

1 **Contrasting photosynthetic characteristics of forest versus**
2 **savanna species (Far North Queensland, Australia)**
3

4 Keith J. Bloomfield^{1,*}, Tomas F. Domingues^{2,**}, Gustavo Saiz^{3,***}, Michael I. Bird⁴,
5 Darren M. Crayn⁵, Andrew Ford⁶, Daniel J. Metcalfe⁶, Graham D. Farquhar¹ and
6 Jon Lloyd^{7,8,*}
7

8 [1] (Division of Plant Sciences, Research School of Biology, The Australian National University,
9 Canberra, ACT 0200, Australia)

10 [2] (Universidade de São Paulo, Faculdade de Filosofia Ciências e Letras de Ribeirão Preto, Brazil)

11 [3] (Karlsruhe Institute of Technology, Institute of Meteorology and Climate Research, Garmisch-
12 Partenkirchen, Germany)

13 [4] (School of Earth and Environmental Sciences and Centre for Tropical Environmental and
14 Sustainability Science, James Cook University, Cairns, Queensland, Australia)

15 [5] (Australian Tropical Herbarium and Centre for Tropical Biodiversity and Climate Change, James Cook
16 University, Cairns, Queensland, Australia)

17 [6] (CSIRO Ecosystem Sciences Tropical Forest Research Centre, Atherton, Queensland, Australia)

18 [7] (Department of Life Sciences, Imperial College London, Silwood Park Campus, SL5 7PY, United
19 Kingdom)

20 [8] (School of Marine and Tropical Biology and Centre for Tropical Environmental and Sustainability
21 Science, James Cook University, Cairns, Queensland, Australia)

22
23 [*] (Previously at: School of Geography, University of Leeds, LS2 9JT, United Kingdom)

24 [**] (Previously at: School of Geosciences, University of Edinburgh, EH8 9XP, United Kingdom)

25 [***] (Previously at: School of Geography & Geosciences, University of Saint Andrews, KY16
26 9AL, United Kingdom)

27
28

1 **Abstract**

2 Forest and savanna are the two dominant vegetation types of the tropical regions with very few
3 tree species common to both. At a broad scale, it has long been recognised that the distributions
4 of these two biomes are principally governed by precipitation and its seasonality, but with soil
5 physical and chemical properties also potentially important. For tree species drawn from a range
6 of forest and savanna sites in tropical far north Queensland, Australia, we compared leaf traits of
7 photosynthetic capacity, structure and nutrient concentrations. Area-based photosynthetic
8 capacity was higher for the savanna species with a steeper slope to the photosynthesis ↔
9 nitrogen (N) relationship compared with the forest group. Higher leaf mass per unit leaf area for
10 the savanna trees derived from denser rather than thicker leaves and did not appear to restrict
11 rates of light-saturated photosynthesis when expressed on either an area- or mass-basis. Median
12 ratios of foliar N to phosphorus (P) were relatively high (>20) at all sites, but we found no
13 evidence for a dominant P-limitation of photosynthesis for either forest or savanna trees. A
14 parsimonious mixed-effects model of area-based photosynthetic capacity retained vegetation
15 type and both N and P as explanatory terms. Resulting model-fitted predictions suggested a good
16 fit to the observed data ($R^2=0.82$). The model's random component found variation in area-
17 based photosynthetic response to be much greater among species (71% of response variance)
18 than across sites (9%). These results suggest that, on a leaf area-basis, savanna trees of far north
19 Queensland, Australia are capable of photosynthetically out-performing forest species at their
20 common boundaries.

21 **Key symbols and abbreviations**

| | |
|-----------------|---|
| FNQ | Far north Queensland |
| N | Nitrogen |
| P | Phosphorus |
| V | Vegetation type |
| F | Forest |
| S | Savanna |
| Φ | Plant functional type |
| A_{\max} | Rate of CO ₂ assimilation, light and CO ₂ saturated |
| A_N | Photosynthetic nitrogen-use efficiency |
| A_P | Photosynthetic phosphorus-use efficiency |
| g_s | Stomatal conductance to CO ₂ diffusion |
| d | Leaf (lamina) thickness |
| ρ | Leaf density |
| M_a | Leaf mass per unit area |
| ξ | Leaf dry matter content |
| Subscripted "a" | Per unit leaf area |
| Subscripted "m" | Per unit leaf dried mass |

1 **1. Introduction**

2 Forests and savannas dominate the tropical vegetated regions of the Earth covering 15-20% of
3 the Earth's surface (Torello-Raventos et al., 2013). At a broad scale, it has been long recognised
4 that the distribution of these two biomes, each with its own structural characteristics and species
5 composition, is principally governed by precipitation and its seasonality (Schmimper, 1903), but
6 with soil chemical characteristics also important (Lloyd et al., 2008; Lloyd et al., 2009; Lehmann
7 et al., 2011). Edaphic conditions are especially influential in regions where the two biomes
8 intersect – often referred to as 'ecotones' or 'Zones of (Ecological) Tension' (ZOT) - both forest
9 and savanna existing as discrete 'patches' under similar climatic conditions (Cochrane, 1989;
10 Ratter, 1992; Thompson et al., 1992; Hoffmann et al., 2009; Lehmann et al., 2011; Saiz et al.,
11 2012; Veenendaal et al., 2014). The patchiness of the ZOT mosaic at small spatial scales has led
12 some to argue that disturbances, principally fire, must interact with climatic/edaphic boundaries
13 in determining transition between the two alternative vegetation types (e.g. Lehmann et al., 2011;
14 Hoffmann et al., 2012). Whatever the drivers, feedbacks associated with changes to distributions
15 of these biomes in response to anthropogenic climate change have the potential to substantially
16 modify the rate of future global warming (e.g. Malhi et al., 2009).

17 The underlying causes of variation in photosynthetic carbon acquisition across and within these
18 two biomes remain, however, poorly understood. There is, nevertheless, accumulating evidence
19 that for tropical forest species phosphorus (P) availability may limit photosynthetic rates and
20 productivity (Vitousek, 1984; Domingues et al., 2010; Mercado et al., 2011; Quesada et al., 2012).
21 Whereas in savanna ecosystems nitrogen (N) may be more important as a limiting nutrient
22 (Lloyd et al., 2009). Soils in Australia are generally highly weathered with the consequence that
23 plant performance, even in the sub-tropical and temperate regions, is often considered more
24 limited by the supply of P than of N (Beadle, 1962, 1966; Webb, 1968). Nevertheless, in far
25 north Queensland (FNQ), where almost all the Australian tropical forest occurs, recent volcanic
26 activity (0.01 to 4.5 Ma BP) has produced some very young soils. These young basaltic soils
27 cover about 60% of FNQ's land area (Whitehead et al., 2007) and display higher levels of organic
28 matter and total P when compared with other parent material groupings such as granitic or
29 metamorphic (Spain, 1990). Direct links from soil P status to measures of forest productivity are
30 not straightforward, however, and inter-specific variations in P use efficiency are likely to have
31 contributed to the varied composition of local plant communities (Gleason et al., 2009). Likely
32 selection pressures, on infertile soils, for enhanced P use efficiency coupled with FNQ's recent
33 volcanic history mean that primacy here for P as the major limiting nutrient to photosynthetic
34 capacity is still hypothetical.

35 The forest and savanna vegetation types (\mathbb{V}) have very few plant species in common (Torello-
36 Raventos et al., 2013) and the edaphic determinants of the ZOT are of particular interest in
37 Australia (Beadle, 1962, 1966; Russell-Smith et al., 2004). The savannas of FNQ are distinctive
38 globally being dominated by eucalypts (Myrtaceae). Here, species of the closely related genera
39 *Eucalyptus* and *Corymbia* are characterised by sclerophyllous (hard) leaves with relatively low leaf
40 [N], but a high oil content and correspondingly high heat of combustion (Beadle, 1966) – traits
41 that contribute to a highly flammable leaf litter. In contrast to the tree species of the moist
42 forests, such evergreen savanna species are expected to be able to withstand periods of water
43 shortage and high water vapour pressure deficit. Sclerophylly imparts both structural and

1 physiological leaf traits, but, to date, most research has focused on the structural aspects: leaf
2 thickness and density combining in the ratio leaf mass per unit area (M_a , g m⁻²). Sclerophyllous
3 leaves are often amphistomatous (i.e. with stomata abundant on both the adaxial and abaxial
4 surfaces) displaying an isobilateral mesophyll distribution (Burrows, 2001) - characteristics
5 thought to be associated with both high photosynthetic potentials (Mott et al., 1982) and high
6 insolation environments (Pyykko, 1966; Parkhurst, 1978). Such eco-physiological associations are
7 expected on theoretical grounds, especially under conditions of low water availability (Buckley et
8 al., 2002) and it may be that sclerophyllous eucalypt leaves display an ‘investment strategy’ that at
9 once combines resilience with high photosynthetic return (Cernusak et al., 2011).

10 Broad overlap in leaf traits has been reported for savanna and forest tree species, but, due to
11 different soils and the drier conditions typical of savannas, we might expect differences between
12 the two vegetation types in leaf N and P content, rates of photosynthesis, morphology and
13 longevity (Wright et al., 2001; Buckley et al., 2002; Meir et al., 2007; Domingues et al., 2010). It
14 remains unclear, however, whether such differences persist within the relatively narrow
15 boundaries of a ZOT. In addition, within individual tropical forest stands, leaves can vary
16 markedly in their physiological and structural properties depending on canopy position and the
17 availability of gaps (Popma et al., 1992; Lloyd et al., 2010). Indeed tropical forest tree species are
18 often grouped according to their degree of shade tolerance and/or light requirement (e.g. Swaine
19 and Whitmore, 1988; but see Poorter, 1999). Shade tolerant species in the understory may
20 receive less than 2% of full light at the canopy crown (Chazdon, 1992) whilst pioneer species
21 typically require high light for germination and survival and depend on the creation of canopy
22 gaps (Turner, 2001). As a result, species characteristic of differing light niches have commonly
23 been considered to display distinctive photosynthetic traits linked to nutrient investment,
24 allocation and leaf architecture (Niinemets, 1997; Carswell et al., 2000; Wright et al., 2001). In the
25 context of P availability, a study in the forest of Guyana, on relatively infertile Ferralsols, found
26 that pioneer tree species exhibited higher photosynthetic nitrogen and phosphorus use
27 efficiencies than neighbouring climax species (defined here as those species whose seeds can
28 germinate and establish in the shade) (Raaimakers et al., 1995). Fyllas et al. (2012), in describing
29 forest tree species of the Amazon Basin, derived four discrete PFTs aligned with the species’
30 stature, canopy position and pioneering ability. It remains to be seen whether such an attractively
31 simple system can be applied to tropical forest species of FNQ.

32 In this study we contrast leaf photosynthetic traits for tree species from forest and savanna
33 communities of northern Australia addressing the following questions:

- 34 1. Are there differences in photosynthetic capacity and nutrient use efficiency between adjacent
35 forest and savanna vegetation types?
- 36 2. And if so, are these distinctions associated with systematic differences in leaf structural traits?
- 37 3. Is there evidence of a greater role for P rather than N (or vice versa) in determining
38 photosynthetic capacity across both sites and species?
- 39 4. Can a simple classification system based on light requirement and adult stature help to
40 describe observed variation in photosynthetic traits of tropical forest trees?

1 **2. Materials and method**

2 *2.1. Sites and species*

3 A series of sites was selected in FNQ, Australia in an arc from the Atherton Tablelands, inland
4 from Cairns, to Cape Tribulation, north of the Daintree River. The series, which forms part of
5 the *Tropical Biomes in Transition* (TROBIT) network, was designed to provide a contrast of
6 vegetation types, specifically forest (\mathcal{F}) and savanna (\mathcal{S}), and occurring on diverse soils (Table 1).
7 Further descriptions of all sites and the rationale (both structural and floristic) for our distinction
8 between \mathcal{F} and \mathcal{S} are available in Torello-Raventos et al. (2013). Seven sites were visited in six
9 weeks of fieldwork during April and May 2009 and measurements were performed on 125 trees
10 representing 30 species. A full list of species by site is presented in Table S1 (Supplementary
11 information).

12 *2.2. Gas exchange measurements*

13 Leaf gas exchange measurements were performed using a portable photosynthesis system (Li-
14 Cor 6400, Li-Cor, Lincoln, NE, USA) on young, fully developed leaves. During the
15 measurements, chamber conditions were set with block temperature (mean 27 °C) held slightly
16 above ambient air temperature to avoid problems of condensation; relative humidity remained
17 close to ambient (mean = 67%). The rate of air flow to the sample cell was held constant at 500
18 $\mu\text{mol s}^{-1}$, but exceptionally, and when faced with very low stomatal conductance, this was
19 reduced (minimum 250 $\mu\text{mol s}^{-1}$). Light ($A \leftrightarrow Q$) curves were generated for each tree species to
20 determine the saturating light level for adoption in subsequent CO_2 response curves ($A_a \leftrightarrow C_i$
21 curves). Those saturating light levels ranged from 500 to 2,000 $\mu\text{mol m}^{-2} \text{s}^{-1}$. Measurements of
22 light-saturated net CO_2 assimilation per unit leaf area (A_a) were then obtained for a range of
23 intercellular CO_2 concentrations (C_i) by varying chamber CO_2 concentration (C_a). The $A_a \leftrightarrow C_i$
24 curves provided area based values of light-saturated photosynthesis under ambient and elevated
25 atmospheric $[\text{CO}_2]$ ($A_{\text{sat},a}$ and $A_{\text{max},a}$ respectively). For the purposes of modelling photosynthetic
26 capacity we focus on variations in A_{max} - preferred over A_{sat} in this context as less susceptible to
27 limitations of stomatal conductance (g_s).

28 In the absence of cranes or suspended walkways, branches had to be cut from trees. Sun
29 exposed branches for low trees were reached using hand-held secateurs or forestry shears on
30 telescopic poles; for taller trees branches were pulled down using a weighted line shot from a
31 catapult. Trees of sub-canopy species were rarely found growing in full sunlight and so these
32 leaves, although sampled from upper branches free of self-shading, had developed in a relatively
33 low-light environment. Once detached, the stem was re-cut under water in order to re-establish
34 the xylem water column (Domingues et al., 2010). Performing gas exchange measurements on
35 excised branches can affect subsequent calculations where stomatal conductance is heavily
36 depressed (Santiago and Mulkey, 2003). The $A_a \leftrightarrow C_i$ curves were reviewed for such instances
37 and where necessary the data excluded from all further analysis ($n=11$). A further check on data
38 integrity proposed by Kattge et al. (2009) rejects those measurements where $A_{\text{sat}}/[\text{N}]_m$ is < 2
39 $\mu\text{mol CO}_2 \text{g}^{-1} \text{N s}^{-1}$: any such curves were likewise excluded ($n=2$).

2.3. Leaf morphology and nutrient analysis

At the completion of gas exchange measurements, the leaf (gas leaf) was cut from the branch and leaf thickness (d , μm) taken by averaging repeated measurements (Mitutoyo dial thickness gauge, $n=6$) alternating back and forward across the mid-vein and proceeding down the lamina from tip to base. A series of discs (6.6 mm, diameter) was then punched from the leaf avoiding veins and necrosis or other damage. The discs with the remainder of the leaf (petiole and mid-vein discarded) were oven dried at 70°C for a minimum of 48 h before their dried mass was recorded. The combination for the discs of known area and dried mass allowed the calculation of leaf mass per unit area (M_a , g m^{-2}). Leaf density (ρ , g cm^{-3}) was estimated using the equation:

$$\rho = M_a / d \quad (1)$$

In addition to the gas leaf, the opposing leaf was also cut from the branch, petiole and mid-vein discarded, and placed in a zip-lock plastic bag with moist cotton-wool until fresh mass could be measured that evening (or exceptionally the next day). The leaf was then placed in an envelope, oven-dried as above and dried mass recorded. The ratio of the leaf's dried to fresh mass is termed leaf dry matter content (ζ). All subsequent references to ζ relate to opposing and not gas leaves. Logistical constraints imposed by repeated changes of base camp and lack of electricity supply meant that delays were experienced between harvesting the leaves and oven drying (maximum delay 30 days).

Oven-dried material was used for determination of total leaf [N] and [P]: dried ground leaf material was acid-peroxide digested before colorimetric analysis using a segmented flow analyser (Skalar San⁺ System, Breda, The Netherlands). The photosynthetic efficiency of nutrient use was estimated as the maximal rate of carbon gain per unit of leaf N and P (A_N and A_P respectively).

2.4. Statistical analyses

All statistical analysis and modelling was conducted using the open-source statistical environment R (R Development Core Team, 2011). As initial data exploration revealed wide variation in many trait values across the different sites, non-parametric Kruskal-Wallis tests were used to test for differences among the categorical factors of Site and \mathcal{V} using the `coin` package (Hothorn et al., 2008). Where significant, differences among factor levels were assigned using Tukey's HSD post-hoc tests ($p < 0.05$) applied to data rankings. After exclusion of poor $A_a \leftrightarrow C_i$ curves ($n=13$, described above) and replicates with other missing values ($n=3$), the revised dataset of 109 leaf measurements contained many more observations for \mathcal{F} ($n=85$) than \mathcal{S} ($n=24$) and so there is an element of imbalance in the test specification where \mathcal{V} is adopted as the fixed factor. Bivariate relationships were described using Standardised Major Axis (SMA) line fits using `smatr-3` (Warton et al., 2006). Relationships between replicated foliar traits (photosynthesis, nutrient content) and site-dependent variables (soil, climate) were quantified using Kendall's non-parametric rank-order correlation (τ , τ); especially appropriate in cases with many replicated response values for each value of the predictor variable (Legendre and Legendre, 2012).

1 2.4.1. Mixed effects linear model of photosynthetic capacity

2 The study involved replicated measurements of tree species within and across forest and savanna
3 plots. Such a design introduced the strong likelihood that measurements within the same site
4 would be influenced by spatial proximity. In specifying a model that attempted to explain
5 differences in photosynthetic capacity between \mathcal{V} it was necessary to recognise this hierarchical
6 structure in order to avoid systematic variation in the residuals leading to potentially biased
7 interpretation (Zuur et al., 2009). The sites and species selected, rather than considered of
8 primary interest per se, were viewed as representative of a wider population and focus placed on
9 their variance. The model's random component therefore included the categorical variables of
10 species nested within sites. Unfortunately, because not all tree species at all sites were measured
11 with replication (see instances of $n=1$ at Table S1, Supplementary information), convergence
12 problems meant that the random component of the model could not accommodate differing
13 slopes as well as intercepts for species within a site.

14 The final model (fitted using the `n1me` package in R) may be expressed as:

$$15 \quad A_{\max,a[ijk]} = \alpha + \beta_1 B_{jk} + \beta_2 [N]_{a[ijk]} + \beta_3 [P]_{a[ijk]} + a_k + a_{j|k} + \varepsilon_{ijk} . \quad (2)$$

16 Here response variable $A_{\max,a[ijk]}$ denotes the maximum rate of area-based photosynthesis for
17 observation i of species j at site k with B being a categorical variable taking a base value of 0 for
18 species in plots classified as “forest” (\mathcal{F}) and a value of 1 for “savanna” (\mathcal{S}). The term a_k is a
19 random intercept and allows for variation among sites. The term $a_{j|k}$ allows for interspecific
20 variation at the same site. The term ε_{ijk} is the residual (unexplained) error and represents the
21 variation within-site, variation among plants of the same species and measurement error. Each of
22 the variation terms is assumed to be normally distributed with mean zero. With the independent
23 covariates centred (i.e., zeroed on the population mean), the fitted intercept term α thus
24 represents the predicted forest tree $A_{\max,a}$ at the ($\mathcal{F} + \mathcal{S}$) population mean $[N]_a$ and $[P]_a$. The term
25 $\beta_1 B$ represents the difference in predicted $A_{\max,a}$ between the two \mathcal{V} (in this case $A_{\max,a}[\mathcal{S}] -$
26 $A_{\max,a}[\mathcal{F}]$). The predicted \mathcal{S} tree $A_{\max,a}$ at the population mean $[N]_a$ and $[P]_a$ values is therefore α
27 $+ \beta_1$.

28 2.5. Plant functional types

29 \mathcal{F} species were assigned to one of four plant functional types (Φ) depending on their adult
30 stature and light requirements for recruitment (Veenendaal et al., 1996). Three of the authors
31 (DMC, AF and DJM), each with an extensive knowledge of Australian tropical forest trees, made
32 independent Φ assignments before combined results were consolidated and minor discrepancies
33 resolved. The Φ descriptions are provided in Table S2 (Supplementary information) and the
34 relevant species designations in Table S1 (Supplementary information). All of the tropical moist
35 forest species in this study are obligate evergreen.

1 **3. Results**

2 **3.1. Key leaf traits: Forest versus Savanna trees**

3 There was a ten-fold range across the dataset in photosynthetic capacity per unit leaf area ($A_{\max,a}$)
4 from 4.9 to 52.0 $\mu\text{mol m}^{-2} \text{s}^{-1}$ (Figure 1a). Mean values (treating each sampled tree as an
5 independent variable) differed among sites ($p < 0.0001$) and were significantly higher in the \mathcal{S}
6 plots (Figure 1b; $p < 0.0001$). When expressed per unit leaf dried mass, photosynthetic capacity
7 ($A_{\max,m}$) was also highly variable with significant plot-to-plot differences (Figure 1c; $p < 0.0001$).
8 Overall there was, however, no difference in mean $A_{\max,m}$ between \mathcal{F} and \mathcal{S} (Figure 1d; $p = 0.11$).
9 There were striking inter-site differences in leaf mass per unit area (M_a ; Figure 1e) which was
10 highly variable, spanning a five-fold range from lowest to highest observations. Overall, M_a was
11 higher for \mathcal{S} than for \mathcal{F} (Figure 1f; $p < 0.0001$). That contrast in M_a derived chiefly from variable
12 leaf dry matter content ratios (ζ) with a three-fold variation in ζ observed. Of note were
13 pronounced differences among sites (Figure 1g; $p < 0.0001$) and, like M_a , higher mean values for \mathcal{S}
14 than for \mathcal{F} (Figure 1h; $p = 0.0035$). By contrast, there was no difference between \mathcal{S} and \mathcal{F} for
15 measures of leaf thickness (d) (Figure 1j; $p = 0.70$) which was also much less variable.

16 Both area- and mass-based leaf nitrogen levels were highly variable and this was most
17 pronounced within the \mathcal{F} plots. On an area basis, differences among sites (Figure 1m; $p < 0.0001$)
18 produced higher mean $[\text{N}]_a$ values for \mathcal{S} than for \mathcal{F} (Figure 1n; $p = 0.0002$) with this inter-site N
19 variability even more pronounced on a mass basis (Figure 1o; $p < 0.0001$). Overall, mean $[\text{N}]_m$ was
20 higher for \mathcal{F} than for \mathcal{S} (Figure 1p; $p < 0.0001$). Broadly similar trends were observed for total
21 leaf phosphorus with higher mean $[\text{P}]_a$ for \mathcal{S} versus \mathcal{F} (Figure 1r; $p = 0.003$), and with that ranking
22 also reversed when expressed on a mass basis (Figure 1t; $p = 0.0004$). Foliar ratios $[\text{N}]:[\text{P}]$ ranged
23 from 10.1 (*Neisosperma poweri*, KBL-01) to 39.1 (*Symplocos hayesii*, KBL-03), but on average did not
24 differ across plots (Figure 1k; $p = 0.29$) and with no systematic differences between \mathcal{F} and \mathcal{S}
25 (Figure 1l; $p = 0.74$).

26 Considering the dataset as a whole (i.e. \mathcal{F} and \mathcal{S} trees combined), significant correlations of $A_{\max,a}$
27 with environmental variables such as elevation, temperature and soil cation status were found.
28 But with these mirrored by significant correlations of the same sign for both leaf $[\text{P}]_a$ and $[\text{N}]_a$
29 (Table S3 Supplementary Information). In investigating the underlying sources of our dataset's
30 tree-to-tree variation in photosynthetic properties, we therefore focussed (using the mixed
31 effects model) on associated tree-to-tree variations in leaf-based nitrogen and phosphorus
32 concentrations; checking for any edaphic or climatic effect beyond that through an examination
33 of model residuals in relation to the site-associated climate and soil covariates.

34 **3.2. Leaf nutrient relationships**

35 There was a strong positive linear relationship between leaf $[\text{P}]_a$ and $[\text{N}]_a$ for both \mathcal{F} and \mathcal{S}
36 (Figure 2a), but with a steeper slope observed for \mathcal{S} . The shallower slope for \mathcal{F} differs ($p = 0.031$)
37 from the mean relationship for tropical forests as suggested by Reich et al. (2009) as depicted by
38 the dotted line passing through the origin in Figure 2a (slope = $10^3/18.8$). When expressed on a
39 mass basis, there was a single, common P \leftrightarrow N relationship for both \mathcal{V} (Figure 2b). Leaf

1 nutrient investment on an area basis showed positive relationships with M_a for \mathcal{F} only (Figure 2c
2 and d).

3 To test for differences in the photosynthesis-nutrient relationships between the two \mathcal{V} , a series
4 of standardised major axis (SMA) analyses were undertaken with photosynthetic capacity (\mathcal{A}_{\max})
5 as the response variable and leaf chemistry ([N] and [P]) as the explanatory bivariate (Table 2).
6 For the combined dataset, linear relationships were strong for both nutrients irrespective of
7 whether variables were expressed on a mass- or area-basis (r values ranging from 0.63 to 0.70,
8 Table S4, Supplementary information).

9 The $\mathcal{A}_{\max,a} \leftrightarrow [\text{N}]_a$ association as shown in Figure 3a suggests two important differences between
10 the two \mathcal{V} . First: across the (pooled) dataset the lowest 0.3 fraction $[\text{N}]_a$ are confined to \mathcal{F}
11 associated trees (as can also be inferred from Figures 1 and 2). Second: for the lowest $[\text{N}]_a$ for \mathcal{S}
12 associated trees (*ca.* 1.6 g m⁻²) similar $\mathcal{A}_{\max,a}$ are observed for both \mathcal{S} and \mathcal{F} , but as $[\text{N}]_a$ increases
13 beyond that point $\mathcal{A}_{\max,a}$ for \mathcal{S} rises with a sensitivity nearly three times that observed for \mathcal{F} .
14 There was no difference between the two \mathcal{V} in either the slope or the intercept of the $\mathcal{A}_{\max,a} \leftrightarrow$
15 M_a association (Table S4, Supplementary information) and a single line ($r^2 = 0.3$) describes the
16 common positive relationship (Figure 3c).

17 3.3. Nutrient use efficiency and leaf structure

18 Elevated rates of photosynthesis per unit N yielded higher photosynthetic use efficiencies (\mathcal{A}_N)
19 for \mathcal{S} species (Figure 4a). Of the \mathcal{F} trees only the tall pioneers showed an equivalent \mathcal{A}_N to the \mathcal{S}
20 species and there was a significant difference between tall pioneers and shade tolerant species.
21 On the other hand, although there was a clear distinction between \mathcal{F} and \mathcal{S} species in \mathcal{A}_p , no
22 such distinction could be made among the four \mathcal{F} types.

23 The higher M_a values for the \mathcal{S} trees (Figure 1f) suggested underlying differences in leaf structure
24 between the two \mathcal{V} and the range of M_a values for separate Φ classes showed distributions
25 centred at different points along the M_a axis (Figure S1, Supplementary information). Whilst we
26 found positive relationships for M_a with d and ρ , the slopes of those relationships were \mathcal{V} -
27 dependent (Figure 5a and b) and the association was much stronger for the \mathcal{S} group. Over
28 common ranges of M_a , the \mathcal{F} trees displayed the thicker leaves (measured fresh) – a result heavily
29 influenced by the upper canopy group. Indeed, M_a was seldom greater than 75 g m⁻² for either
30 the small pioneer or sub-canopy classes, but for any given M_a the ratio of leaf dry mass to water
31 content (φ , as an index of sclerophylly) was much higher for the small pioneers (Figure 5c). The
32 slope of the relationship $\varphi \leftrightarrow M_a$ also differed among the Φ classes ($p < 0.0001$) – being
33 shallowest for the \mathcal{S} trees (an increased ratio of 0.005 per unit M_a) and steepest for the small
34 pioneers (an increase of 0.019). Such structural differences among classes of \mathcal{V} and Φ appeared
35 unrelated, however, to our measure of photosynthetic N allocation: \mathcal{A}_N was independent of d
36 ($p = 0.46$) and an overall correlation with M_a ($p = 0.0009$) disappeared upon controlling for \mathcal{V} and
37 Φ (Figure 5d).

1 3.4. Modelling photosynthetic capacity

2 In attempting to model variation in $\mathcal{A}_{\max,a}$ the starting, or maximal, fixed component of our linear
3 mixed effects model (Model 1, Table 3) included, along with a \mathcal{V} dependent term, those
4 continuous variables spanning leaf morphology and chemistry suggested by pair-wise correlation
5 plots at Figure S2 (Supplementary Information). The optimal fixed term, on AIC and likelihood
6 ratio criteria, was provided by Model 4: with vegetation type a fitted categorical variable
7 (through the $\beta_1 B$ term) along with the continuous variables $[\text{N}]_a$ and $[\text{P}]_a$. Model performance
8 was not improved by adding interaction terms (see model variants 8 through 12, Table 3). Also
9 of note is Model 6 which in dropping the B term produces a significantly inferior model
10 ($p=0.002$). That is to say, we could not account for the intrinsically higher area-based
11 photosynthetic capacities of \mathcal{S} affiliated trees (Fig. 1b) through systematic \mathcal{F} - \mathcal{S} differences in any
12 of our measured foliar traits. Comparisons against a model variant (Model 7, Table 3), excluding
13 the B term, but reinstating available traits relating to leaf chemistry and structure confirmed that
14 B could not be substituted in this way.

15 The full model, fixed and random terms combined, explained 82% of variation in the observed
16 values of $\mathcal{A}_{\max,a}$ and with all four forest Φ reasonably well predicted (Figure 6a). A plot of model
17 residuals against fitted values raised no concerns for the model assumptions (Figure S3,
18 Supplementary information). These model results also suggest, other things being equal, that
19 photosynthetic capacities are intrinsically higher for \mathcal{S} than \mathcal{F} species (Figure 1b) - with estimated
20 values at the dataset trait means of 29 and 18 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ (Table 4). Despite the relatively
21 low t value for the fitted $[\text{P}]_a$ term, the predictive power of the overall model was improved by
22 retaining $[\text{P}]_a$ as a fixed term (compare models 4 and 5 in Table 3). The greater significance
23 attached by the final model to the $[\text{N}]_a$ term is confirmed when the two slope coefficients are
24 standardised to allow for the difference in underlying units: $\beta_2' [\text{N}]_a = 0.294$ and $\beta_3' [\text{P}]_a = 0.172$
25 (interpreted as the relative effect on $\mathcal{A}_{\max,a}$ of one standard deviation change in the independent
26 variable).

27 Model output indicated that, after controlling for B and leaf nutrient levels, less than 10% of
28 variation in $\mathcal{A}_{\max,a}$ was attributable to site effects. The environmental influence on photosynthetic
29 capacity noted above (Table S3, Supplementary information) was, however, adequately captured
30 by our mixed model's fixed term (which incorporates leaf N and P), as shown by the absence of
31 any relationship between model residuals and those same site variables relating to climate and
32 soil conditions (Table S5 and Figure S4, Supplementary information).

33 Broadly similar results were obtained when the mixed modelling approach was repeated for
34 mass-based leaf traits (Table S6, Supplementary information). The fixed component of the final
35 parsimonious model again included B , $[\text{N}]_m$ and $[\text{P}]_m$, but this time with a significant N:P
36 interaction. The preferred random term was unchanged from that presented above.

37 3.5. Plant functional types

38 The area-based model's nested random component, which recognises the grouping of species
39 within sites, allows the variance of the response ($\mathcal{A}_{\max,a}$) to be partitioned among the available
40 terms. The fraction attributable to variance among plots (9%) is dwarfed by that among species

1 (71%). Variation in $A_{\max,a}$ around the population mean was therefore influenced much more by a
2 tree's taxonomy than the plot in which it was growing. One factor that may help explain this
3 inter-specific variation is plant functional type (Φ , related to a tree's growth strategy and light
4 requirements as described in Table S2, Supplementary information). We examined model output
5 for any Φ -related pattern in random intercepts for the different species. We found that tall
6 pioneers displayed higher intercepts than sub-canopy species ($p=0.0326$) as is shown in Figure
7 6b. Thus, at any given $[N]_a$ and $[P]_a$ tall pioneers typically achieved a higher $A_{\max,a}$ than shade
8 tolerant forest trees - as confirmed by their higher A_N and A_P (Figure 4). Small pioneers and
9 shade tolerant canopy species were intermediate between these two extremes and showed
10 intercepts close to the population mean.

11 **4. Discussion**

12 The main aim of this study was to compare photosynthetic traits for the tree species typical of
13 adjacent tropical moist forest and savanna plots – a dynamic boundary potentially sensitive to
14 changes in global climate (Sitch et al., 2008; Booth et al., 2012; Gloor et al., 2012; Huntingford et
15 al., 2013). Our findings include higher photosynthetic capacity and nutrient use efficiencies for
16 the savanna species, but our prediction of a primary photosynthetic role for P rather than N
17 across the forest sites was not supported. Our preferred area-based model retained only three
18 fixed terms: vegetation type, leaf N and P yet accounted for 82% of variation in $A_{\max,a}$. Model
19 output revealed a stronger relationship $A \leftrightarrow N$ than $A \leftrightarrow P$ and found variability among species
20 much more pronounced than among sites. For \mathcal{F} there was qualified support for the expectation
21 that pioneer species would show higher photosynthetic traits of $A_{\max,a}$ and A_N compared to late
22 successional shade tolerant species (Raaimakers et al., 1995).

23 **4.1. Forest and savanna comparisons**

24 Values reported here for key leaf traits such as $A_{\max,a}$, M_a and levels of foliar N and P fell within
25 previously published ranges for \mathcal{F} and \mathcal{S} trees (e.g. Medina, 1984; Prior et al., 2005; Harrison et
26 al., 2009; Cernusak et al., 2011). There were, however, significant differences among sites and
27 between \mathcal{F} and \mathcal{S} in all these traits (Figure 1). In particular, a recently cyclone-affected \mathcal{F} site
28 south of Ravenshoe (KBL-01) stood out as high in leaf nutrients and photosynthetic capacity
29 when expressed on a mass basis. Due to lower M_a , however, that prominence was all but
30 removed when area-based traits were examined.

31 Our study included measurements of 30 tree species across seven sites; these sites and species
32 being viewed as representative of wider populations and our modelling treatment of those terms
33 therefore focusing on their influence on the variance of the photosynthetic response rather than
34 mean values. The linear mixed effects model (Table 4), through its random component of
35 species nested within sites, showed that most of the variance in the data occurred among species
36 (71%). Once levels of leaf N and P had been included in the model, variation among sites
37 represented less than 10% of total variation. This corresponds with the findings of other
38 Australian studies where within-site variation has proved much larger than that across sites (e.g.
39 Wright et al., 2004; Asner et al., 2009). For this study, it could be argued, however, that the
40 climatic and topographical ranges spanned by the seven sites (Table 1) were rather modest –

1 mean annual precipitation, for example, is nowhere lower than 1.3 m and the range in mean
2 annual temperatures is only 6 °C.

3 Lower mass-based leaf nutrient values for \mathcal{S} species have traditionally been linked to their higher
4 M_a associated with contrasts in leaf longevity and economic strategy. In the Australian literature,
5 these species are widely described as sclerophyllous, characterised by tough leaves and
6 adaptations to limit water loss. We argue, however, that on theoretical grounds it is area- rather
7 than mass-based concentrations of N (and presumably also P, where relevant for photosynthetic
8 carbon gain) that should be modulated by differences in water availability. With declining
9 precipitation therefore, an increase in area-based photosynthetically important nutrients (in our
10 case $\mathcal{S} > \mathcal{F}$) seems to be the general case (Buckley et al., 2002; Cernusak et al., 2011; Domingues
11 et al., 2014; Schrodtt et al., 2014). It is non-systematic variations in M_a with precipitation, in turn
12 probably attributable to differences in rainfall seasonality, that produce contradictions in mass-
13 based N \leftrightarrow precipitation relationships (Schrodtt et al., 2014).

14 **4.2. Linking leaf structure to metabolism**

15 At the leaf level, \mathcal{A}_N is dependent upon a number of factors including N allocation, conductance
16 and Rubisco kinetics and no single cause has been found to explain observed interspecific
17 differences (Hikosaka, 2004; Hikosaka and Shigeno, 2009). The idea that species with high M_a
18 exhibit low \mathcal{A}_N due to greater structural investment (e.g. Takashima et al., 2004) has been
19 countered by later studies that found no relationship between \mathcal{A}_N and the proportion of leaf N
20 allocated to cell walls (Harrison et al., 2009; Hikosaka and Shigeno, 2009). Indeed, our general
21 positive association between area-based photosynthetic capacity and M_a - also observed by
22 Domingues et al. (2014) - challenges the general notion that thick sclerophyllous leaves should be
23 characterised by low photosynthetic rates and/or low photosynthetic nutrient use efficiencies
24 (Wright et al., 2004; Westoby and Wright, 2006). Certainly, it has long been known that typically
25 sclerophyllous eucalypt species can have exceptionally high photosynthetic rates (Larcher, 1969)
26 with Denton et al. (2007) also finding very high nutrient use efficiencies for numerous
27 xeromorphic Proteaceae species that exhibit some of the very highest M_a worldwide. Maximov
28 (1929) noted: “the drier the habitat, the smaller and more leathery the leaves, and the higher their
29 rate of transpiration”.

30 Interestingly, our best-fit photosynthesis model was not improved by the inclusion of
31 morphological traits such as M_a , leaf thickness or the ratio of leaf dry to fresh mass. Although
32 M_a was much higher for the \mathcal{S} plots, there was no difference in leaf thickness between \mathcal{S} and \mathcal{F}
33 sites ($p=0.95$) suggesting that most of the difference in M_a between the two vegetation types was
34 attributable to a higher leaf density for \mathcal{S} . But, as noted above, with no adverse effects on
35 photosynthetic nutrient use efficiencies.

36 **4.3. The role of phosphorus**

37 It has long been considered that vegetation differentiation in Australia is strongly influenced by
38 edaphic constraints and specifically soil P status (Webb, 1968; Russell-Smith et al., 2004). The
39 widely observed positive correlation leaf [N] \leftrightarrow [P] (e.g. Wright et al., 2001) is evident here as
40 well (Figure 2a), but the slope of the relationship differed between \mathcal{F} and \mathcal{S} . Despite their

1 situation in the Atherton basalt province (Whitehead et al., 2007), the mean foliar N:P ratios for
2 all of the sites visited in this study were far above thresholds believed to constitute P-deficiency
3 (Townsend et al., 2007; Cernusak et al., 2010). This is particularly striking since the forests
4 studied here, even after excluding the sub-canopy trees, had median $[N]_a$ of only 1.63 g m^{-2} . Such
5 N levels are low compared to other tropical forests for which extensive data have recently
6 become available - see Table 2 of Domingues et al. (2014). Values here are lower, for example,
7 than Cameroon (2.12 g m^{-2}) and lower even than for trees growing on soils of low nutrient
8 availability in the Amazon Basin (1.90 g m^{-2}). Foliar P levels, however, were lower still despite
9 concentrations of total soil extractable P (P_{ex}) for our forest plots being relatively high ($165\text{-}958$
10 $\mu\text{g g}^{-1}$, Table 1). For what have been classed 'high nutrient' soils in the Amazon Basin, for
11 example, Fyllas et al. (2009) reported median foliar P of 1.11 mg g^{-1} and 0.11 g m^{-2} with
12 equivalent soil P_{ex} ranging from $52\text{-}422 \mu\text{g g}^{-1}$ (Quesada et al., 2010). In our study, however,
13 upper canopy forest trees displayed median P values of only 0.76 mg g^{-1} and 0.08 g m^{-2} , much
14 closer to Amazon forest trees growing on 'low nutrient status' soils for which soil P_{ex} ranges
15 from $14\text{-}87 \mu\text{g g}^{-1}$ (Quesada et al., 2010) with foliar P median values of 0.7 mg g^{-1} and
16 0.06 g m^{-2} (Fyllas et al., 2009).

17 This apparent 'inability' of Australian forest trees to utilise high soil P availabilities could perhaps
18 be related to their unique evolutionary history. Essentially of Gondwanan origin (Crisp et al.,
19 1999), today's forests represent remnants of more temperate moist forests that covered much of
20 the continent until the mid-Miocene (Adam, 1992). Presumably arising from a flora already
21 adapted to the characteristically P-limited soils of much of Australia (McKenzie et al., 2004) it
22 may be that many of the species occurring within the Australian tropical forest region suffer
23 from an 'evolutionary hangover' lacking the ability to utilise high levels of soil P even where
24 available. There is in addition, especially for the lowlands, a considerable Asia-derived element in
25 the Australian tropical forest flora (Sniderman and Jordan, 2011; Crayn et al., 2014) and many
26 soils of the Asian lowland tropical forest region are also of relatively low nutrient status (Acres et
27 al., 1975; Tessins and Jusop, 1983; Ohta and Effendi, 1992; Banin et al., 2014). By comparison,
28 despite the generally lower P status of the savanna soils (Table 1), the savanna trees in our study
29 had a slightly higher median $[P]_a$ than those of the forest ($\mathcal{S} = 0.08 \text{ g m}^{-2}$ and $\mathcal{F} = 0.06 \text{ g m}^{-2}$) and
30 this was true also for $[N]_a$ ($\mathcal{S} = 2.09 \text{ g m}^{-2}$ and $\mathcal{F} = 1.62 \text{ g m}^{-2}$). This finding for the Australian
31 species contrasts with previous work in West Africa and Cameroon where area-based N and P
32 concentrations were lower for savanna than for forest species (Domingues et al., 2014; Schrodte
33 et al., 2014).

34 Despite these differences in area-based nutrient concentrations, there are notable consistencies
35 between our results and the African studies mentioned above. First, albeit with different model
36 constructs, there is clear evidence of a role for both N and P in the modulation of
37 photosynthetic rates in the field. Second, other things being equal, it seems that savanna trees
38 have higher N use efficiency than their forest counterparts. This higher \mathcal{A}_N (Figure 4) may
39 reflect differences in leaf construction linked to the higher light environment. Earlier studies
40 have suggested that lower \mathcal{A}_N values for sclerophytes may be caused by limitations to internal
41 conductance caused by leaf structural factors linked to greater leaf longevity (e.g. Warren, 2008).
42 Mesophyll conductance (g_m) is, however, the complex and variable product of at least three
43 phases acting in series (Flexas et al., 2008): conductance through intercellular air spaces (g_{ias}),

1 through cell walls (g_w) and through the liquid and membrane phases inside cells (g_{liq}). The most
2 important constraint on g_m is thought to be g_{liq} which is the phase least affected by leaf structure.

3 Overall, our results suggest a complex effect of P on photosynthetic capacity for these Australian
4 tree species. Foliar $[P]_a$ was only marginally significant in the preferred linear mixed effects
5 model (Table 4), but its inclusion improved the overall predictive power. When examined on a
6 mass basis, P did, however, appear more critical and with an N:P interaction term also included
7 in the optimal model (**Error! Reference source not found.**). This mass- versus. area-based
8 inconsistency in the apparent importance of P as a modulator of photosynthetic rates was also
9 noted by Domingues et al. (2014) who likewise found their mass-based models to include an N:P
10 interaction term not present in the area-based version. The sudden appearance of apparently
11 significant terms when transforming area-based entities to a mass basis is, however, to be
12 expected (Lloyd et al., 2013).

13 *4.4. Describing trait variation using plant functional types*

14 Where possible in the \mathcal{F} sites, tree species were selected in order to provide a contrast of light
15 environment as described in the assigned categories of plant functional type (Φ). Such
16 categorisation is often problematic and especially in the setting of boundaries from one group to
17 another (e.g. Poorter, 1999). For Australian tropical moist forests, Webb developed a
18 classification system of 20 structural vegetation types along climatic and edaphic gradients
19 (1968). Faced with such complexity and subjectivity, many authors have instead argued for a
20 spectrum of vegetation types or habits (e.g. Coste et al., 2005). The current study used Φ to
21 attempt to explain residual patterns in the data after controlling for \mathcal{V} (Figure 6b). As
22 hypothesised, there was evidence that pioneer trees of the \mathcal{F} showed higher photosynthetic
23 capacity and nutrient use efficiency than those shade tolerant species which persist in the
24 understory (Figure 4).

25 *4.5. Conclusions*

26 Comparing tree species of neighbouring forest and savanna sites in far north Queensland,
27 Australia we found higher photosynthetic capacity for savanna species linked to greater nitrogen
28 use efficiency. Higher leaf mass per unit area for the savanna trees derived from greater leaf
29 density rather than thickness, but did not inhibit photosynthetic capacity or \mathcal{A}_N . Our mixed
30 effects model accounted for 82% of variation in observed photosynthetic response and
31 confirmed the importance of the $\mathcal{A} \leftrightarrow \mathcal{N}$ relationship. The model, whilst retaining leaf P as a key
32 explanatory term, did not support the prediction of a dominant role for P rather than N in
33 determining rates of photosynthesis for Australian tropical forest species.

1 **Acknowledgments**

2 We would like to thank Jeanette Kemp (Australian Wildlife Conservancy) and Matt Bradford
3 (CSIRO, Atherton) for all their help with logistics and in the field. Thanks also to Prof. Owen
4 Atkin (Australian National University, Canberra) for many fruitful discussions. This work was
5 supported by the UK Natural Environment Research Council (Reference NE/F002165/1), a
6 Royal Society of London UK-Australia Exchange Award to Jon Lloyd, and the Australian
7 Research Council (Reference DP0986823).

References

- Acres, B. D., Bower, R. P., Burrough, P. A., Folland, C. J., Kalsi, M. S., Thomas, P., and Wright, P. S.: The soils of Sabah, Ministry of Overseas Development, Surrey, England, 1975.
- Adam, P.: Australian rainforests, Clarendon Press, Oxford, 1992.
- Asner, G. P., Martin, R. E., Ford, A. J., Metcalfe, D. J., and Liddell, M. J.: Leaf chemical and spectral diversity in Australian tropical forests, *Ecological Applications*, 19, 236-253, 2009.
- Banin, L., Lewis, S. L., Lopez-Gonzalez, G., Baker, T. R., Quesada, C. A., Chao, K.-J., Burslem, D. F. R. P., Nilus, R., Abu Salim, K., Keeling, H. C., Tan, S., Davies, S. J., Monteagudo Mendoza, A., Vásquez, R., Lloyd, J., Neill, D. A., Pitman, N., and Phillips, O. L.: Tropical forest wood production: a cross-continental comparison, *Journal of Ecology*, n/a-n/a, 2014.
- Beadle, N. C. W.: Soil phosphate and delimitation of plant communities in Eastern Australia II, *Ecology*, 43, 281-288, 1962.
- Beadle, N. C. W.: Soil phosphate and its role in molding segments of Australian flora and vegetation with special reference to xeromorphy and sclerophylly, *Ecology*, 47, 992-1007, 1966.
- Booth, B. B. B., Jones, C. D., Collins, M., Totterdell, I. J., Cox, P. M., Sitch, S., Huntingford, C., Betts, R. A., Harris, G. R., and Lloyd, J.: High sensitivity of future global warming to land carbon cycle processes, *Environmental Research Letters*, 7, 2012.
- Buckley, T. N., Miller, J. M., and Farquhar, G. D.: The mathematics of linked optimisation for water and nitrogen use in a canopy, *Silva Fennica*, 36, 639-669, 2002.
- Burrows, G. E.: Comparative anatomy of the photosynthetic organs of 39 xeromorphic species from subhumid New South Wales, Australia, *International Journal of Plant Sciences*, 162, 411-430, 2001.
- Carswell, F. E., Meir, P., Wandelli, E. V., Bonates, L. C. M., Kruijt, B., Barbosa, E. M., Nobre, A. D., Grace, J., and Jarvis, P. G.: Photosynthetic capacity in a central Amazonian rain forest, *Tree Physiol.*, 20, 179-186, 2000.
- Cernusak, L. A., Hutley, L. B., Beringer, J., Holtum, J. A. M., and Turner, B. L.: Photosynthetic physiology of eucalypts along a sub-continental rainfall gradient in northern Australia, *Agricultural and Forest Meteorology*, 151, 1462-1470, 2011.
- Cernusak, L. A., Winter, K., and Turner, B. L.: Leaf nitrogen to phosphorus ratios of tropical trees: experimental assessment of physiological and environmental controls, *New Phytologist*, 185, 770-779, 2010.
- Chazdon, R. L.: Photosynthetic plasticity of two rain forest shrubs across natural gap transects, *Oecologia*, 92, 586-595, 1992.
- Cochrane, T. T.: Chemical properties of native savanna and forest soils in central Brazil, *Soil Science Society of America Journal*, 53, 139-141, 1989.
- Coste, S., Roggy, J. C., Imbert, P., Born, C., Bonal, D., and Dreyer, E.: Leaf photosynthetic traits of 14 tropical rain forest species in relation to leaf nitrogen concentration and shade tolerance, *Tree Physiol.*, 25, 1127-1137, 2005.
- Crayn, D. M., Costion, C., and Harrington, M.: The Sahul-Sunda floristic exchange: dated molecular phylogenies document post-Miocene intercontinental dispersal dynamics, *Journal of Biogeography*, In review, 2014.
- Crisp, M. D., West, J. G., and Linder, H. P.: Biogeography of the terrestrial flora. In: *Flora of Australia*, Orchard, A. E. and Thompson, H. S. (Eds.), CSIRO, Melbourne, 1999.
- Denton, M. D., Veneklaas, E. J., Freimoser, F. M., and Lambers, H.: *Banksia* species (Proteaceae) from severely phosphorus-impooverished soils exhibit extreme efficiency in the use and re-mobilization of phosphorus, *Plant Cell Environ.*, 30, 1557-1565, 2007.
- Domingues, T. F., Ishida, F. Y., Feldpausch, T. R., Grace, J., Meir, P., Saiz, G., Sene, O., Schrodte, F., Sonké, B., Taedoumg, H., Veenendaal, E. M., Lewis, S., and Lloyd, J.:

1 Biome-specific effects of nitrogen and phosphorus on the photosynthetic characteristics
2 of trees at a forest-savanna boundary in Cameroon, *Oecologia*, In submission, 2014.

3 Domingues, T. F., Meir, P., Feldpausch, T. R., Saiz, G., Veenendaal, E. M., Schrodte, F., Bird, M.,
4 Djagbletey, G., Hien, F., Compaore, H., Diallo, A., Grace, J., and Lloyd, J.: Co-limitation
5 of photosynthetic capacity by nitrogen and phosphorus in West Africa woodlands, *Plant
6 Cell Environ.*, 33, 959-980, 2010.

7 Flexas, J., Ribas-Carbo, M., Diaz-Espejo, A., Galmes, J., and Medrano, H.: Mesophyll
8 conductance to CO₂: current knowledge and future prospects, *Plant Cell Environ.*, 31,
9 602-621, 2008.

10 Fyllas, N. M., Patino, S., Baker, T. R., Nardoto, G. B., Martinelli, L. A., Quesada, C. A., Paiva, R.,
11 Schwarz, M., Horna, V., Mercado, L. M., Santos, A., Arroyo, L., Jimenez, E. M., Luizao,
12 F. J., Neill, D. A., Silva, N., Prieto, A., Rudas, A., Silviera, M., Vieira, I. C. G., Lopez-
13 Gonzalez, G., Malhi, Y., Phillips, O. L., and Lloyd, J.: Basin-wide variations in foliar
14 properties of Amazonian forest: phylogeny, soils and climate, *Biogeosciences*, 6, 2677-
15 2708, 2009.

16 Fyllas, N. M., Quesada, C. A., and Lloyd, J.: Deriving Plant Functional Types for Amazonian
17 forests for use in vegetation dynamics models, *Perspectives in Plant Ecology Evolution
18 and Systematics*, 14, 97-110, 2012.

19 Gleason, S. M., Read, J., Ares, A., and Metcalfe, D. J.: Phosphorus economics of tropical
20 rainforest species and stands across soil contrasts in Queensland, Australia:
21 understanding the effects of soil specialization and trait plasticity, *Functional Ecology*,
22 23, 1157-1166, 2009.

23 Gloor, M., Gatti, L., Brienen, R., Feldpausch, T. R., Phillips, O. L., Miller, J., Ometto, J. P.,
24 Rocha, H., Baker, T., de Jong, B., Houghton, R. A., Malhi, Y., Aragao, L., Guyot, J. L.,
25 Zhao, K., Jackson, R., Peylin, P., Sitch, S., Poulter, B., Lomas, M., Zaehle, S.,
26 Huntingford, C., Levy, P., and Lloyd, J.: The carbon balance of South America: a review
27 of the status, decadal trends and main determinants, *Biogeosciences*, 9, 5407-5430, 2012.

28 Harrison, M. T., Edwards, E. J., Farquhar, G. D., Nicotra, A. B., and Evans, J. R.: Nitrogen in
29 cell walls of sclerophyllous leaves accounts for little of the variation in photosynthetic
30 nitrogen-use efficiency, *Plant Cell Environ.*, 32, 259-270, 2009.

31 Hikosaka, K.: Interspecific difference in the photosynthesis-nitrogen relationship: patterns,
32 physiological causes, and ecological importance, *Journal of Plant Research*, 117, 481-494,
33 2004.

34 Hikosaka, K. and Shigeno, A.: The role of Rubisco and cell walls in the interspecific variation in
35 photosynthetic capacity, *Oecologia*, 160, 443-451, 2009.

36 Hoffmann, W. A., Adasme, R., Haridasan, M., de Carvalho, M. T., Geiger, E. L., Pereira, M. A.
37 B., Gotsch, S. G., and Franco, A. C.: Tree topkill, not mortality, governs the dynamics of
38 savanna-forest boundaries under frequent fire in central Brazil, *Ecology*, 90, 1326-1337,
39 2009.

40 Hoffmann, W. A., Geiger, E. L., Gotsch, S. G., Rossatto, D. R., Silva, L. C. R., Lau, O. L.,
41 Haridasan, M., and Franco, A. C.: Ecological thresholds at the savanna-forest boundary:
42 how plant traits, resources and fire govern the distribution of tropical biomes, *Ecology
43 Letters*, 15, 759-768, 2012.

44 Hothorn, T., Hornik, K., van de Wiel, M. A. V., and Zeileis, A.: Implementing a Class of
45 Permutation Tests: The coin Package, *Journal of Statistical Software*, 28, 1-23, 2008.

46 Huntingford, C., Zelazowski, P., Galbraith, D., Mercado, L. M., Sitch, S., Fisher, R., Lomas, M.,
47 Walker, A. P., Jones, C. D., Booth, B. B. B., Malhi, Y., Hemming, D., Kay, G., Good, P.,
48 Lewis, S. L., Phillips, O. L., Atkin, O. K., Lloyd, J., Gloor, E., Zaragoza-Castells, J., Meir,
49 P., Betts, R., Harris, P. P., Nobre, C., Marengo, J., and Cox, P. M.: Simulated resilience of
50 tropical rainforests to CO₂-induced climate change, *Nature Geoscience*, 6, 268-273,
51 2013.

- 1 Kattge, J., Knorr, W., Raddatz, T., and Wirth, C.: Quantifying photosynthetic capacity and its
2 relationship to leaf nitrogen content for global-scale terrestrial biosphere models, *Glob.*
3 *Change Biol.*, 15, 976-991, 2009.
- 4 Larcher, W.: Effect of environmental and physiological variables on carbon dioxide gas exchange
5 of trees, *Photosynthetica*, 3, 167-&, 1969.
- 6 Legendre, P. and Legendre, L.: Numerical ecology, Elsevier, Amsterdam, 2012.
- 7 Lehmann, C. E. R., Archibald, S. A., Hoffmann, W. A., and Bond, W. J.: Deciphering the
8 distribution of the savanna biome, *New Phytologist*, 191, 197-209, 2011.
- 9 Lloyd, J., Bird, M. I., Vellen, L., Miranda, A. C., Veenendaal, E. M., Djangbletey, G., Miranda, H.
10 S., Cook, G., and Farquhar, G. D.: Contributions of woody and herbaceous vegetation to
11 tropical savanna ecosystem productivity: a quasi-global estimate, *Tree Physiol.*, 28, 451-
12 468, 2008.
- 13 Lloyd, J., Bloomfield, K., Domingues, T. F., and Farquhar, G. D.: Photosynthetically relevant
14 foliar traits correlating better on a mass vs an area basis: of ecophysiological relevance or
15 just a case of mathematical imperatives and statistical quicksand?, *New Phytologist*, 199,
16 311-321, 2013.
- 17 Lloyd, J., Goulden, M., Ometto, J. P., Fyllas, N. M., Quesada, C. A., and Patiño, S.:
18 Ecophysiology of forest and savanna vegetation. In: Amazonia and Global Change,
19 Keller, M., Gash, J., and Silva Dias, P. (Eds.), American Geophysical Union, Washington,
20 D.C., 2009.
- 21 Lloyd, J., Patiño, S., Paiva, R. Q., Nardoto, G. B., Quesada, C. A., Santos, A. J. B., Baker, T. R.,
22 Brand, W. A., Hilke, I., Gielmann, H., Raessler, M., Luizao, F. J., Martinelli, L. A., and
23 Mercado, L. M.: Optimisation of photosynthetic carbon gain and within-canopy
24 gradients of associated foliar traits for Amazon forest trees, *Biogeosciences*, 7, 1833-
25 1859, 2010.
- 26 Malhi, Y., Aragao, L., Galbraith, D., Huntingford, C., Fisher, R., Zelazowski, P., Sitch, S.,
27 McSweeney, C., and Meir, P.: Exploring the likelihood and mechanism of a climate-
28 change-induced dieback of the Amazon rainforest, *Proceedings of the National Academy*
29 *of Sciences of the United States of America*, 106, 20610-20615, 2009.
- 30 Maximov, N. A.: The plant in relation to water: a study of the physiological basis of drought
31 resistance, George Allen and Unwin, London, 1929.
- 32 McKenzie, N. J., Jacquier, D., Isbell, R. F., and Brown, K.: Australian soils and landscapes: an
33 illustrated compendium, CSIRO Publishing, Collingwood, Victoria, 2004.
- 34 Medina, E.: Nutrient balance and physiological processes at the leaf level. In: Physiological
35 ecology of plants of the wet tropics, Medina, E., Mooney, H. A., and Vázquez-Yanes, C.
36 (Eds.), Dr W. Junk Publishers, The Hague, 1984.
- 37 Meir, P., Levy, P. E., Grace, J., and Jarvis, P. G.: Photosynthetic parameters from two
38 contrasting woody vegetation types in West Africa, *Plant Ecology*, 192, 277-287, 2007.
- 39 Mercado, L. M., Patiño, S., Domingues, T. F., Fyllas, N. M., Weedon, G. P., Sitch, S., Quesada,
40 C. A., Phillips, O. L., Aragao, L., Malhi, Y., Dolman, A. J., Restrepo-Coupe, N., Saleska,
41 S. R., Baker, T. R., Almeida, S., Higuchi, N., and Lloyd, J.: Variations in Amazon forest
42 productivity correlated with foliar nutrients and modelled rates of photosynthetic carbon
43 supply, *Philosophical Transactions of the Royal Society B-Biological Sciences*, 366, 3316-
44 3329, 2011.
- 45 Mott, K. A., Gibson, A. C., and Oleary, J. W.: The adaptive significance of amphistomatic leaves,
46 *Plant Cell Environ.*, 5, 455-460, 1982.
- 47 Niinemets, U.: Role of foliar nitrogen in light harvesting and shade tolerance of four temperate
48 deciduous woody species, *Functional Ecology*, 11, 518-531, 1997.
- 49 Ohta, S. and Effendi, S.: Ultisols of "lowland *Dipterocarp* forest" in East Kalimantan, Indonesia
50 II. Status of carbon, nitrogen and phosphorus, *Soil Science and Plant Nutrition*, 38, 207-
51 216, 1992.

- 1 Parkhurst, D. F.: The adaptive significance of stomatal occurrence on one or both surfaces of
2 leaves, *Journal of Ecology*, 66, 367-383, 1978.
- 3 Poorter, L.: Growth responses of 15 rain-forest tree species to a light gradient: the relative
4 importance of morphological and physiological traits, *Functional Ecology*, 13, 396-410,
5 1999.
- 6 Popma, J., Bongers, F., and Werger, M. J. A.: Gap-dependence and leaf characteristics of trees in
7 a tropical lowland rain-forest in Mexico, *Oikos*, 63, 207-214, 1992.
- 8 Prior, L. D., Bowman, D., and Eamus, D.: Intra-specific variation in leaf attributes of four
9 savanna tree species across a rainfall gradient in tropical Australia, *Australian Journal of*
10 *Botany*, 53, 323-335, 2005.
- 11 Pyykko, M.: The leaf anatomy of East Patagonian xerophytic plants, *Ann. bot. fennici*, 3, 453-
12 622, 1966.
- 13 Quesada, C. A., Lloyd, J., Schwarz, M., Patino, S., Baker, T. R., Czimczik, C., Fyllas, N. M.,
14 Martinelli, L., Nardoto, G. B., Schmerler, J., Santos, A. J. B., Hodnett, M. G., Herrera, R.,
15 Luizao, F. J., Arneith, A., Lloyd, G., Dezzee, N., Hilke, I., Kuhlmann, I., Raessler, M.,
16 Brand, W. A., Geilmann, H., Moraes, J. O., Carvalho, F. P., Araujo, R. N., Chaves, J. E.,
17 Cruz, O. F., Pimentel, T. P., and Paiva, R.: Variations in chemical and physical properties
18 of Amazon forest soils in relation to their genesis, *Biogeosciences*, 7, 1515-1541, 2010.
- 19 Quesada, C. A., Phillips, O. L., Schwarz, M., Czimczik, C. I., Baker, T. R., Patino, S., Fyllas, N.
20 M., Hodnett, M. G., Herrera, R., Almeida, S., Davila, E. A., Arneith, A., Arroyo, L., Chao,
21 K. J., Dezzee, N., Erwin, T., di Fiore, A., Higuchi, N., Coronado, E. H., Jimenez, E. M.,
22 Killeen, T., Lezama, A. T., Lloyd, G., Lopez-Gonzalez, G., Luizao, F. J., Malhi, Y.,
23 Monteagudo, A., Neill, D. A., Vargas, P. N., Paiva, R., Peacock, J., Penuela, M. C., Cruz,
24 A. P., Pitman, N., Priante, N., Prieto, A., Ramirez, H., Rudas, A., Salomao, R., Santos, A.
25 J. B., Schmerler, J., Silva, N., Silveira, M., Vasquez, R., Vieira, I., Terborgh, J., and Lloyd,
26 J.: Basin-wide variations in Amazon forest structure and function are mediated by both
27 soils and climate, *Biogeosciences*, 9, 2203-2246, 2012.
- 28 R Development Core Team: R: A language and environment for statistical computing. R
29 Foundation for Statistical Computing, Vienna, Austria, 2011.
- 30 Raaijmakers, D., Boot, R. G. A., Dijkstra, P., Pot, S., and Pons, T.: Photosynthetic rates in
31 relation to leaf phosphorus content in pioneer versus climax tropical rainforest trees,
32 *Oecologia*, 102, 120-125, 1995.
- 33 Ratter, J. A.: Transition between Cerrado and forest vegetation in Brazil. In: *Nature and*
34 *dynamics of forest-savanna boundaries*, Furley, P., Proctor, J., and Ratter, J. A. (Eds.),
35 Chapman and Hall, London, 1992.
- 36 Reich, P. B., Oleksyn, J., and Wright, I. J.: Leaf phosphorus influences the photosynthesis-
37 nitrogen relation: a cross-biome analysis of 314 species, *Oecologia*, 160, 207-212, 2009.
- 38 Russell-Smith, J., Stanton, P. J., Whitehead, P. J., and Edwards, A.: Rain forest invasion of
39 eucalypt-dominated woodland savanna, iron range, north-eastern Australia: I.
40 Successional processes, *Journal of Biogeography*, 31, 1293-1303, 2004.
- 41 Saiz, G., Bird, M. I., Domingues, T., Schrodt, F., Schwarz, M., Feldpausch, T. R., Veenendaal, E.,
42 Djagbletey, G., Hien, F., Compaore, H., Diallo, A., and Lloyd, J.: Variation in soil carbon
43 stocks and their determinants across a precipitation gradient in West Africa, *Glob.*
44 *Change Biol.*, 18, 1670-1683, 2012.
- 45 Santiago, L. S. and Mulkey, S. S.: A test of gas exchange measurements on excised canopy
46 branches of ten tropical tree species, *Photosynthetica*, 41, 343-347, 2003.
- 47 Schimper, A. F. W.: *Plant-geography upon a Physiological Basis*, Clarendon Press, Oxford,
48 1903.
- 49 Schrodt, F., Domingues, T. F., Feldpausch, T. R., Saiz, G., Quesada, C. A., Schwarz, M., Ishida,
50 F. Y., Compaore, H., Diallo, A., Djagbletey, G., Hien, F., Hiernaux, P., Mougou, E.,
51 Sonké, B., Zapfack, L., Bird, M., Lewis, S. L., Meir, P., Phillips, O. L., Grace, J.,

- 1 Veenendaal, E. M., and Lloyd, J.: Foliar trait contrasts between African forest and
2 savanna trees: Genetic versus environmental effects, *Functional Plant Biology*, In
3 submission, 2014.
- 4 Sitch, S., Huntingford, C., Gedney, N., Levy, P. E., Lomas, M., Piao, S. L., Betts, R., Ciais, P.,
5 Cox, P., Friedlingstein, P., Jones, C. D., Prentice, I. C., and Woodward, F. I.: Evaluation
6 of the terrestrial carbon cycle, future plant geography and climate-carbon cycle feedbacks
7 using five Dynamic Global Vegetation Models (DGVMs), *Glob. Change Biol.*, 14, 2015-
8 2039, 2008.
- 9 Sniderman, J. M. K. and Jordan, G. J.: Extent and timing of floristic exchange between
10 Australian and Asian rain forests, *Journal of Biogeography*, 38, 1445-1455, 2011.
- 11 Spain, A. V.: Influence of environmental conditions and some soil chemical properties on the
12 carbon and nitrogen contents of some tropical Australian rainforest soils, *Australian*
13 *Journal of Soil Research*, 28, 825-839, 1990.
- 14 Swaine, M. D. and Whitmore, T. C.: On the definition of ecological species groups in tropical
15 rain forests, *Vegetatio*, 75, 81-86, 1988.
- 16 Takashima, T., Hikosaka, K., and Hirose, T.: Photosynthesis or persistence: nitrogen allocation
17 in leaves of evergreen and deciduous *Quercus* species, *Plant Cell Environ.*, 27, 1047-
18 1054, 2004.
- 19 Tessins, E. and Jusop, S.: Quantitative relationships between mineralogy and properties of
20 tropical soils, Universiti Pertanian Malaysia, Serdang, Selangor, 1983.
- 21 Thompson, J., Viana, J., Proctor, J., and Ratter, J. A.: Contrasting forest-savanna boundaries on
22 Maraca Island, Roraima, Brazil. In: *Nature and dynamics of forest-savanna boundaries*,
23 Furley, P., Proctor, J., and Ratter, J. A. (Eds.), Chapman and Hall, London, 1992.
- 24 Torello-Raventos, M., Feldpausch, T. R., Veenendaal, E., Schrodte, F., Saiz, G., Domingues, T. F.,
25 Djagbletey, G., Ford, A., Kemp, J., Marimon, B. S., Marimon, B. H., Lenza, E., Ratter, J.
26 A., Maracahipes, L., Sasaki, D., Sonke, B., Zapfack, L., Taedoumg, H., Villarroel, D.,
27 Schwarz, M., Quesada, C. A., Ishida, F. Y., Nardoto, G. B., Affum-Baffoe, K., Arroyo,
28 L., Bowman, D., Compaore, H., Davies, K., Diallo, A., Fyllas, N. M., Gilpin, M., Hien,
29 F., Johnson, M., Killeen, T. J., Metcalfe, D., Miranda, H. S., Steininger, M., Thomson, J.,
30 Sykora, K., Mougouin, E., Hiernaux, P., Bird, M. I., Grace, J., Lewis, S. L., Phillips, O. L.,
31 and Lloyd, J.: On the delineation of tropical vegetation types with an emphasis on
32 forest/savanna transitions, *Plant Ecology & Diversity*, 6, 101-137, 2013.
- 33 Townsend, A. R., Cleveland, C. C., Asner, G. P., and Bustamante, M. M. C.: Controls over foliar
34 N : P ratios in tropical rain forests, *Ecology*, 88, 107-118, 2007.
- 35 Turner, I. M.: *The ecology of trees in the tropical rain forest*, Cambridge University Press,
36 Cambridge, 2001.
- 37 Veenendaal, E. M., Swaine, M. D., Lecha, R. T., Walsh, M. F., Abebrese, I. K., and
38 OwusuAfrিয়ে, K.: Responses of West African forest tree seedlings to irradiance and soil
39 fertility, *Functional Ecology*, 10, 501-511, 1996.
- 40 Veenendaal, E. M., Torello-Raventos, M., Feldpausch, T. R., Domingues, T. F., Gerard, F.,
41 Schrodte, F., Saiz, G., Quesada, C. A., Djagbletey, G., Ford, A., Kemp, J., Marimon, B. S.,
42 Marimon-Junior, B. H., Lenza, E., Ratter, J. A., Maracahipes, L., Sasaki, D., Sonké, B.,
43 Zapfack, L., Villarroel, D., Schwarz, M., Yoko Ishida, F., Gilpin, M., Nardoto, G. B.,
44 Affum-Baffoe, K., Arroyo, L., Bloomfield, K., Ceca, G., Compaore, H., Davies, K.,
45 Diallo, A., Fyllas, N. M., Gignoux, J., Hien, F., Johnson, M., Mougouin, E., Hiernaux, P.,
46 Killeen, T., Metcalfe, D., Miranda, H. S., Steininger, M., Sykora, K., Bird, M. I., Grace, J.,
47 Lewis, S., Phillips, O. L., and Lloyd, J.: Structural, physiognomic and aboveground
48 biomass variation in savanna-forest transition zones on three continents. How different
49 are co-occurring savanna and forest formations?, *Biogeosciences Discuss.*, 11, 4591-
50 4636, 2014.

- 1 Vitousek, P. M.: Litterfall, nutrient cycling and nutrient limitation in tropical forests, *Ecology*, 65,
2 285-298, 1984.
- 3 Warren, C. R.: Stand aside stomata, another actor deserves centre stage: the forgotten role of the
4 internal conductance to CO₂ transfer, *Journal of Experimental Botany*, 59, 1475-1487,
5 2008.
- 6 Warton, D. I., Wright, I. J., Falster, D. S., and Westoby, M.: Bivariate line-fitting methods for
7 allometry, *Biological Reviews*, 81, 259-291, 2006.
- 8 Webb, L. J.: Environmental relationships of the structural types of Australian rain forest
9 vegetation, *Ecology*, 49, 296-&, 1968.
- 10 Westoby, M. and Wright, I. J.: Land-plant ecology on the basis of functional traits, *Trends in*
11 *Ecology & Evolution*, 21, 261-268, 2006.
- 12 Whitehead, P. W., Stephenson, P. J., McDougall, I., Hopkins, M. S., Grahams, A. W., Collerson,
13 K. D., and Johnson, D. P.: Temporal development of the Atherton Basalt Province,
14 north Queensland, *Australian Journal of Earth Sciences*, 54, 691-709, 2007.
- 15 Wright, I. J., Reich, P. B., and Westoby, M.: Strategy shifts in leaf physiology, structure and
16 nutrient content between species of high- and low-rainfall and high- and low-nutrient
17 habitats, *Functional Ecology*, 15, 423-434, 2001.
- 18 Wright, I. J., Reich, P. B., Westoby, M., Ackerly, D. D., Baruch, Z., Bongers, F., Cavender-Bares,
19 J., Chapin, T., Cornelissen, J. H. C., Diemer, M., Flexas, J., Garnier, E., Groom, P. K.,
20 Gulias, J., Hikosaka, K., Lamont, B. B., Lee, T., Lee, W., Lusk, C., Midgley, J. J., Navas,
21 M. L., Niinemets, U., Oleksyn, J., Osada, N., Poorter, H., Poot, P., Prior, L., Pyankov, V.
22 I., Roumet, C., Thomas, S. C., Tjoelker, M. G., Veneklaas, E. J., and Villar, R.: The
23 worldwide leaf economics spectrum, *Nature*, 428, 821-827, 2004.
- 24 Zuur, A. F., Ieno, E. N., Walker, N. J., Saveliev, A. A., and Smith, G. M.: Mixed effects models
25 and extensions in ecology with R, Springer Science, New York, USA, 2009.

26

27

Table 1. Plot coordinates after Torello-Raventos et al. (2013), vegetation classification \mathcal{V} , elevation above sea level E_V , mean annual temperature T_A , mean annual precipitation P_A , 0.95 quantile upper stratum canopy height H^* , upper stratum canopy area index C_U , soil pH, soil exchangeable cations, soil extractable phosphorus and Wold Reference Base (WRB) soil classification for the study sites. Soil values represent the top 0.3 m of soil.

| Plot | Lat. | Long. | \mathcal{V} | E_V (m) | T_A (°C) | P_A (m) | H^* (m) | C_U (m ² m ⁻²) | pH | [Al] _e | [Ca] _{ex} | [K] _{ex} | [Mg] _{ex} | [Na] _{ex} | [P] _{ex} | WRB Soil Classification |
|--------|---------|----------|--------------------------|--------------|---------------|--------------|--------------|--|------|-------------------------------------|--------------------|-------------------|--------------------|--------------------|-----------------------|--|
| | | | | | | | | | | mmol _{eq} kg ⁻¹ | | | | | (μg g ⁻¹) | |
| CTC-01 | 16.103S | 145.447E | Tall forest | 90 | 25.2 | 3.20 | 38.9 | 2.36 | 5.56 | 0.48 | 17.94 | 0.71 | 7.65 | 0.65 | 208 | Haplic Cambisol (Hyperdystric, Alumatic, Skeletic) |
| KBL-01 | 17.764S | 145.544E | Tall forest | 761 | 20.5 | 1.75 | 38.0 | 1.45 | 4.79 | 0.27 | 4.71 | 0.35 | 2.58 | 0.36 | 952 | Haplic Regosol (Siltic, Hyperdystric) |
| KBL-03 | 17.685S | 145.535E | Tall forest | 1055 | 19.1 | 1.34 | 35.8 | 2.30 | 4.38 | 4.84 | 0.90 | 0.47 | 1.82 | 0.62 | 227 | Haplic Nitisol (Hyperdystric, Rhodic) |
| KCR-01 | 17.107S | 145.604E | Tall forest | 813 | 20.5 | 1.96 | 44.0 | 2.21 | 5.40 | 0.60 | 16.11 | 0.78 | 5.71 | 0.39 | 165 | Haplic Cambisol (Dystric, Alumatic) |
| DCR-01 | 17.026S | 145.597E | Tall savanna woodland | 683 | 21.2 | 1.45 | 26.2 | 1.63 | 5.65 | 0.90 | 8.78 | 0.71 | 6.93 | 0.71 | 79 | Haplic Cambisol (Orthodystric, Alumatic) |
| DCR-02 | 17.021S | 145.584E | Tall savanna woodland | 653 | 21.3 | 1.46 | 22.1 | 0.70 | 5.52 | 1.27 | 8.82 | 0.55 | 5.74 | 0.33 | 56 | Arenic Cambisol (Epieutric) |
| KBL-02 | 17.849S | 145.532E | Tall savanna woodland | 860 | 20.1 | 1.43 | 28.1 | 0.77 | 5.28 | 0.26 | 4.78 | 0.17 | 7.66 | 0.77 | 216 | Geric Acrisol (Hyperdystric, Rhodic) |

Table 2: Coefficients for standardised major axis (SMA) bivariate relationships. Vegetation contrasts: forest and savanna; n, number of observations; r^2 , correlation coefficient and associated p -value; Intercept; Slope and 95% confidence interval.

| Response | Bivariate | Vegetation class | n | r^2 | p | Intercept | Slope | Slope: | |
|-----------------------|------------------|------------------|-----|-------|---------|-----------|--------|------------|-------------|
| | | | | | | | | Low 95% CI | High 95% CI |
| [P] _a ↔ | [N] _a | Forest | 81 | 0.82 | <0.0001 | -11.01 | 47.92 | 43.58 | 52.68 |
| | | Savanna | 24 | 0.51 | <0.0001 | -70.96 | 75.86 | 55.87 | 102.99 |
| [P] _m ↔ | [N] _m | All | 105 | 0.81 | <0.0001 | -0.097 | 0.046 | 0.042 | 0.050 |
| [N] _a ↔ | M_a | Forest | 81 | 0.43 | <0.0001 | 0.299 | 0.018 | 0.015 | 0.021 |
| | | Savanna | 24 | 0.08 | 0.1848 | | | | |
| [P] _a ↔ | M_a | Forest | 81 | 0.38 | <0.0001 | 3.31 | 0.84 | 0.71 | 1.00 |
| | | Savanna | 24 | 0.00 | 0.8207 | | | | |
| $A_{max,a}$ ↔ | [N] _a | Forest | 81 | 0.47 | <0.0001 | -11.43 | 17.01 | 14.46 | 20.00 |
| | | Savanna | 24 | 0.17 | 0.0442 | -54.48 | 42.15 | 28.47 | 62.40 |
| $A_{max,a}$ ↔ | [P] _a | Forest | 81 | 0.47 | <0.0001 | -7.52 | 0.35 | 0.30 | 0.42 |
| | | Savanna | 24 | 0.09 | 0.1522 | | | | |
| $A_{max,a}$ ↔ | M_a | All | 105 | 0.30 | <0.0001 | -6.03 | 0.29 | 0.25 | 0.34 |
| $A_{max,m}$ ↔ | [N] _m | Forest | 81 | 0.63 | <0.0001 | -192.81 | 18.47 | 16.12 | 21.16 |
| | | Savanna | 24 | 0.31 | 0.0049 | -100.47 | 21.87 | 15.26 | 31.35 |
| $A_{max,m}$ ↔ | [P] _m | Forest | 81 | 0.61 | <0.0001 | -148.85 | 404.67 | 352.06 | 465.15 |
| | | Savanna | 24 | 0.30 | 0.0061 | -14.60 | 394.68 | 274.50 | 567.47 |
| $A_{max,m}$ ↔ | M_a | Forest | 81 | 0.08 | 0.0132 | 607.15 | -4.95 | -6.13 | -4.00 |
| | | Savanna | 24 | 0.30 | 0.0056 | 628.00 | -2.73 | -3.93 | -1.90 |
| DM:H ₂ O ↔ | M_a | Forest | 85 | 0.40 | <0.0001 | -0.063 | 0.009 | 0.007 | 0.010 |
| | | Savanna | 24 | 0.30 | 0.0056 | 0.086 | 0.005 | 0.003 | 0.007 |
| d ↔ | M_a | Forest | 82 | 0.26 | <0.0001 | 107.77 | 3.91 | 3.23 | 4.72 |
| | | Savanna | 24 | 0.74 | <0.0001 | 152.77 | 1.86 | 1.48 | 2.32 |
| ρ ↔ | M_a | Forest | 82 | 0.45 | <0.0001 | 0.040 | 0.0020 | 0.0017 | 0.0024 |
| | | Savanna | 24 | 0.66 | <0.0001 | 0.141 | 0.0014 | 0.0011 | 0.0018 |
| A_N ↔ | M_a | Forest | 81 | 0.00 | 0.6122 | | | | |
| | | Savanna | 24 | 0.00 | 0.9268 | | | | |

Table 3. Stepwise selection process for the fixed component of the linear mixed effect model: photosynthetic capacity ($\mathcal{A}_{\max,a}$) as response variable. Categorical variable B has two levels: forest and savanna for the contrasting vegetation types. Continuous explanatory variables are $[N]_a$, $[P]_a$, leaf mass per unit area (M_a), leaf dry matter content (ξ) and leaf thickness (d). The effect of dropping sequential terms was tested by comparing the nested model variants. Model variants were all run using the Maximum Likelihood method; the models' random component was identical in all variants. Test parameters and statistics are df, degrees of freedom; AIC, Akaike information criteria; BIC, Bayesian information criteria; logLik, Maximum Likelihood; the Likelihood Ratio statistic and associated p -value. Models with the same degrees of freedom are not nested one in the other.

| Model | Fixed component | df | AIC | BIC | logLik | Test | L.Ratio | p-value |
|-------|--|----|--------|--------|---------|---------|---------|--------------|
| 1 | B + $[N]_a$ + $[P]_a$ + M_a + d + ξ | 11 | 735.51 | 765.12 | -356.76 | | | |
| 2 | B + $[N]_a$ + $[P]_a$ + M_a + d | 10 | 733.52 | 760.44 | -356.76 | 1 vs 2 | 0.011 | 0.915 |
| 3 | B + $[N]_a$ + $[P]_a$ + M_a | 9 | 731.94 | 756.16 | -356.97 | 2 vs 3 | 0.417 | 0.519 |
| 4 | B + $[N]_a$ + $[P]_a$ | 8 | 730.58 | 752.11 | -357.29 | 3 vs 4 | 0.640 | 0.424 |
| 5 | B + $[N]_a$ | 7 | 732.96 | 751.80 | -359.48 | 4 vs 5 | 4.383 | 0.036 |
| 6 | $[N]_a$ + $[P]_a$ | 7 | 738.39 | 757.23 | -362.20 | 4 vs 6 | 9.814 | 0.002 |
| 7 | $[N]_a$ + $[P]_a$ + M_a + d + ξ | 10 | 742.94 | 769.86 | -361.47 | 1 vs 7 | 9.432 | 0.002 |
| 8 | B + $[N]_a$ + $[P]_a$ + B: $[N]_a$ + B: $[P]_a$ + $[N]_a$: $[P]_a$ + B: $[N]_a$: $[P]_a$ | 12 | 734.73 | 767.02 | -355.36 | 4 vs 8 | 3.852 | 0.426 |
| 9 | B + $[N]_a$ + $[P]_a$ + B: $[N]_a$ + B: $[P]_a$ | 10 | 733.15 | 760.07 | -356.58 | 7 vs 9 | 2.426 | 0.297 |
| 10 | B + $[N]_a$ + $[P]_a$ + B: $[N]_a$ | 9 | 731.37 | 755.59 | -356.68 | 8 vs 10 | 0.213 | 0.645 |
| 11 | B + $[N]_a$ + $[P]_a$ + B: $[P]_a$ | 9 | 731.49 | 755.71 | -356.75 | 8 vs 11 | 0.339 | 0.561 |
| 12 | B + $[N]_a$ + $[P]_a$ + $[N]_a$: $[P]_a$ | 9 | 732.53 | 756.75 | -357.26 | 4 vs 12 | 0.050 | 0.823 |

Table 4. Output of the linear mixed effects model (Eq. 2): fixed effects (top) and random effects (bottom). The top section shows fixed effect parameter estimates and associated standard error, degrees of freedom, test statistic and associated p -value. The final ‘optimal’ model is compared against a simpler ‘null’ model that includes only vegetation type (B) in the fixed component, but has an identical random term of Species nested within Site. Continuous explanatory variables were centred on their respective means (i.e. zero re-set to the trait average).

| Fixed effect | Final model | | | | | Null model | | | | | |
|---|---------------------|-----------------|-------------------|---------|-----------|--------------|-----------------|-------------------|---------|-----------|---------|
| | Estimate | S.E. | DF | t value | p value | Estimate | S.E. | DF | t value | p value | |
| Forest (if other variables were zero) | | 18.06 | 1.57 | 73 | 11.537 | <0.0001 | 17.08 | 2.47 | 75 | 6.925 | <0.0001 |
| Savanna (Vegetation contrast) | β_1 | 11.18 | 3.08 | 5 | 2.076 | 0.0151 | 14.15 | 4.54 | 5 | 3.119 | 0.0263 |
| Leaf[N] _a | β_2 | 6.66 | 2.07 | 73 | 3.472 | 0.0020 | | | | | |
| Leaf[P] _a | β_3 | 0.07 | 0.03 | 73 | 1.618 | 0.0502 | | | | | |
| Random effect | | Variance | % of total | | | | Variance | % of total | | | |
| Intercept variance: Among sites | a_k | 3.56 | 9.4% | | | 12.06 | 13.9% | | | | |
| Intercept variance: Among species | $a_j k$ | 26.77 | 70.8% | | | 66.39 | 76.2% | | | | |
| Residual (within species, within sites) | ε_{ijk} | 7.49 | 19.8% | | | 8.64 | 9.9% | | | | |
| | | 37.82 | 100.0% | | | 87.10 | 100.0% | | | | |
| AIC | | | | 726.7 | | | | | 746.5 | | |
| Likelihood ratio test | | | | -355.4 | | | | | -367.2 | | |

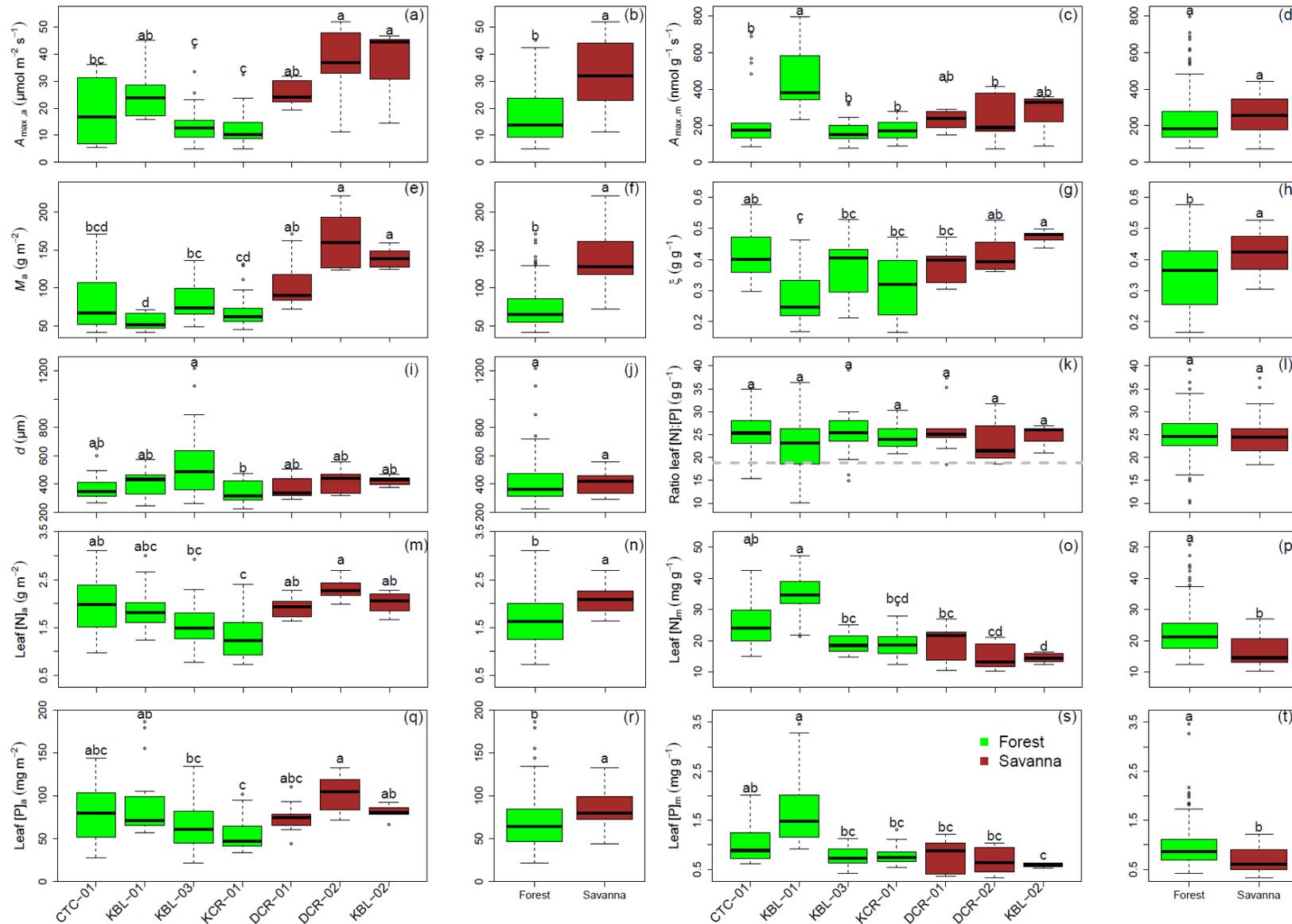


Figure 1 Paired boxplots of key leaf traits (untransformed data) by site and vegetation type (\mathcal{V}). The two \mathcal{V} classes are distinguished by colour: green for forest and brown for savanna. Site abbreviations are laid out in Table 1. Leaf traits are photosynthetic capacity (a,b) per unit leaf area and (c,d) per unit leaf dried mass; (e,f) leaf mass per unit area; (g,h) leaf dried matter content; (i,j) leaf thickness; (k,l) ratio of total leaf nitrogen to phosphorus; total leaf nitrogen (m,n) per unit leaf area and (o,p) per unit leaf dried mass; total leaf phosphorus (q,r) per unit leaf area and (s,t) per unit leaf dried mass. The box and whiskers show the median result as a thick horizontal band, the ends of the box denote the inter-quartile range; the whiskers extend 1.5 times the inter-quartile range or to the most extreme value, whichever is smaller; any points outside these values are shown as outliers. The grey dashed line in plot k represents a mean N:P ratio of 18.8 reported for tropical forests by Reich et al. (2009).

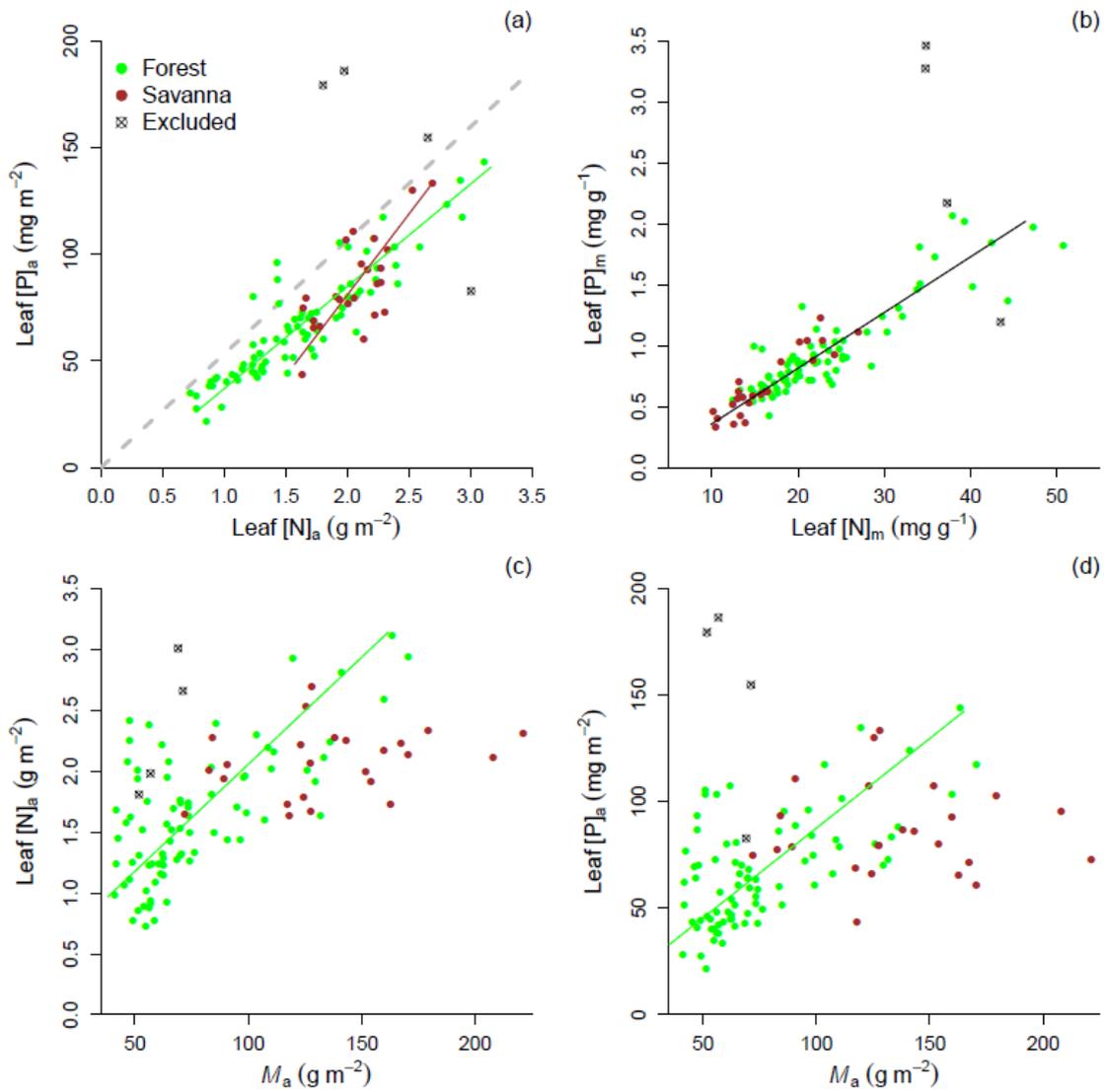


Figure 2. Scatterplots of the relationships between leaf phosphorus and leaf nitrogen (a) on an area basis and (b) on a mass basis. Plot (c) shows the relationship between leaf nitrogen on an area basis and leaf mass per unit area; plot (d) shows the equivalent relationship for leaf phosphorus. Each point corresponds to a single tree and vegetation types are distinguished by colour: green for \mathcal{F} and brown for \mathcal{S} . Standardised major axis (SMA) fitted lines are shown for the two vegetation types only where the bivariate relationship proved significant $p < 0.05$. Pearson's correlations testing the assumption of linearity are given at Table S3 (Supplementary Information) together with Likelihood ratio and Wald statistics testing the H_0 of common slope, elevation and axis shift for the two \mathcal{V} classes. Intercept, slope and r^2 values for the SMA fitted lines are given in Table 2. In plot (a) a third fitted line (grey, dashed) displays a slope based on a mean N:P ratio of 18.8 typical for tropical forests as reported by Reich et al. (2009) and passing through the origin. In plot (b) there was a common slope to the $[P]_m \leftrightarrow [N]_m$ relationship for both vegetation types (black line). In reviewing residual plots of initial SMA fits (not shown), four data points were identified as outliers (crossed circles). The four outliers have been excluded from the re-run SMA fits shown here.

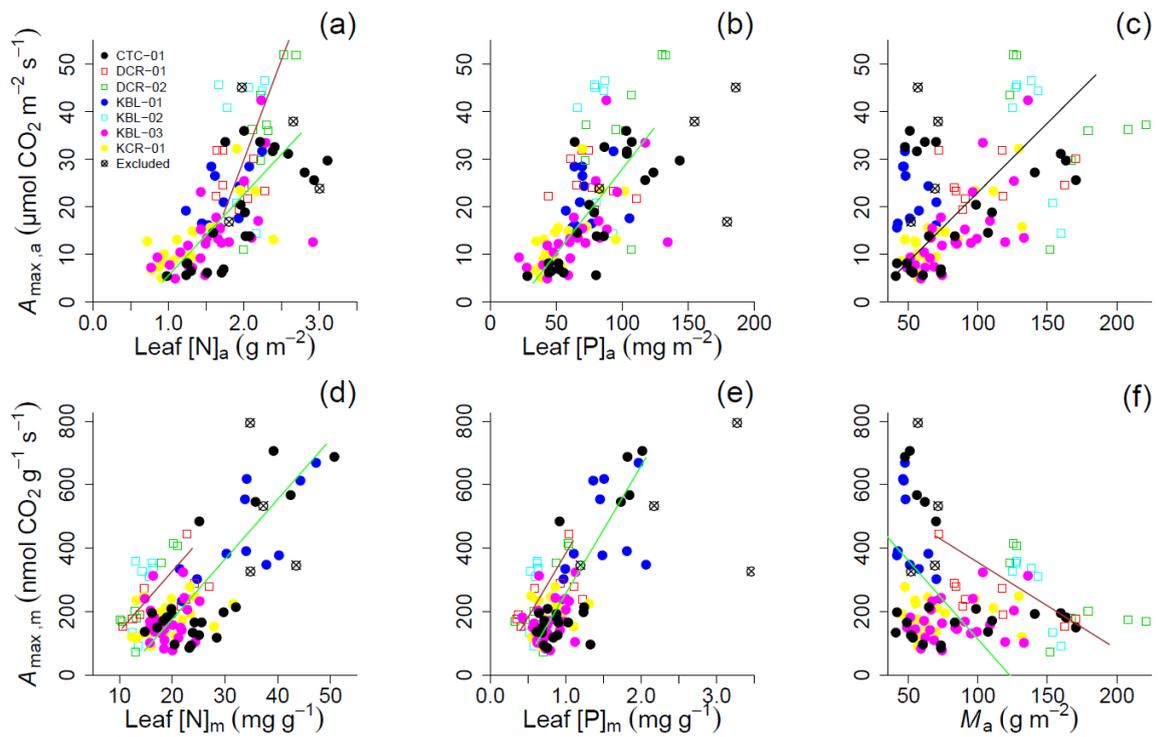


Figure 3. Top panel (plots a, b and c), scatterplots of the area-based relationships between photosynthetic capacity ($A_{\max,a}$) and (a) leaf nitrogen; (b) leaf phosphorus; (c) leaf mass per unit area (M_a). Bottom panel (plots d, e and f) the equivalent plots are expressed per unit leaf dried mass. Each point represents a single leaf; dots denote *V* sites and square symbols denote *S* sites; individual sites are distinguished by colour: CTC-01 black, DCR-01 red, DCR-02 green, KBL-01 royal blue, KBL-02 turquoise, KBL-03 pink, KCR-01 yellow. Standardised major axis (SMA) fitted lines are shown where significant: *V* (green), *S* (brown). Pearson's correlations testing the assumption of linearity are given at Table S3 (Supplementary Information) together with Likelihood ratio and Wald statistics testing the H_0 of common slope, elevation and axis shift for the two *V* classes. In plot (c) there was no difference in slope between the two *V* and so a common line is fitted (black). SMA intercept, slope and r^2 values are given in Table 2.

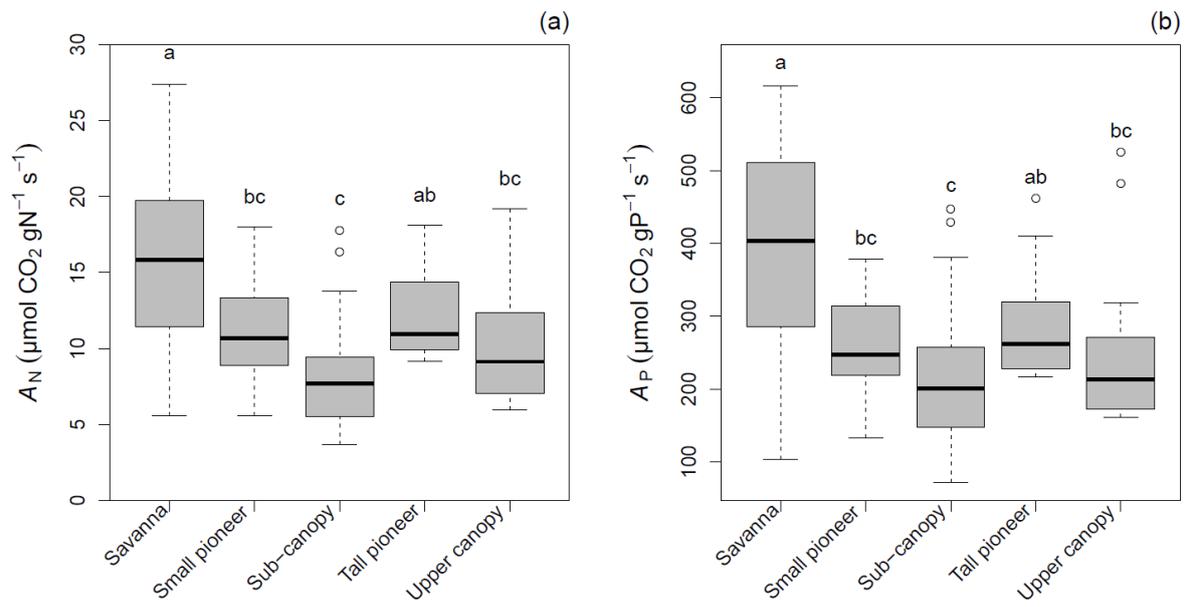


Figure 4. Boxplots of photosynthetic N use efficiency (A_N) and P use efficiency (A_P) by Φ . Boxes which share the same letter correspond to means that were not significantly different (Tukey's HSD on ranked values). Boxplot construction is explained in Figure 1.

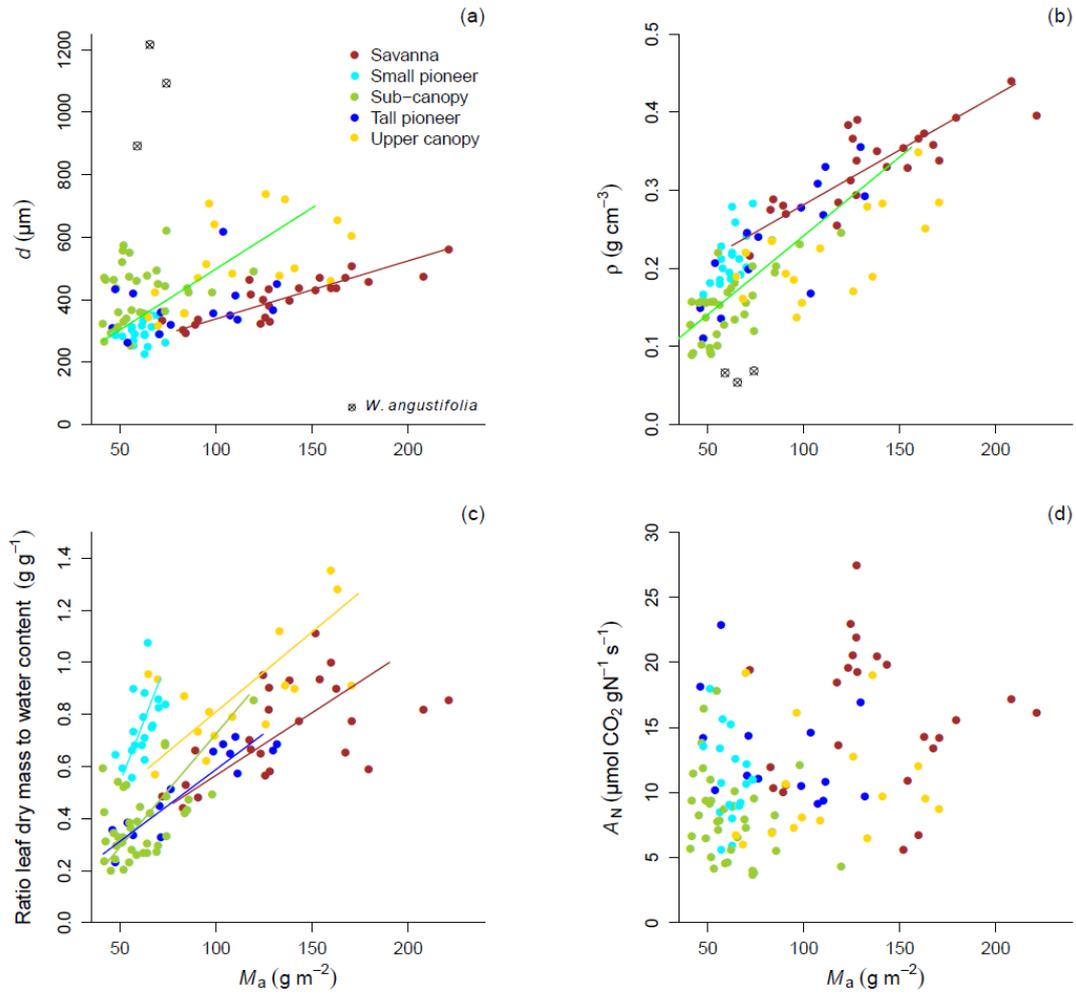


Figure 5: Scatterplots of the relationship with leaf mass per unit area for each of (a) leaf thickness (measured fresh), (b) leaf density (derived as M_a/d), (c) the ratio of leaf dry mass to water content and (d) photosynthetic N use efficiency. Each point represents one tree and separate Φ are distinguished by colour. Standardised major axis (SMA) fitted lines are shown in the top panel for the two vegetation types: \mathcal{F} in green and \mathcal{S} in brown. Outlying values for *Wilkiea angustifolia* (crossed circles) were excluded from the SMA analyses of plots (a) and (b). In plot (c) a separate fitted line is shown for each Φ . Pearson's correlations testing the assumption of linearity are given at Table S3 (Supplementary Information) together with Likelihood ratio and Wald statistics testing the H_0 of common slope, elevation and axis shift for the two \mathcal{V} classes. SMA intercept, slope and r^2 values are given in Table 2. There was no relationship $A_N \leftrightarrow M_a$ for individual levels of \mathcal{V} or Φ .

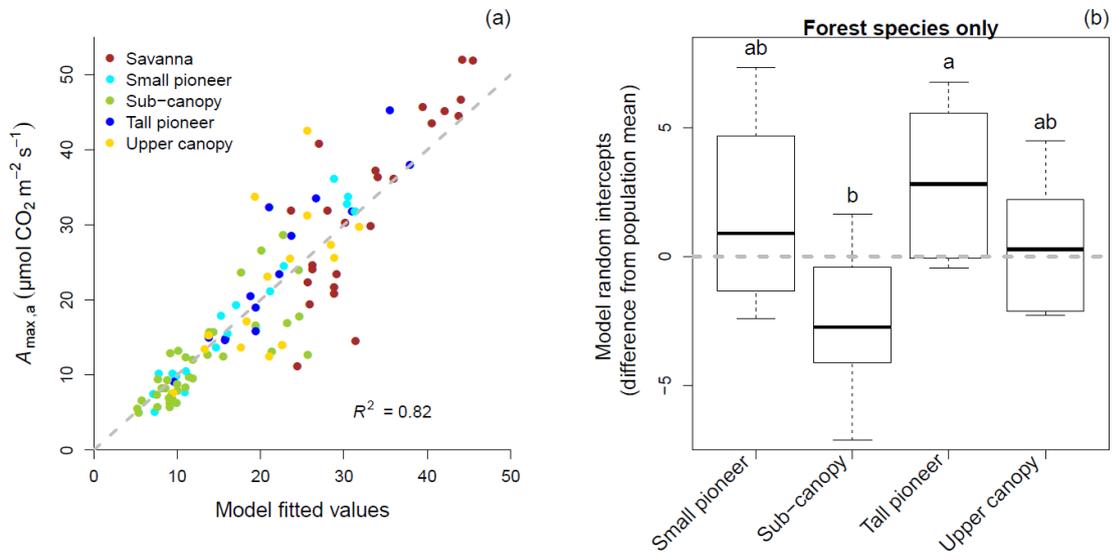


Figure 6. Output from linear mixed effects model (Eq. 2): (a) scatterplot of observed photosynthetic capacity ($A_{\max,a}$) against the model fitted values as an indication of goodness of fit and (b) boxplot of model random intercepts ($A_{\max,a} \leftrightarrow [N]_a$) by Φ for the \mathcal{F} subset; boxes which share the same letter correspond to means that were not significantly different (Tukey's HSD on ranked values). Boxplot construction is explained in Figure 1.