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# Carbon sequestration in managed temperate coniferous forests under climate change

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## Abstract

Management of temperate forests has the potential to increase carbon sinks and mitigate climate change. However, those opportunities may be confounded by negative climate change impacts. We therefore need a better understanding of climate change alterations to temperate forest carbon dynamics before developing mitigation strategies. The purpose of this project was to investigate the interactions of species composition, fire, management and climate change on the Copper–Pine creek valley, a temperate coniferous forest with a wide range of growing conditions. To do so, we used the LANDIS-II modelling framework including the new Forest Carbon Succession extension to simulate forest ecosystems under four different productivity scenarios, with and without climate change effects, until 2050. Significantly, the new extension allowed us to calculate the Net Sector Productivity, a carbon accounting metric that integrates above and below-ground carbon dynamics, disturbances, and the eventual fate of forest products. The model output was validated against literature values. The results implied that the species optimum growing conditions relative to current and future conditions strongly influenced future carbon dynamics. Warmer growing conditions led to increased carbon sinks and storage in the colder and wetter ecoregions but not necessarily in the others. Climate change impacts varied among species and site conditions and this indicates that both of these components need to be taken into account in when considering climate change mitigation activities and adaptive management. The introduction of a new carbon indicator – Net Sector Productivity, promises to be useful in assessing management effectiveness and mitigation activities.

## 1 Introduction

As a global society, we depend on forests and land to take up about  $2.5 + 1.3 \text{ Pg C y}^{-1}$ , about one-third of our fossil emissions (Ciais et al., 2013). A reduction in the size of these sinks could accelerate global change by further increasing the accumulation rate

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of greenhouse gases in the atmosphere. However, even a minor improvement to these biological sinks could help mitigate climate change because of their large scale.

Temperate forests offer many opportunities for increasing carbon sinks; however the risk of negative climate change effects and poor management decisions may limit these opportunities. For example, starting from 2000 a bark beetle outbreak (*Dendroctonus ponderosae*) caused in part by climate change (warmer winters), combined with the management response (increased logging) created a large carbon emission in the central interior of the province of British Columbia (BC), Canada (Kurz et al., 2008). In contrast, increased tree species productivity due to climate change effects could help create a net carbon sink, even with an increase in wildfire (Metsaranta et al., 2011). Without an integrated, landscape-scale understanding of climate change impacts on forests, we are limited in our management capacity to maintain the existing carbon storage or enhance sink strength.

Forest carbon dynamics depend on the management regime, expected growth and mortality rates, regeneration ingress, decomposition rates, and natural disturbances (Canadell and Raupach, 2008). Existing literature documents the complexity of forest carbon dynamics to potential rising temperatures, changing precipitation patterns, increasing atmospheric CO<sub>2</sub> and nitrogen availability. For example, stand-level modelling of future conditions in Colorado found that projected carbon stocks varied with future climate scenarios, and in some cases stocks decreased as the area became non-forested due to a loss of tree species viability (Buma and Wessman, 2013). In their study, adaptive management maintained forest carbon stocks in most climate scenarios, but with different species and lower tree densities than currently occur in the ecosystem. In contrast, results from Oregon using an earth system model projected increased net primary productivity and net biome productivity in the future forest ecosystem although, more intensive management increased net emissions (Hudiburg et al., 2013). Other studies have found minor climate change effects on net primary productivity and forest carbon stocks; and that greater differences were caused by local

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variation in growing conditions (e.g. Scheller et al., 2012). Because of these divergent results, climate change effects on temperate forests are not yet generalizable.

An additional aspect of forest carbon dynamics typically excluded from ecosystem studies is the storage of carbon in harvested wood products. The storage and emissions from wood products has been shown to be important for considering emissions due to forest management, climate change mitigation activities, and life cycle assessments (e.g. Hennigar et al., 2008; Smyth et al., 2014; Lamers et al., 2014). While the combination of ecosystem and wood product carbon dynamics are recognized as important, there is a mixture of indicators (typically stocks) and terms in the literature. Here we propose a new metric: Net Sector Productivity, to facilitate calculation and comparison among studies. This metric is based on the Net Ecosystem Productivity minus emissions from disturbances and wood products.

Our purpose was to improve our understanding of the interactions of species composition, climate change, fire, and management on temperate forest ecosystem carbon dynamics. The Copper–Pine creek valley in north-western BC provides an exemplar landscape because it includes a variety of forest ecosystems with naturally varying climate envelopes, tree species composition, management activities and natural disturbance rates within a relatively small area of under 750 km<sup>2</sup>. Furthermore, a recent study in a neighbouring area by Nitschke et al. (2012) demonstrated stand-level responses to climate change as an interaction of species-response, existing stand conditions, disturbance type, competition, and resource availability. To achieve our purpose, we had the following objectives: (1) project species productivity on different site types using down-scaled circulation model projections and a mechanistic tree species productivity model, (2) parameterize a new extension of the LANDIS-II landscape model that estimates ecosystem carbon dynamics, (3) assess model behavior by comparing it with available literature on carbon stocks and fluxes, (4) project ecosystem dynamics until 2050 under different productivity scenarios; and, (5) assess the landscape scale responses of carbon fluxes and stocks under climate change.

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production generally follow the Biomass Succession (v2) extension and the methods outlined in Scheller and Mladenoff (2004). The primary exceptions are that we added root pools and their growth, turnover, and mortality dynamics; and added greater user control over disturbance impacts. For the Copper–Pine creek study area, root parameters were based on literature values (Li et al., 2003; Mokany et al., 2006; Yuan and Chen, 2010). The modelling of decay in dead organic matter and soil pools generally follows the methods described in Kurz et al. (2009). That paper also provided the decay parameters for the Copper–Pine creek study. More detail is available in the user’s guide (Dymond et al., 2015). Terminology follows Chapin et al. (2006) and positive values of NEP and NBP indicate forest sinks.

The ForCSv2 extension is integrated with harvesting, fire and wind extensions of LANDIS-II. When a disturbance occurs, species-age cohorts may be killed by the disturbance extension. The transfers of carbon from biomass pools to dead organic matter, air or the forest products sector are controlled by user input. In addition, disturbances can trigger emissions and transfers from the dead organic matter or soil pools. For the Copper–Pine creek study area, wildfire impacts on carbon pools were based on Campbell et al. (2007). For harvest impacts, the model transferred 80 % of the merchantable-sized wood biomass out of the ecosystem during an event; any other killed biomass was transferred to the DOM pools.

LANDIS-II has stochastic processes including wildfires and natural regeneration. Therefore, we calculated landscape averages and standard deviations from 20 Monte Carlo replicates to conduct *t* tests comparing the results without climate change effects against the results from the average productivity with climate change scenario in 2050.

The harvested carbon output from ForCSv2 was run through the British Columbia Harvested Wood Product (v1) model (Dymond, 2012) to estimate storage and emissions on an annual basis. Those wood product emission estimates and wildfire emissions were subtracted from NEP to calculate the Net Sector Productivity (NSP).



## 2.3 Model input data

### 2.3.1 Growth and reproduction

For the Copper–Pine creek study area we gathered species life history parameters required by ForCSv2 from the literature (Table 2). The main sources of information were Klinka et al. (2000) and Burns et al. (1990). Additional information for *Populus tremuloides* (At, trembling aspen) and *P. balsamifera* (Ac, poplar) was available from Peterson et al. (1996). However, these reviews provided insufficient information for parameterizing the seed dispersal algorithm in ForCSv2. We found additional information on seed dispersal for *Picea Engelmannii* X *glauca* (Sx, interior spruce) (Squillace, 1954; Roe, 1967), *Pinus contorta* var. *latifolia* (Pl, lodgepole pine) (Boe, 1956; Dahms, 1963), trembling aspen (McDonough, 1986), *Tsuga heterophylla* (Hw, western hemlock) (Pickford, 1929; Beach and Halpern, 2001), poplar (Zasada et al., 1981), *Abies amabilis* (Ba, amabilis fir) (Heatherington, 1965), and *Betula papyrifera* (Ep, paper birch) (Bjorkbom, 1971; Greene and Johnson, 1995). Longevities were capped at the maximum ages documented in the local forest inventory to reflect local conditions.

The spatial forest inventory dataset maintained by the Government of BC provided the plant species and age information for the initial communities map (BC MFLNR, 2011). The leading species in the inventory was most frequently *Abies lasiocarpa* (Bl, subalpine fir) (62 %) and the second most frequent was lodgepole pine (14 %). Most stands did not have a second species listed (76 % of area). When it was listed, the second species was most frequently interior spruce.

We obtained 144 model simulations from five global climate models from the Pacific Climate Impacts Consortium (PCIC, 2012). Climate change is projected to increase the study area's mean annual temperature by 1 to 3.5 °C by the 2040–2069 period, depending on the global climate models (PCIC, 2012). Mean annual precipitation projections are more variable with models showing increasing, decreasing or unchanging precipitation. The monthly minimum, and maximum temperatures and precipitation

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were used to model the probability of establishment ( $P_{est}$ ), maximum aboveground net primary productivity (ANPP) and maximum biomass inputs for ForCSv2.

We used the Tree and Climate Assessment Tool Establishment Model (TACA-EM) to estimate the  $P_{est}$  through natural regeneration based on parameters in Table 3 (Nitschke and Innes, 2008; Nitschke et al., 2012). TACA-EM estimates the probability of a tree species to regenerate naturally given soil and climate site conditions (Nitschke and Innes, 2008). The TACA-EM probabilities are for a 3-year period, so we divided them by three to get the annual input to ForCSv2.

We used the Tree and Climate Assessment Tool Growth and Productivity model (TACA-GAP) model to estimate maximum ANPP and maximum biomass variables for each species in each ecoregion. TACA-GAP uses the growth and response functions in the BRIND (Shugart and Noble, 1981) and ZELIG++ (Burton and Cumming, 1995) models but is run at a daily time step to incorporate the snow, soil moisture and phenology components of TACA-EM (Nitschke et al., 2012). The TACA-GAP simulated individual species growth potential (biomass) over a range of soil and climate conditions (Table 3). The model does not simulate stand dynamics and interspecific competition rather the effects of temperature; drought, frost and soil moisture on growth. Species parameterization followed Nitschke et al. (2012). Weather stations are not located in the parkland ecoregions (i.e. 1 and 2) therefore regeneration and biomass variables were set to 50 % of the non-parkland ecoregion values (i.e. 3 and 4 respectively). From the ensemble of 144 future climate projections, we generated an average and standard deviation for productivity for the 2040–2069 period for each species in each ecoregion. To represent the uncertainty in future productivity, we defined the average productivity, low productivity (average minus one standard deviation), and high productivity (average plus one standard deviation) as scenarios.

### 2.3.2 Disturbances

To parameterize the fire regimes we used a combination of available information and scenarios representing possible disturbance regimes. Natural resource managers in

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the study area typically assume rates of natural disturbance based on the biogeoclimatic zones (BC Environment, 1995). We analyzed the fire maps maintained by the Government of BC from the study area and the surrounding region indicated a much lower fire cycle than is assumed by managers (data not shown). Furthermore, studies by Haughain et al. (2012) and Boulanger et al. (2012) also indicate a low fire hazard in the region. Based on the climate parameters and spatial arrangement in the study area, the ecoregions were grouped into the fire regime zones listed in Table 1. The disturbance return intervals for the fire regime zones were assumed to be double those used for forest management. Climate change alterations to the fire regimes are expected to be small, and therefore none were simulated (Haughain et al., 2012).

Natural resource management in the study area is primarily focused on harvesting, recreation, and cultural values. In BC constraints on harvesting include wildlife trees, old-growth retention requirements, adjacency requirements, visual quality concerns, water quality, and recreation activities. Therefore, we used different management zones in simulating a range of harvesting and reforestation activities. Harvesting and planting prescriptions were based on the forest stewardship plans for the Wetzink'wa Community Forest Corporation (2009) and BC Timber Sales – Babine (2007) (Table 4). Local forest managers reviewed the harvest parameters and results for accuracy.

### 3 Results

To determine the credibility of our model results, we conducted a model comparison based on literature values (Table 5). However, the literature review demonstrated that carbon stocks in forests are highly variable with site type and age. The ForCSv2 carbon stock estimates for Copper–Pine creek were within the range of other published values for temperate coniferous forests except for the coldest ecoregions (1 and 2) which were relatively low. Likewise, carbon fluxes can vary depending on site type, age, inter-annual weather patterns, disturbances, and different models. The ForCSv2 results seem reasonable compared to the literature values, except again for ecoregion





of carbon due to disturbances by using NBP lessened the differences between the simulations with or without climate change. The landscape was projected to have a NBP closer to zero under the average productivity scenario compared with a sink under no change.

Climate change was projected to have no effect on the ability of forest managers to achieve the harvest as currently planned (Fig. 8a). However, the harvest rate markedly affected estimates of net carbon fluxes with the lowest flux values in the first decade when harvest rates were highest (Fig. 8b). Similarly, the difference between the NSP and NBP is greatest during that first decade when harvest rates are high and therefore considering the storage of carbon in wood products created a noticeable difference at the landscape scale. However, there were no visible trends in the NSP between the no climate change scenario and the average productivity scenario, although only one replicate is shown (Fig. 8c).

Despite our efforts to model climate change effects for each there were no apparent changes to the distribution of the leading species (Fig. S4). There was however, a marked reduction of subalpine fir and an increase of lodgepole pine and interior spruce as leading species through management activity. In contrast, the climate change scenarios did show a marked change in aboveground biomass stocks and spatial distribution of western hemlock (Fig. 9).

## 4 Discussion

The purpose of this study was to improve our understanding of the interactions of species composition, climate change, fire, and management on temperate forest ecosystem carbon dynamics. Therefore we simulated the climate change impacts on productivity and natural regeneration interacting with management and wildfires within a region with steep elevational gradients using a new extension for LANDIS-II. Our results indicate that the effects of climate change on forest productivity and ecosystem carbon dynamics may be significant and substantial, but not uniform. The direction

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drier and continental type climate under the HadGEM1 A1B. The warmer and moisture climate would likely coincide with an increase of western hemlock biomass, as found in our study. This species easily regenerates naturally in suitable conditions and is a co-dominant in the warmer and moisture forests to the west of the study area.

## 4.2 Management implications

The projected leading species of the study area was, to a great extent, driven by management activities, planting in particular. This result reinforces the opportunities identified by others to adapt to climate change through management (e.g. Steenberg et al., 2011; Buma and Wessman, 2013). Adaptation may take the form of planting species currently viable, but provenances more suitable to future climatic conditions than the ones in the local geographic area (Rehfeldt et al., 1999). That action could also provide climate change mitigation if it prevents declines in productivity. In addition, increasing tree species diversity may increase resilience to forest health damage or as a strategy for dealing with the uncertainty in future projections (Dymond et al., 2014).

The harvest rate in our study was highly variable over time due to the mortality caused by mountain pine beetle triggering salvage logging in the near term in the Wetzink'wa Community Forest (Fig. 8a). Similarly, BC Timber Sales anticipates logging rates decreasing within the study area by about 2020 in part because they operate across a much larger area. The planned harvest was achieved in the simulations despite declining productivity in some areas. This was likely due to the age class distribution of the forest being over 100 years old (Fig. 2). The near-term harvest relies on trees that have already reached maturity, and therefore the growing stock already exists on the landscape. A longer simulation period that incorporates harvesting of second growth stands may have different results. The changing productivity could lead to changes in harvest rates. If monitoring substantiates the projected productivity increases in ecoregions 3 and 4, there may be capacity to increase harvest. This would be consistent with the results found by Steenberg et al. (2011) that sustainable harvest could increase assuming higher productivity under climate change.

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The NSP provides a metric that is sensitive to management changes in the forest, as indicated by the larger difference between the NSP and NBP when harvest rates were higher (Fig. 8b). Based on the wood product model documented behavior (Dymond, 2012), the NSP will likely also be sensitive to the lifespan of products and their disposal.

5 Therefore, we suggest this metric would be particularly useful when assessing climate change mitigation options available to the forest industry.

### 4.3 Modelling confidence and caveats

This study not only assessed climate change impacts on the Copper–Pine creek valley, but also provided a test case for the ForCSv2 extension to LANDIS-II. Unfortunately, 10 whether the model is based on allometric equations (field plots), flux tower data, or more complex simulation models, it is nearly impossible to directly measure carbon stocks or fluxes and so we must rely on model inter-comparisons. The comparison of carbon stocks and fluxes with literature values in Table 5 provides some confidence that the ForCSv2 output is reasonable, although the variability is large. Therefore, this model is likely most useful for assessing differences between climate, management, or 15 disturbance scenarios, rather than for predicting absolute values.

The LANDIS-II modelling of aboveground biomass, tree species growth, competition, and natural regeneration has been extensively investigated and the strengths and weaknesses are understood (e.g. Simons-Legaard et al., 2015). The landscape NPP and aboveground biomass are highly sensitive to the input variables: maximum NPP and maximum biomass for each species in each ecoregion and the growth parameter  $r$ . Also, they found the aboveground biomass tended to increase as the duration of the simulation increased over 30 years. Since the ForCSv2 extension biomass dynamics are based on the Biomass Succession extension analyzed in their study, we can 20 assume a similar sensitivity for NPP, aboveground biomass, NEP, NBP and NSP.

The ForCSv2 DOM and soil dynamics are built from the CBM-CFS3 (Kurz et al., 2009). The CBM-CFS3 has also been investigated for parameter sensitivity (e.g. White et al., 2008), compared with field estimates of carbon stocks (Shaw et al., 2014) and

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with other estimates of NEP (e.g. Wang et al., 2011). White et al. (2008) found that the DOM and soil carbon stocks and stock changes were most sensitive to the base decay rates for the above- and belowground slow pools and the transfer to air for the above- and belowground very fast pools. Shaw et al. (2013) found that the CBM-CFS3 model output was reliable for estimating total ecosystems stocks for the forests of Canada. However, they did find it overestimated deadwood and underestimated forest floor and mineral soil carbon stocks, primarily in stands of balsam fir, white and black spruce due to the model not representing moss. Those stand types are not found in the Copper–Pine Creek study area. Wang et al. (2011) demonstrated the large uncertainty between different estimates of NEP among six models over eight years for a relatively small area around a flux tower ( $-200$  to  $+850$   $\text{g C m}^{-2} \text{y}^{-1}$ ). The CBM-CFS3 results were within the range of other estimates.

The productivity estimates used as input to ForCSv2 did not include the positive impact of  $\text{CO}_2$  or N fertilization (Wu et al., 2014) or negative impact of provenance (local adaptation) (e.g. O'Neill and Nigh, 2011). These would increase the uncertainty of model outputs.

## 5 Conclusions

The results indicated that the relative position of species optimum to current and future site conditions strongly influenced projections of landscape carbon dynamics. Those productivity rates interacted with respiration and disturbance rates to shape the dynamics of net carbon fluxes of the ecosystem, biome and sector. Climate change effects on forests vary with species, site conditions, management and fire regime, therefore all of these components need to be considered when planning climate change mitigation and adaptive management. This type of future research may consider ForCSv2 as a viable model within the LANDIS-II framework.



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**Table 1.** Ecoregions for LANDIS-II, biogeoclimatic variant names as used in BC, and fire regime zones as used in this study.

Ecoregion number <sup>a</sup>	Biogeoclimatic variants <sup>b</sup>	Climate <sup>c</sup> MAT <sup>d</sup> (°C)	1961–90 MAP <sup>e</sup> (mm)	Climate MAT (°C)	2040–69 MAP (mm)	Fire regime zone	Fire return interval
1	Engelmann Spruce – Subalpine Fir, Moist Cold Parkland	0.3	1307	2.8	1404	Upper slopes	700
2	Engelmann Spruce – Subalpine Fir, Wet Very Cold Parkland	0.5	1602	2.9	1732	Upper slopes	700
3	Engelmann Spruce – Subalpine Fir, Moist Cold	1.4	1081	3.8	1161	Upper slopes	700
4	Engelmann Spruce – Subalpine Fir, Wet Very Cold	1.6	1291	4.0	1395	Upper slopes	700
5	Sub-Boreal Spruce, Moist Cold, Babine	2.2	851	4.6	910	Lower slopes	400
6	Interior Cedar – Hemlock, Moist Cold, Nass	2.3	899	4.7	964	Lower slopes	400
7	Sub-Boreal Spruce, Dry Cool	3.1	521	5.5	548	SBSdk	200

<sup>a</sup> Ecoregion number based on rank order of mean annual temperature.

<sup>b</sup> BC Environment (1995).

<sup>c</sup> Source: PCIC (2012).

<sup>d</sup> Mean annual temperature.

<sup>e</sup> Mean annual precipitation.

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**Table 2.** Life history attributes for LANDIS-II.

Species Code	Species	Longevity (years)	Sexual maturity (years)	Shade tolerance class	Fire tolerance class	Effective seed dispersal (m)	Maximum seed dispersal (m)	Probability of resprouts	Minimum age for re-sprouting	Maximum age for re-sprouting	Post-fire regeneration
Ac	<i>Populus trichocarpa</i>	200	10	1	3	50	199	0.75	10	199	resprout
At	<i>Populus tremuloides</i>	200	10	1	3	50	499	0.5	10	149	resprout
Ba	<i>Abies amabilis</i>	340	25	5	3	38	120	0	0	0	none
Bl	<i>Abies lasiocarpa</i>	400	20	3	3	38	99	0	0	0	none
Ep	<i>Betula papyrifera</i>	200	30	2	2	50	470	0	15	199	resprout
Hw	<i>Tsuga heterophylla</i>	325	20	5	3	50	1399	0	0	0	none
Pa	<i>Pinus albicaulis</i>	325	10	1	3	50	101	0	0	0	none
Pl	<i>Pinus contorta</i>	300	7	1	3	20	199	0	0	0	serotiny
Sb	<i>Picea mariana</i>	250	10	4	3	20	101	0	0	0	none
Sx	<i>Picea Engelmannii X glauca</i>	325	30	2	3	30	299	0	0	0	none

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**Table 3.** Life history attributes for TACA-EM and TACA-GAP. See Table 1 for species codes.

Species code	Base temp (°C)	Bud burst (GDD <sup>a</sup> )	Chilling req. <sup>b</sup> (days)	Lethal temp. (°C)	Drought tol. <sup>c</sup>	GDD min	GDD max	Frost tol.	Frost days	Wet soils tol.	AHMI <sup>d</sup>	<i>D</i> max <sup>e</sup> (cm)	<i>H</i> max <sup>f</sup> (m)	<i>A</i> max <sup>g</sup> (y)	Shade tol.
Ac	4.6	175	70	−60	0.13	258	5263	0.5	295	0.55	62.3	200	4500	250	1
At	3.5	189	70	−80	0.4	227	4414	0.9	284	0.3	40	95	3900	200	1
Ba	4.3	307	91	−35	0.4	206	3877	0.3	305	0.55	41.4	182	6200	440	2
Bl	2.6	119	60	−67	0.25	198	5444	0.9	320	0.75	28.7	150	4100	320	2
Ep	3.7	231	77	−80	0.3	237	4122	0.9	285	0.3	40	76	3000	140	1
Hw	4.1	277	56	−39	0.25	328	5861	0.1	265	0.55	36.8	225	8000	500	2
Pa	3	120	70	−55	0.4	216	3352	0.9	320	0.05	34.2	200	3500	600	1
Pl	2.9	116	63	−85	0.42	186	3374	0.9	320	0.5	37.9	130	4500	335	1
Sb	3	123	56	−69	0.3	144	3060	0.9	305	1	42.7	46	2700	250	2
Sx	2.9	146	45	−58	0.3	139	3331	0.9	305	0.5	43.2	171	5100	430	2

<sup>a</sup> GDD is Growing Degree Days.

<sup>b</sup> Req. is requirement.

<sup>c</sup> Tol. is Tolerance.

<sup>d</sup> AHMI is annual heat moisture index.

<sup>e</sup> *D* max is maximum diameter.

<sup>f</sup> *H* max is maximum height.

<sup>g</sup> *A* max is maximum age.

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Name	Time period	Harvest rate (% y – 1)	Planting
Pine-targeted clear cut	2012–2017	1 to 1.8	interior spruce, subalpine fir, lodgepole pine
WCF-Clearcut early	2012–2017	1	interior spruce, subalpine fir, lodgepole pine
WCF-Clearcut	2018–2060	0.33	interior spruce, subalpine fir, lodgepole pine
BCTS-Clearcut north-west	2015–2035	2 to 4	interior spruce, subalpine fir, lodgepole pine
BCTS-Clearcut south-west	2012–2060	0.8 to 1.2	interior spruce, subalpine fir, lodgepole pine
Forest health patch-cut (1 ha)	2012–2060	0.08 to 0.3	interior spruce, subalpine fir, lodgepole pine



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**Table 5.** Model comparison of various temperate forest carbon indicators between published values and this study. Means  $\pm$  SD. Units are  $\text{gCm}^{-2}$  or  $\text{gCm}^{-2}\text{y}^{-1}$ .

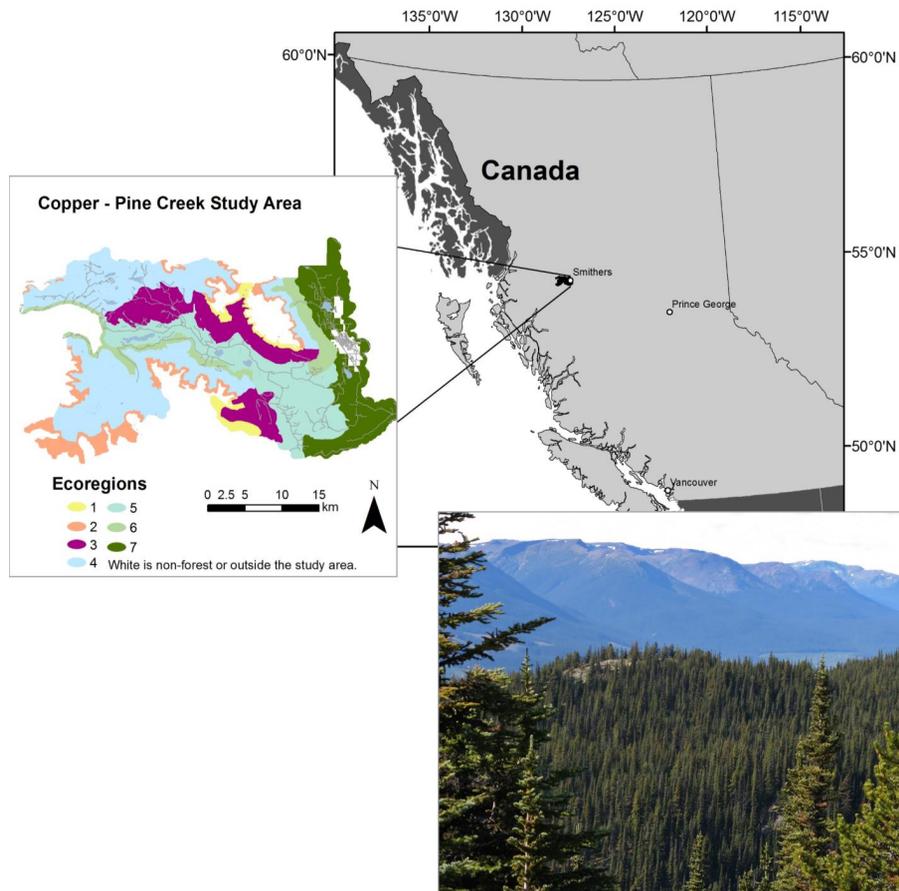
Forest carbon indicator	Stand models <sup>a</sup>	Eddy covariance studies <sup>b</sup>	Stock change model <sup>c</sup>	This study, 2012 Ecoregion 1	Ecoregion 4	Ecoregion 7
Aboveground biomass	2500 to 36 000	4952 $\pm$ 3417	8472 to 9786	1160 $\pm$ 489	4454 $\pm$ 2048	9770 $\pm$ 2132
Roots	800 to 8000	1209 $\pm$ 875	1876 to 2050	339 $\pm$ 207	1301 $\pm$ 600	2853 $\pm$ 623
DOM and Soil	6700 to 16 850		16 016 to 27 619	2384 $\pm$ 840	15 855 $\pm$ 5157	27 300 $\pm$ 6655
Total ecosystem	23 900 to 30 900		28 114 to 41 290	3883 $\pm$ 1230	21 610 $\pm$ 6848	39 922 $\pm$ 8607
NPP		281 $\pm$ 127	463 to 541	37.8 $\pm$ 24	197 $\pm$ 126	642 $\pm$ 161
Rh		396 $\pm$ 155	397 to 578	38.4 $\pm$ 13	253 $\pm$ 94	563 $\pm$ 117
NEP		93 $\pm$ 185	−36 to 75	−0.55 $\pm$ 17	−56.7 $\pm$ 89	79.4 $\pm$ 134
NBP			−93 to 71	−0.55 $\pm$ 17	−75.6 $\pm$ 375	56.9 $\pm$ 541

<sup>a</sup> Fredeen et al. (2005) and Kranabetter (2009) sites are in or near the Copper–Pine creek study area. Gower and Grier (1989), Pregitzer and Euskirchen (2004).

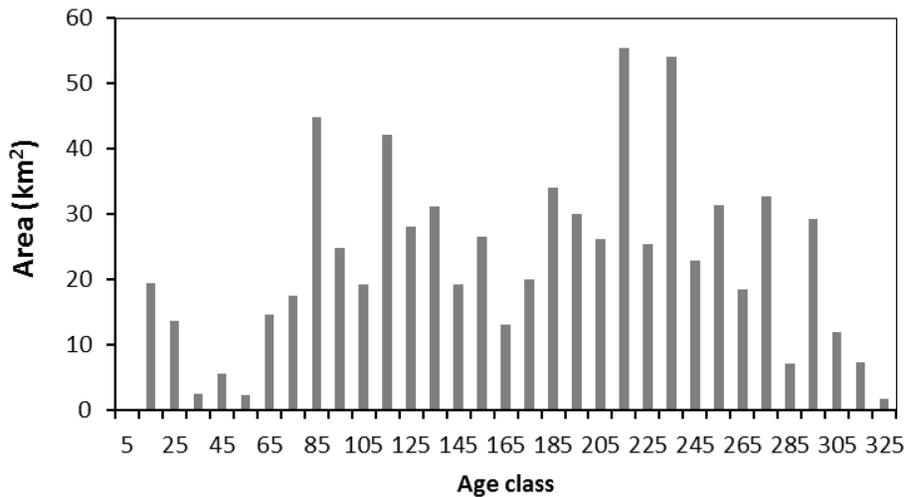
<sup>b</sup> Luyssaert et al. (2007), needle-leaved, boreal humid sites.

<sup>c</sup> Stinson et al. (2011), Bulkley Valley Timber Supply Area results extracted from the results database. Includes Copper–Pine creek study area except ecoregions 1 and 2.





**Figure 1.** The Copper–Pine creek study area (black polygon) near Smithers, Canada, ecoregions for LANDIS-II modelling and photograph looking south–west across part of the study area. See Table 1 for ecoregion descriptions.



**Figure 2.** Age class distribution in 2011 for the Copper–Pine creek study area.

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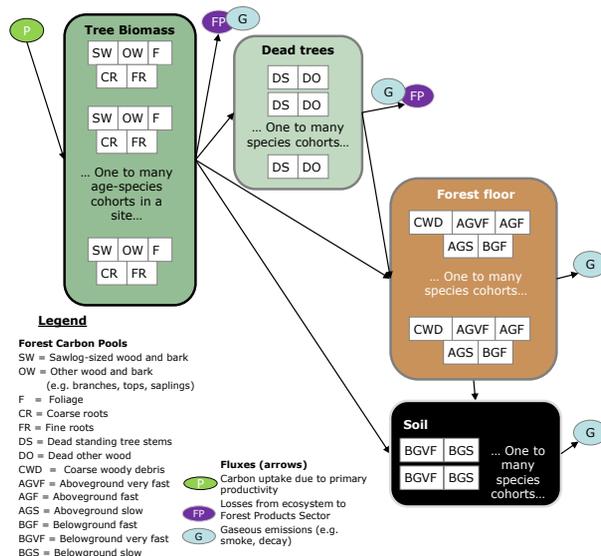
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**Figure 3.** Simplified pools and fluxes represented in the Forest Carbon Succession module (v2) for LANDIS-II. In the left panel, carbon accumulates in the tree biomass pools based on the primary productivity input data. When mortality of a whole or part of a tree occurs, the carbon is transferred to the dead organic matter and soil pools in the three right-hand panels, or may be removed from the ecosystem through harvesting or combustion. As decay occurs, carbon is transferred among the dead organic matter and soil pools, eventually entering the belowground slow pool (BGS) or being emitted from the ecosystem. Fire and harvesting can also cause transfers or emissions from the dead organic matter pools.

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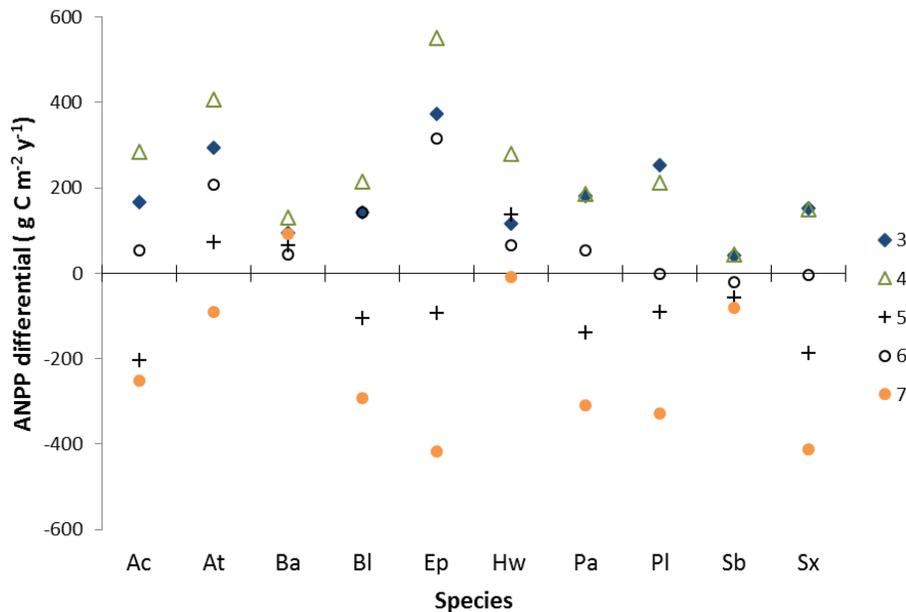
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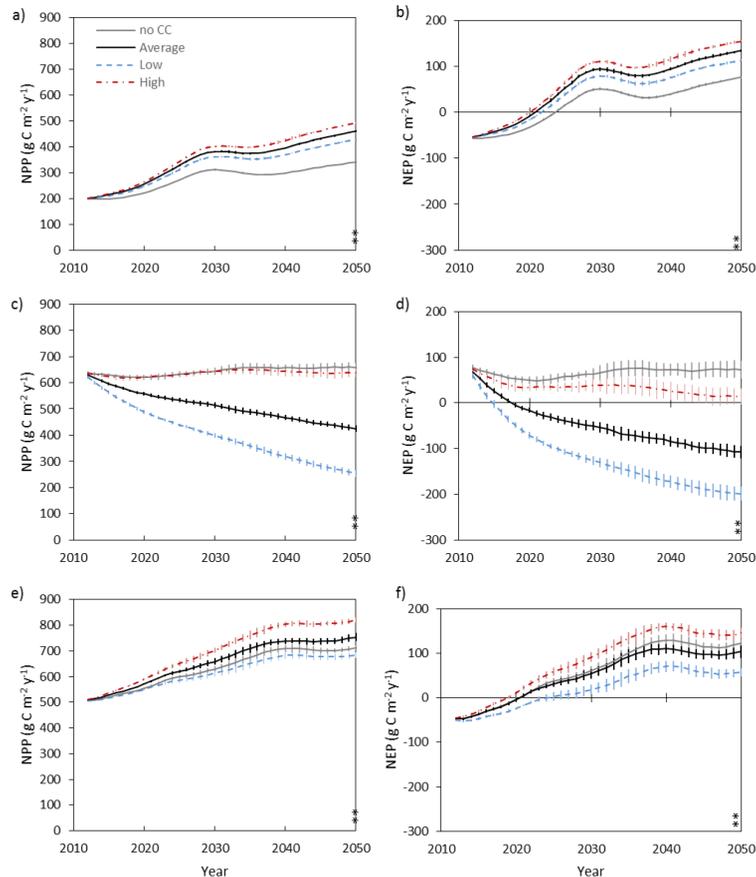


**Figure 4.** Maximum ANPP differential from 1961–90 climate to 2040–69 climate average in estimated by the TACA-GAP model for the five main modelling ecoregions in the study area. Input ANPP for ecoregions 1 and 2 were set at 50 % of regions 3 and 4 respectively.

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**Figure 5.** Climate change impact projections on the NPP and NEP (average + SD) for ecoregions 4 (a, b), 7 (c, d), and 6 (e, f). Asterisk notes  $t$  tests that were significantly different between the no change scenario (no CC) and climate change average productivity (\*\*  $P < 0.01$ ) in 2050. Note,  $y$  axes vary.

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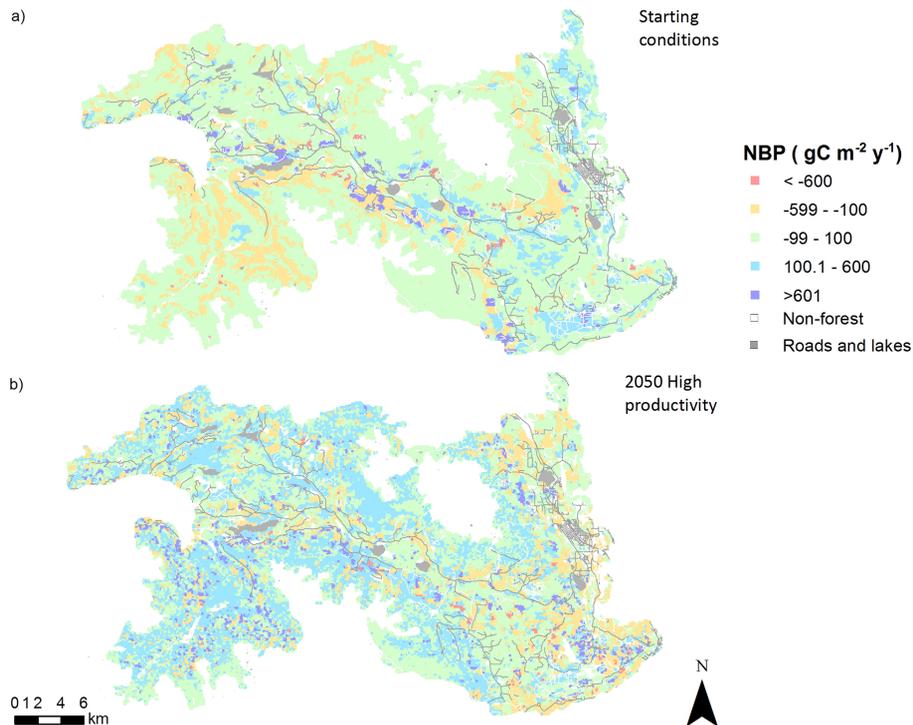
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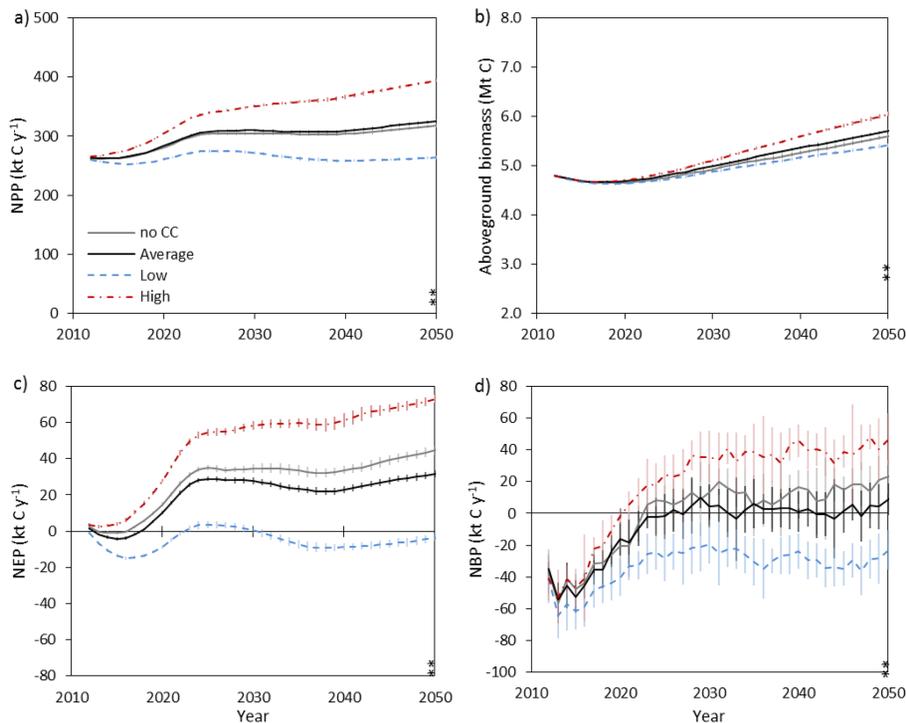
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**Figure 6.** Spatial distribution of NBP under the starting conditions **(a)** and in 2050 under the high productivity scenario **(b)**.

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**Figure 7. (a–d)** Landscape total carbon fluxes and aboveground biomass stocks (average + SD) for no climate change, average, high or low productivity scenarios. Asterisk notes *t* tests that were significantly different between the no change scenario (no CC) and average productivity scenario (\*\*  $P < 0.01$ ) in 2050.

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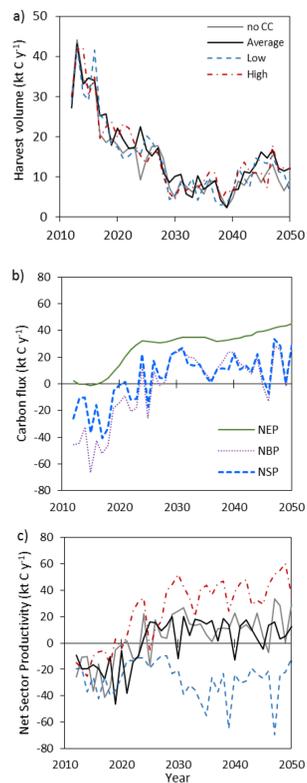
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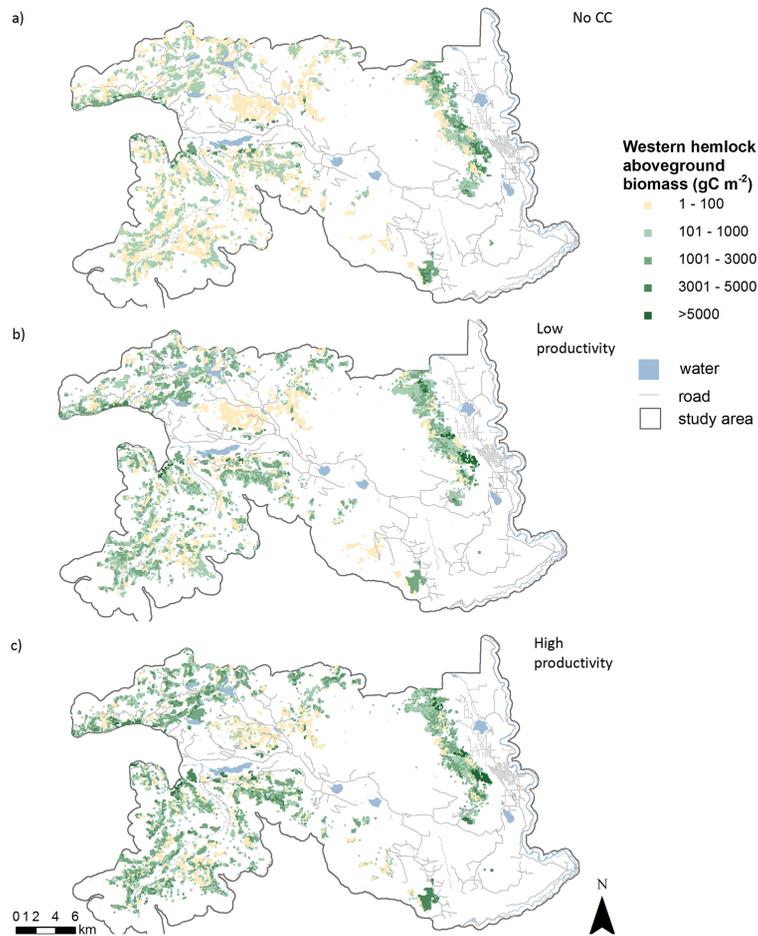
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Interactive Discussion



**Figure 8.** Relationship between harvest rate and carbon fluxes for a single replicate. Removal of carbon from the ecosystem through logging (a). NEP, NBP and NSP for a single replicate without climate change (b). Net sector productivity for a single replicate of each scenario (c).



**Figure 9.** Western hemlock biomass distribution in 2050 with no climate change (no CC), high and low productivity scenarios.