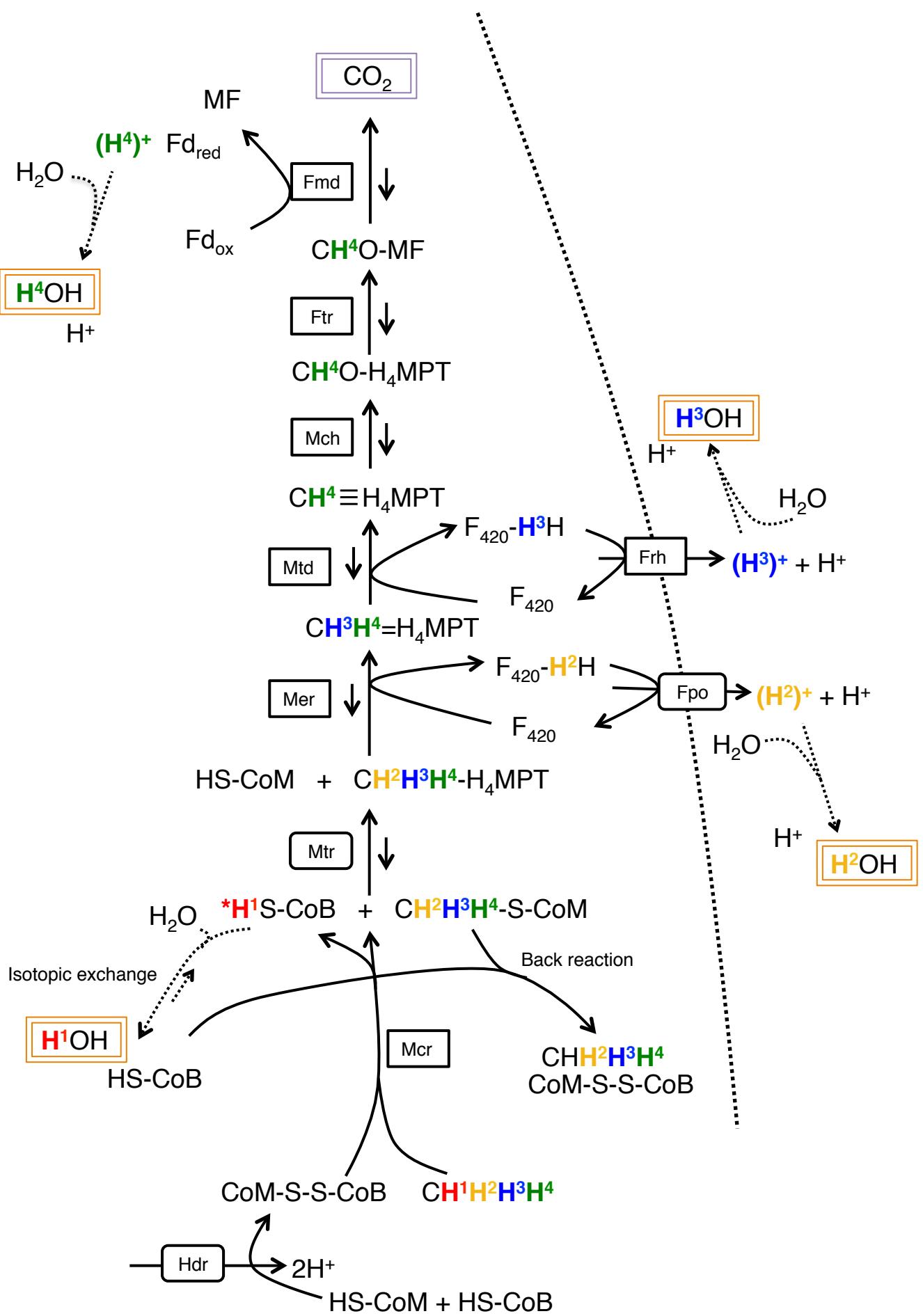


Figure 4

Aerobic Methanotrophy Pathway

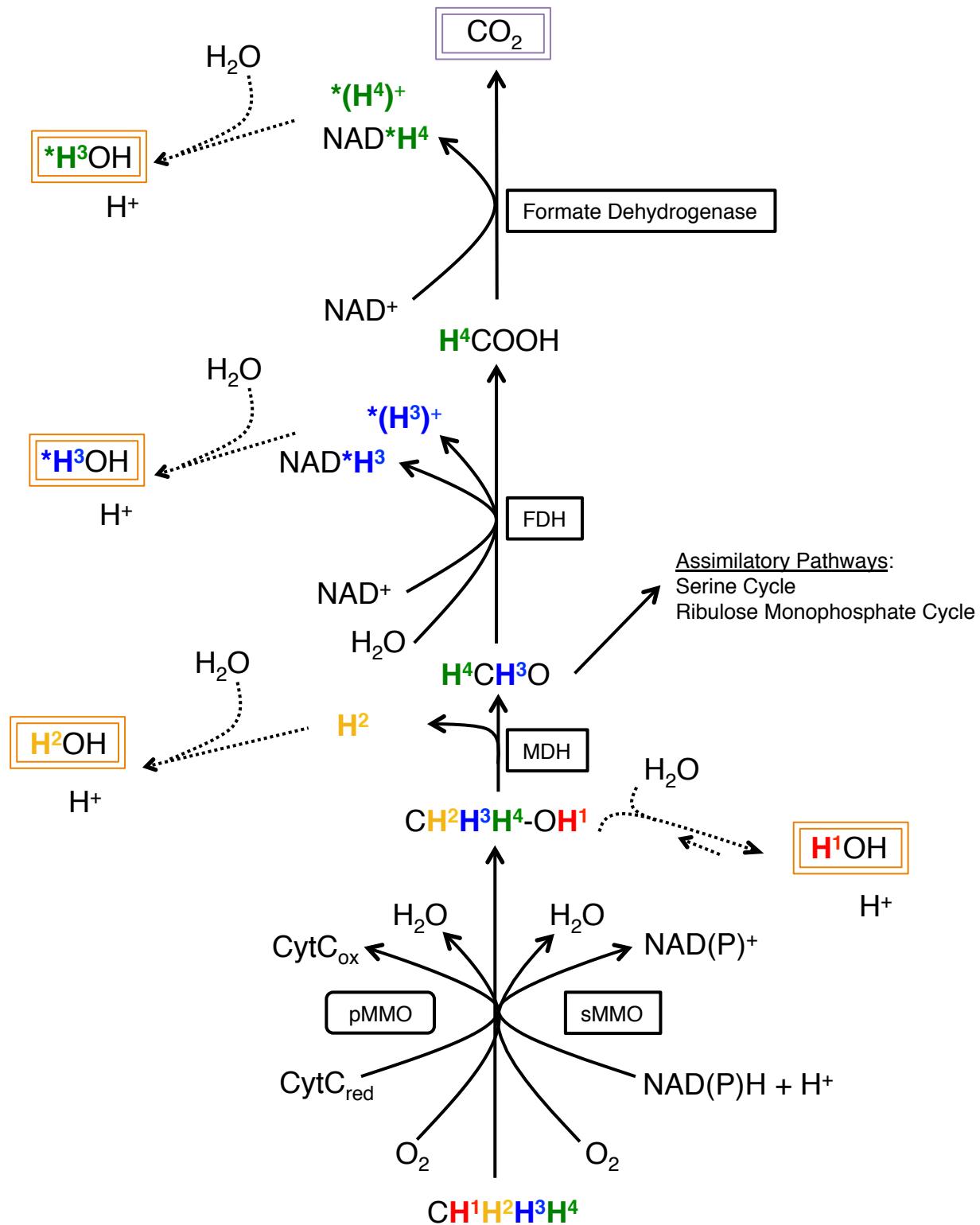


Figure 5

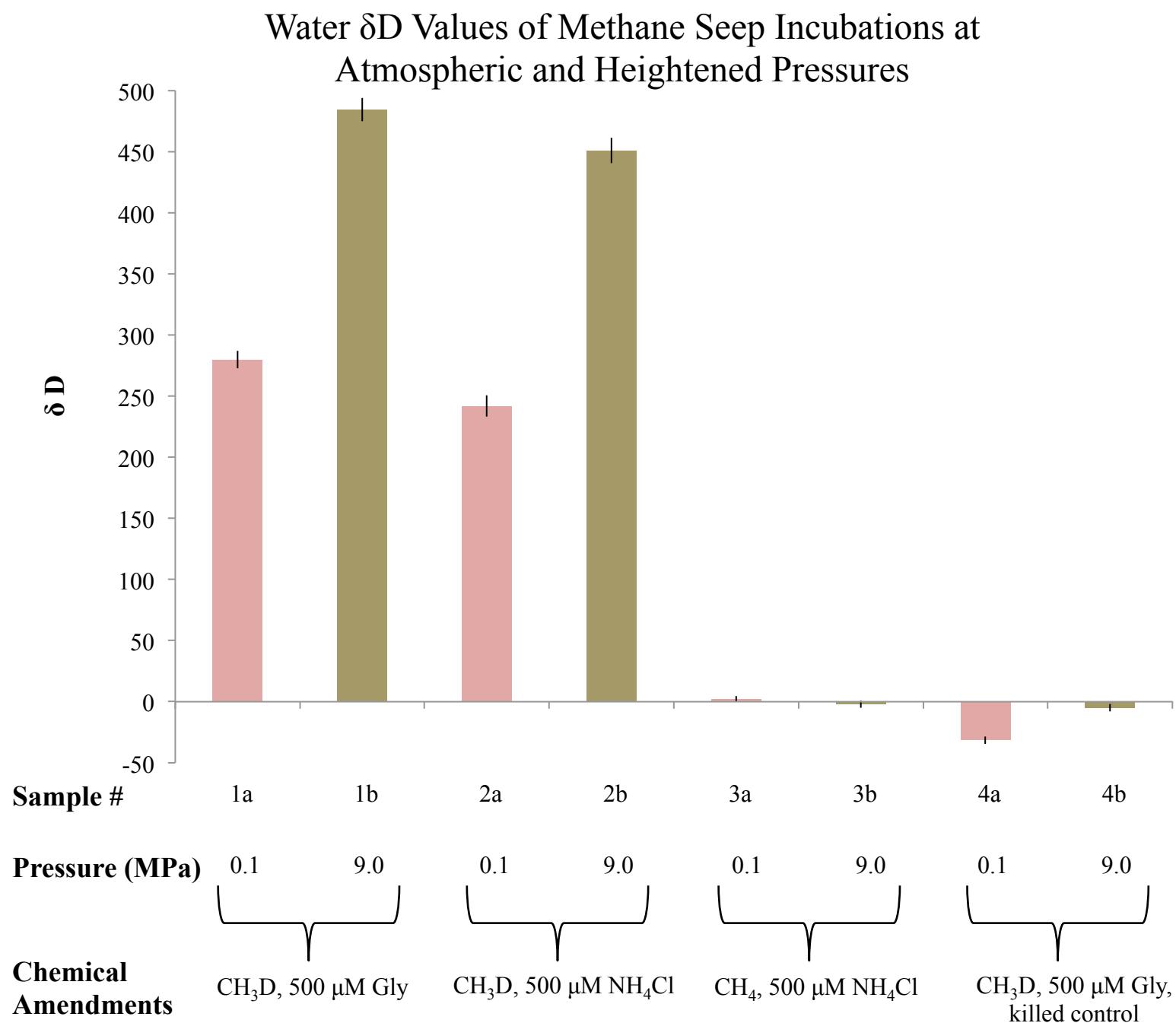
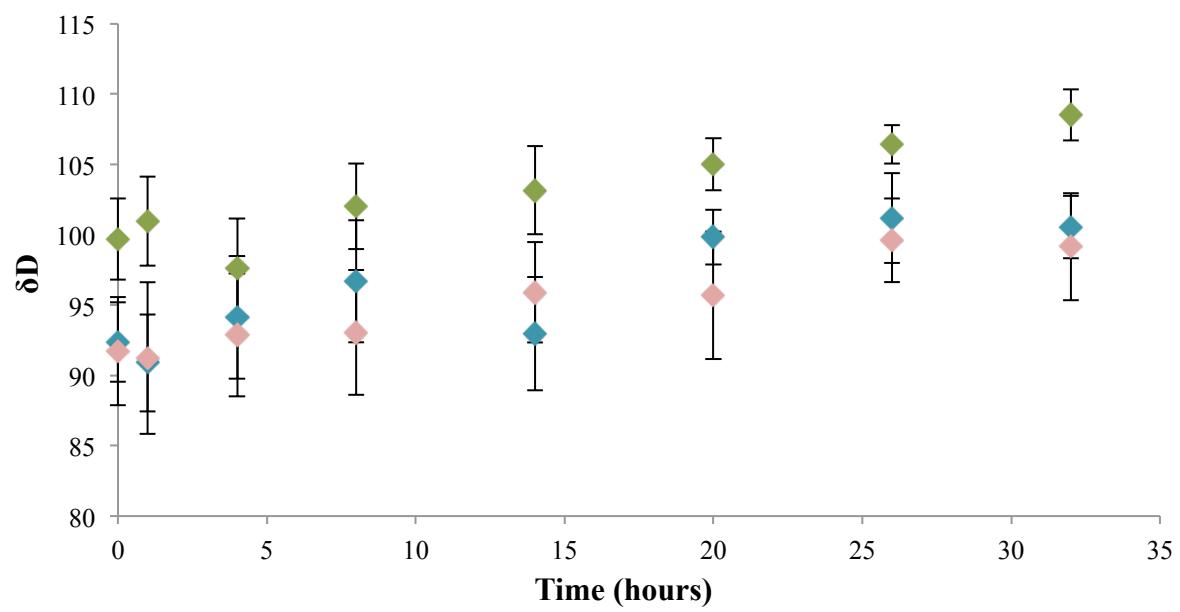


Figure S1



Supplementary Material for ““Monodeuterated Methane: An Isotopic Probe to Measure Biological Methane Metabolism Rates and Track Catabolic Exchange Reactions”

Table S1: Conditions for the aerobic methanotrophy experiments. All sample types were set up in triplicate; pressures shown as partial pressures.

Sample Condition	p(CH ₃ D)	p(CH ₄)	¹⁴ CH ₄	p(O ₂)	p(Ar)	Inoculum Introduced	Killed Cells
1	0.1 MPa			1 x 10 ⁵ Pa		10% v/v	
2	0.1 MPa			1 x 10 ⁵ Pa		10% v/v	Yes
3	0.1 MPa			1 x 10 ⁵ Pa			
4	0.1 MPa				1 x 10 ⁵ Pa	10% v/v	
5		0.1 MPa		1 x 10 ⁵ Pa		10% v/v	
6	0.1 MPa		13 kBq (T1)	1 x 10 ⁵ Pa		10% v/v	
7	0.1 MPa		13 kBq (T2)	1 x 10 ⁵ Pa		10% v/v	
8	0.1 MPa		13 kBq (T3)	1 x 10 ⁵ Pa		10% v/v	
9	0.1 MPa			1 x 10 ⁵ Pa		10% v/v	
10	0.1 MPa		13 kBq (T3)	1 x 10 ⁵ Pa		10% v/v	Yes

Table S2: The experimental set-up for methane seep sediment pressurized rate measurement incubations. The samples ran for 38 days at 4 °C, and each sample was contained in a sealed Mylar bag. Pressure values indicate absolute pressure exerted on the incubated Mylar bags.

Sample #	Sediment	Nitrogen Source	Methane Source	Pressure (MPa)
1a	50 mL	500 uM Glycine	40 mL CH ₃ D	0.1
2a	50 mL	500 uM NH ₄ Cl	40 mL CH ₃ D	0.1
3a	50 mL	500 uM NH ₄ Cl	40 mL CH ₄	0.1
4a	50 mL, killed control	500 uM Glycine	40 mL CH ₃ D	0.1
1b	50 mL	500 uM Glycine	40 mL CH ₃ D	9.0
2b	50 mL	500 uM NH ₄ Cl	40 mL CH ₃ D	9.0
3b	50 mL	500 uM NH ₄ Cl	40 mL CH ₄	9.0
4b	50 mL, killed control	500 uM Glycine	40 mL CH ₃ D	9.0

Fig. S1: To assess the empirical resolving power of the D/H measurement technique, we determined the time points showing non-overlapping confidence intervals for triplicate incubations of A.Sed-5128. Distinct signals were seen at the 20-hour time point for replicate A (teal diamonds) and the 26-hour time point for replicates B (pink diamonds) and C (green diamonds).

