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# Distribution of phytoplankton functional types in high-nitrate low-chlorophyll waters in a new diagnostic ecological indicator model

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Received: 24 April 2013 – Accepted: 2 May 2013 – Published: 15 May 2013

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Published by Copernicus Publications on behalf of the European Geosciences Union.

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## Abstract

Modeling and monitoring plankton functional types (PFTs) is challenged by insufficient amount of field measurements to ground-truth both plankton models and bio-optical algorithms. In this study, we combine remote sensing data and a dynamic plankton model to simulate an ecologically-sound spatial and temporal distribution of phyto-PFTs. We apply an innovative ecological indicator approach to modeling PFTs, and focus on resolving the question of diatom-coccolithophore co-existence in the subpolar high-nitrate and low-chlorophyll regions. We choose an artificial neural network as our modeling framework because it has the potential to interpret complex nonlinear interactions governing complex adaptive systems, of which marine ecosystems are a prime example. Using ecological indicators that fulfill the criteria of measurability, sensitivity and specificity, we demonstrate that our diagnostic model correctly interprets some basic ecological rules similar to ones emerging from dynamic models. Our time series highlight a dynamic phyto-PFT community composition in all high latitude areas, and indicate seasonal co-existence of diatoms and coccolithophores. This observation, though consistent with in situ and remote sensing measurements, was so far not captured by state-of-the-art dynamic models which struggle to resolve this “paradox of the plankton”. We conclude that an ecological indicator approach is useful for ecological modeling of phytoplankton and potentially higher trophic levels. Finally, we speculate that it could serve as a powerful tool in advancing ecosystem-based management of marine resources.

## 1 Introduction

We are yet to obtain a consistent and complete view of the global biogeography of plankton functional types (PFTs), groups of organisms composed of many different species identified by a common biogeochemical function rather than a common phylogeny. We deal with large uncertainty due to insufficient amount of field measurements

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to ground-truth both plankton models (Anderson, 2005) and bio-optical phytoplankton PFT algorithms (Brewin et al., 2011). Knowing how PFT distributions are changing is key to projecting biological responses to global climate change, both in the geological past (Wells et al., 1999), present and future (Beaugrand et al., 2012; Bopp et al., 2005; Boyd and Doney, 2002). A requirement is to investigate regions such as the Southern Ocean, the subarctic North Pacific and the equatorial Pacific, the so-called high-nitrate low-chlorophyll (HNLC) regions (Minas and Minas, 1992) where the ocean acts as a source of CO<sub>2</sub> to the atmosphere (Takahashi et al., 2002) despite being sufficiently productive. According to modeling studies, these regions may also constitute PFT diversity hotspots (Prowe et al., 2012) or, to the contrary, PFT diversity deserts (Barton et al., 2010).

A variety of recently-developed phytoplankton size class (phyto-PSC) and phytoplankton PFT (phyto-PFT) bio-optical algorithms may be used to monitor regional and global phytoplankton distributions at an intraseasonal resolution (Alvain et al., 2008; Bracher et al., 2009; Brewin et al., 2010; Hirata et al., 2011). Among the key phyto-PFTs are silicifiers (diatoms) whose biomass can be estimated from satellite (Sathyendranath et al., 2004; Alvain et al., 2008; Hirata et al., 2013), and calcifiers (coccolithophores) whose biomass estimates up until very recently (Sadeghi et al., 2012) were limited to bloom occurrence and area calculation (Alvain et al., 2008; Moore et al., 2012) and inference from particulate inorganic carbon (PIC) measurements (Balch et al., 2011). However, field measurements of biological processes, on the level of PFTs in particular, are not sufficiently resolved compared to chemical and physical properties of the ocean (Claustre et al., 2010). Thus, synoptic relationships between PFT biomass (derived from in situ phytoplankton cell and pigment abundance/biomass) and optical properties of the surface ocean (derived from remote sensing) cannot be constrained and evaluated consistently over all biogeochemical provinces. Borders between biogeochemical provinces shift dramatically on annual and longer scales (Boyd and Doney, 2002; Devred et al., 2007) challenging the integration of regional algorithms into global models (Cetinić et al., 2012).

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Dynamic plankton models coupled to ocean circulation models take changing environmental conditions into account. Moreover, they derive PFT estimates from knowledge of the underlying mechanistic processes. However, whereas they can discern between several to tens of different groups (Follows et al., 2007), they struggle with the long debated paradox of the plankton (Hutchinson, 1961). To our knowledge, none of the models operating on the PFT level can simulate the observed coexistence of more than one dominant group under limiting resources when their biogeochemical functions are similar yet need to be parameterised separately. For instance, fast-growing diatoms in the subarctic North Pacific and the Southern Ocean outcompete coccolithophores preventing their biomass to build up (e.g. Gregg and Casey, 2007; Le Quéré et al., 2005; Sinha et al., 2010) in contrast to what remote sensing observations suggest (e.g. Alvain et al., 2008; Balch et al., 2011; Sadeghi et al., 2012). A PFT biogeochemical model is sensitive to the type of PFT classification, choice of zooplankton grazing formulation (Hashioka et al., 2012) and, especially for diatoms and coccolithophores, to the mixing formulation in the chosen physical model coupled to the plankton model (Sinha et al., 2010). The MARine Ecosystem Model Intercomparison Project (MAREMIP) (Hashioka et al., 2012) currently examines the ability of these models to identify key processes and evaluate the role of functional groups in the whole ecosystem. Complementary to MAREMIP, the Satellite phyto-PFT Intercomparison Project (Hirata et al., 2012) is investigating the performance of satellite phyto-PFT models in the global ocean.

In a two-dimensional phytoplankton niche space defined by turbulence and nutrient concentration ("Margalef's Mandala"), coccolithophores traditionally fall between diatoms which thrive in well-mixed, high nutrients regimes, and dinoflagellates which dominate the stratified, low-nutrient regimes (Margalef, 1978). Today we know that picoeukaryotic and prokaryotic autotrophs (e.g. *Prochlorococcus*, *Synechococcus*) successfully compete with dinoflagellates for their niche. Light is another important niche descriptor. Balch (2004) suggested that daylength is as important as light intensity for the onset of coccolithophore (and likely other phyto-PFT) blooms, and should form

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the third dimension of phytoplankton mandala. The ability to locate phyto-PFTs in their ecological niches allows us to derive ecological rules that can verify both plankton models and bio-optical algorithms. However, ecological rules can serve another purpose: they help identify key ecological indicators of change of PFTs – a task imposed by the rapidly changing climate.

Having acknowledged the challenges in monitoring and modeling biological processes explicitly, an ecological indicator approach is an alternative means of describing and managing marine ecosystems (e.g. Dale and Beyeler, 2001; Blanchard et al., 2010). In this study, we explore the possibility of applying an ecological indicator approach to simulate an ecologically consistent global distribution of phyto-PFTs, with particular focus on diatoms and coccolithophores in the HNLC regions. We choose an artificial neural network (ANN) as our modeling framework because this artificial intelligence tool has the potential to interpret complex nonlinear interactions governing complex adaptive systems (Holland, 1995) of which marine ecosystems are a prime example (Levin, 1998). In order to enable projection of past and future phyto-PFT states, and their potential application in ecosystem management of marine resources (Palacz, 2012), we select ecological indicators that fulfill the criteria of indicators of Good Environmental Status (GES) (Commission, 2008). These criteria, described by Link et al. (2010) include: (i) measurability – the availability of data to estimate the indicator, (ii) sensitivity – the ability to detect change in an ecosystem, and (iii) specificity – the ability to link the said change in an indicator as a response to a known intervention or pressure.

The idea of using an ecological indicator approach to phyto-PFT modeling is not a new one. For instance, Raitos et al. (2006) attempted to explain variability in North Atlantic blooms of coccolithophores by identifying their ecological indicators of change obtained from a combination of in situ, satellite and model data. Application of ANNs in the ecological approach to phyto-PFT modeling was pioneered by Raitos et al. (2008) who estimated probability of diatom occurrence in the North Atlantic from ecological (e.g. sea surface temperature, photosynthetically available radiation,

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surface chlorophyll *a* concentration) and geographical (e.g. latitude, longitude) indicators. ANNs based on ecological indicators were also used to simulate the distribution of  $p\text{CO}_2$  in the North Atlantic (Telszewski et al., 2009), and to compare patterns of biological production in eastern boundary upwelling regions (Lachkar and Gruber, 2012).

In contrast to these earlier studies, we attempt to use only ecological indicators to simultaneously model biomass distribution of four phyto-PFTs in key biogeochemical provinces, including the open ocean HNLC regions. We hypothesize that our phyto-PFT ecological indicator model (hereafter PhytoANN) will:

- interpret the complex nonlinear interactions between four phyto-PFTs and their ecological indicators in a variety of distinct biogeochemical conditions;
- improve the existing model estimates of monthly climatology and time series distribution of diatoms and coccolithophores in the HNLC regions.

## 2 Methods

### 2.1 Source of indicators

We selected the following ecological indicators as principal inputs into the PhytoANN model: (i) sea surface temperature (SST), (ii) wind speed (Wspd), (iii) photosynthetically available radiation (PAR), (iv) surface chlorophyll *a* concentration (Chl), and (v) mixed layer depth (MLD). Although we considered two additional indicators, modeled surface nitrate ( $\text{NO}_3$ ) and surface iron (Fe) concentration, we did not include them in the final PhytoANN because they displayed significant bias with respect to observations in several biogeographic provinces, and they did not add significantly towards explaining patterns of phyto-PFT variability. We assume that changes in  $\text{NO}_3$  distribution can largely be explained by associated changes in SST and Chl. This assumption has been used to derive a  $\text{NO}_3$  index (Nelson et al., 2004) and  $\text{NO}_3$  maps from satellite data alone (Silió-Calzada et al., 2008). We do not consider geographical and time

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In order to avoid overfitting and assure the generality of our model, we use the early stopping procedure. In this technique, the available data is divided into three subsets: training, validation and testing. The training set (70 % of confirmatory analysis data) is used for computing the gradient and updating the network weights and biases. The validation set (15 % of confirmatory analysis data) is used to monitor the error during the training process because it initially decreases but later typically increases as the network begins to overfit the data. Hence, when the validation error increases for a specified number of iterations (six in this study), the training is stopped, and the weights and biases at the minimum of the validation error are returned. The error of calculated from the testing set (15 % of confirmatory analysis data) is not used during training explicitly. However, it is plotted during the training process to monitor whether the error in the testing set reaches a minimum at a significantly different iteration number than the validation set error. In such cases, it would indicate a poor division of the data set and a need for re-training.

It is not required to select only linearly independent indicators as inputs to the ANN because it performs dimensionality reduction on its own. Similarly, we do not remove any outliers because we assume the ANN is capable of distinguishing between signal and noise after analysing a sufficient number of training examples. We do however need to normalize the input data onto a common min-max scale to prevent indicators with highest absolute values from overpowering the neurons. Moreover, if the indicators or targets have a non-normal distribution, the ANN will produce results biased towards the more populated end of their range. In this study, we thus log-normalize MLD and Chl in the input layer and all four phyto-PFTs in the output layer. All values are then back-transformed to linear scales only after training and simulation are completed.

We also note that climatology and time series results presented in this study are in fact ensemble mean results from 10 PhytoANNs. No two ANNs are exactly the same because they are trained using random weight and bias vector initialization and a random distribution of datapoint indices assigned to training, evaluation or testing. Though individual simulations reveal variable absolute values of Phyto-PFT biomass, their time

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series distribution patterns are very robust. Therefore, we consider ensemble means as optimal representations of PhytoANN results.

In Table 2 we provide the details of the PhytoANN architecture common to all ensemble members, chosen as optimal for this study. Results in Table 3 show how increasing the number of neurons in the hidden layer (ANN complexity) increases the fit to the entire confirmatory data (both log-transformed and linear). In theory, the optimum number of neurons depends on the degree of linear independence of the patterns in hidden layer space (Teoh et al., 2006). We note that a good fit is obtained even using the simplest architecture. However, an inspection of time series distributions reveals that not all expected patterns are well captured by a five-neuron ANN. On the other hand, the 10- and 15-neuron nets are likely overfitted to the training data because they do not increase the fit substantially (Table 3). We concluded that the eight-neuron ANN was well-fitted yet general enough to simulate phyto-PFTs, also in a relatively short time (average of 228 s per training).

Table 4 revealed that the Levenberg–Marquardt and the BFGS Quasi-Newton training algorithms provide equally good fits to the data and take little time to perform. Except for the fact that the Levenberg–Marquardt algorithm is more common in feed-forward type of ANNs, its choice here is arbitrary. We relied on vast literature accounts of the usage of transfer and performance functions and did not test the sensitivity to the choice of these functions here.

In summary, we describe the entire procedure with consecutive steps from data selection and processing to PhytoANN model development and its application:

1. Select ecological indicators that fulfill the criteria of Link et al. (2010).
2. Assemble a matrix of input and target data and place them on a unified spatial and temporal grid.
3. Divide all available data between confirmatory and exploratory datasets.
4. Inspect histograms of individual input and target data, and transform them onto a log-10 scale if their distribution is non-normal.

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5. Normalize all processed input and target data onto a common minimum–maximum range (e.g.  $-1$  to  $1$ ) in order to avoid bias towards high value inputs/outputs.
6. Divide the confirmatory dataset into training (70%), testing (15%) and evaluation (15%) subsets, either randomly or systematically.
7. Set up a feedforward ANN with one input, one hidden layer and one output layer.
8. Select the type of transfer, performance and training function, and initial ANN parameters. Initialize connection weights and biases randomly within the network.
9. Train the network with early stopping.
10. Evaluate ANN performance within confirmatory regions by calculating error statistics.
11. Perform sensitivity analysis on key parameters listed above and retrain the ANN to maximize performance.
12. Apply a total of 10 trained nets to time series from exploratory regions.
13. Normalize phyto-PFT biomass with respect to total Chl to assure conservation of Chl biomass.
14. Compare the ensemble mean output with target as well as independent data.

### 3 Results and discussion

#### 3.1 Ecological niches of PFTs

- 20 The first aim is to investigate how the PhytoANN interprets the interactions between PFTs and their environmental indicators. We break down the problem into four

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questions here. Which interactions are linear and which nonlinear? How do the interpretations vary across PFTs? Does the model capture the same relationships that are described mathematically by the NOBM that was used to train it? How are the weights distributed among the interactions?

- 5 We answer the first three questions using information presented in Figs. 3 and 4. By plotting PhytoANN estimates phyto-PFT biomass against an observed range of one input at a time, we detect very nonlinear relationships in all cases, for all phyto-PFTs. However, some linear patterns emerge after we separate the results according to the spatial domain of origin. Consequently, we notice two distinct physical and/or biogeochemical regimes that can be separated into high and low latitude regions. All four phyto-PFTs exhibit similar responses to conditions in the NEAtl, NorwSea, NEPac and AntAtl (hereafter HighLat Regime), but very different type of responses to conditions in the EqAtl, EEP, WCAtl and CPac (hereafter LowLat Regime). In general, HighLat Regime is characterized by relatively lower SST and lower PAR but higher Wspd and higher Chl compared to the LowLat Regime.

- 15 Within a single regime, phyto-PFT biomass shows stronger relationships to individual inputs. For example, coccolithopore biomass is positively correlated with PAR in the HighLat Regime (Fig. 3d), and cyanobacteria biomass is highly correlated with PAR in the LowLat Regime (Fig. 4c). Compared to coccolithopores and cyanobacteria, diatoms and chlorophytes exhibit little if any visible relationship with individual indicators, except with Chl. We observe characteristic relationships between Chl and all phyto-PFTs contribution to total Chl [%], which are in agreement with patterns of co-variability between Chl and phyto-PFT contribution [%] derived for bio-optical algorithms (Hirata et al., 2011). Hirata et al. (2011) concluded that Chl is not only an index of phytoplankton biomass but also an index of phytoplankton community structure at synoptic scale. In case of diatoms, we note a mean exponential increase in percentage of total Chl as a function of  $\log_{10}$  of Chl. Our spectrum differs from Hirata et al. (2011) in that there is a larger scatter around the main trend towards the high Chl concentration end. This scatter is associated with lower diatom contributions found in the NEAtl, NorwSea

and partially also the NEPac (Fig. 3i). Comparing the contribution of PhytoANN coccolithophores with haptophytes from the bio-optical model, the relationship differs at the higher end of the Chl range. The PhytoANN estimates that coccolithophores make up a greater proportion of total Chl under bloom conditions observed in the subarctic regions (Fig. 3j). This comparison should be viewed with caution because coccolithophores constitute only a fraction of the larger haptophyte phyto-PFT group considered by Hirata et al. (2011).

When comparing % of total Chl in cyanobacteria in the PhytoANN (Fig. 4i) with *Prochlorococcus* in the bio-optical model (Hirata et al., 2011, their Fig. 2i), we see that maximum percentage contribution is associated with low total Chl in both models. Although the shapes of the two distributions along the Chl gradient are very similar, there are some differences in the magnitude of fractional contribution at lowest Chl values. However, this difference would be minimal if we assumed that NOBM's (and thus PhytoANN's) very broad cyanobacteria group overlaps not only with *Prochlorococcus* but partially also with prokaryotes classified separately in the bio-optical model (Hirata et al., 2011, their Fig. 2g).

Chlorophytes are perhaps most difficult to evaluate because they should to some extent functionally resemble more than one group in the bio-optical algorithm. In the PhytoANN, chlorophytes contribute the most to total Chl under moderate Chl levels (Fig. 4j). Such a pattern is in general close to that of green algae in the bio-optical algorithm (Hirata et al., 2011, their Fig. 2h). The only clear discrepancy occurs for the NorwSea region where the PhytoANN predicts very high percentage contribution of chlorophytes even for the highest Chl values. The comparisons with the bio-optical model therefore suggest that the PhytoANN overestimates the contribution to total Chl of both coccolithophores and chlorophytes in the North Atlantic basin.

It is interesting to compare PhytoANN relationships with those in the NOBM (Figs. A1 and A2). In general, we see that the PhytoANN was able to capture the same general relationships described mechanistically by the NOBM. Considering that NOBM's physical model is partially forced with similar sources of data as used for indicators

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of change in the PhytoANN (e.g. SST, wind stress), this result need not be surprising and may suggest that our derived relationships are artefacts of the empirical or mechanistic relationships in the NOBM. Nevertheless, we observe important differences in how phyto-PFTs are distributed along environmental gradients in the two models, especially in areas modeled only during exploratory analysis. The most striking difference is that in the NOBM, NEPac and AntAtl coccolithophores seem to fall into a separate environmental regime, unlike in the PhytoANN. In the NOBM, coccolithophores are poorly correlated with SST, PAR, Wspd and MLD in those regions. Consequently, their large percentage contribution to total Chl is also not associated with high Chl concentration there. This analysis confirms what was previously described by Gregg and Casey (2007), namely that the NOBM does not predict coccolithophore blooms coincident with high diatom biomass concentrations anywhere outside of the nutrient-replete North Atlantic basin.

The results of this analysis also describe the ecological niches of individual phyto-PFTs which can be compared to the traditional phytoplankton mandala. In the PhytoANN, diatom blooms occur under SST between 5 and 15 °C, PAR between 20 and 45 Wm<sup>-2</sup>, Wspd between 5 and 10 ms<sup>-1</sup>, and Chl above 0.25 mgm<sup>-3</sup>. Coccolithophores are in general more abundant under low mixing (shallower MLD) regimes – consistent with what we know of their ecology (Balch, 2004). Still, they occupy a similar niche to diatoms. Their biogeographical extent is thus considerably greater than in the NOBM (Gregg and Casey, 2007) or another global dynamic PFT model – PlankTOM (Sinha et al., 2010). High latitude blooms predicted by PhytoANN both in the Atlantic and the Pacific are in agreement with in situ and remote sensing coccolithophore estimates in surface waters, as well as with geological records of coccoliths in bottom sediments (Balch, 2004, and references therein). Cyanobacteria and chlorophytes dominate the LowLat Regime which in general can be described by high SST, high PAR, low Wspd and shallow MLD conditions. Their ecological niche is consistent with Margalef's mandala and numerous field studies which concluded that the intensity of surface blooms of cyanobacteria is regulated by a combination of climatic factors, such

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as water temperature, solar radiation, and wind speed (Stal et al., 2003; Whitton and Potts, 2000).

Results from Figs. 3 and 4 also indicate the role of interactions between two or more indicators. For example, under same levels of PAR, there is large build up of cyanobacteria biomass in the LowLat regime but no such buildup in the HighLat. This is clearly because of coinciding differences in SST but likely also variable Wspd and MLD conditions (Fig. 4). Of course these interactions are well known and appear also in the NOBM (Figs. A1 and A2). Yet, it is important to note that the PhytoANN is capable of interpreting these complex and often nonlinear interactions between phyto-PFTs and their ecological indicators because it enables us to verify the first hypothesis of this study.

In order to now say what is the distribution of weights assigned to these interactions, we use a Hinton diagram from one of ten PhytoANNs used to form the ensemble. In Fig. 5 we see that Chl, SST, PAR and Wspd are strongly correlated within a single neuron. MLD appears to be the least significant indicator on its own (but could be significant in combination with others). While it may often be closely associated with Wspd, it can also have an opposite sign assigned, as in the 5th and 7th neuron. We note that in this net, only the 6th and 8th neurons store very similar information about the interactions between inputs. In most other nets included in the ensemble, all eight neurons provide unique information. With respect to connections from hidden layer to the output later, sign and strength of correlations differ from one neuron to another. Note that any one neuron may store important information about one or two PFTs but at the same time provide insignificant information about the remaining phyto-PFTs. This indicates a rather unique phyto-PFT response to various combinations of environmental conditions interpreted by the PhytoANN.

### 3.2 Annual average phytoplankton community composition

The relationship between phyto-PFTs and ecological indicators also reveals important regional differences in phytoplankton community composition. In Fig. 6 we compare

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NOBM and PhytoANN annual mean relative contribution to total Chl biomass, and relate it to any available field estimates. In diatoms, we note that PhytoANN's estimates are in line with those of NOBM in areas where NOBM was used for training. However, significant differences are noted elsewhere. In the NEPac and the AntAtl, PhytoANN predicts less than 20% annual contribution of diatoms which is much closer to observations compared to NOBM (Fig. 6). In the EEP, although PhytoANN estimates a much smaller diatom contribution relative to NOBM (18 vs. 65%), it is still twice as high as the less than 10% observed contribution.

In NOBM, high diatom contribution is almost always at the expense of severely underestimating coccolithophores. This is in contrast to PhytoANN which predicts their 60% contribution compared to almost 0% in the NOBM (Fig. 6). It should be noted that NOBM appears to overestimate coccolithophore contribution in two out of four regions used to train the PhytoANN (with two others not evaluated against observations). Hence, it is expected that the PhytoANN overestimates coccolithophores in those and other regions as well (e.g. EEP, NEPac and AntAtl). Nevertheless, the fact that PhytoANN predicts a significant contribution of coccolithophores suggests the existence of suitable ecological niches for coccolithophores in the NEPac and AntAtl, as well as in the EEP.

Similarity of ecological niches of the NEAtl, NEPac and AntAtl are evident both in PhytoANN and NOBM. For example, both models rely on a similar total Chl which is not surprising considering that NOBM assimilates SeaWiFS Chl during its simulation. Therefore, the main difference between the model phyto-PFT distribution in these regions originates primarily from distinct partitioning of Chl between the phyto-PFTs. In NOBM it depends on nutrient uptake and light availability in a dynamical context, while in PhytoANN only on the favourability of environmental conditions.

As for cyanobacteria, the largest difference between the two models is observed in the EEP (Fig. 6). Here, higher biomass estimates from PhytoANN are closer but still significantly lower than observed. On the other hand, though similar to NOBM, PhytoANN biomass estimate is much too low in the EqAtl relative to observations. This







PhytoANN's monthly coccolithophore climatology matches well with bio-optical algorithm results, also in relation to other basins. Furthermore, PhytoANN suggests that chlorophytes constitute a substantial portion of total Chl from April to September (close to 50 %). This is in close agreement with another bio-optical model result of Alvain et al. (2008, their Fig. 10) who based on 1998–2006 monthly climatology predict that diatoms contribute at most 52 % to total Chl and that nanoflagellates (most equivalent to our chlorophytes) contribute over 80 % from April to September in the Southern Ocean (40–70° S, 180°–180° E). We conclude that PhytoANN is able to improve phyto-PFT monthly climatology picture from NOBM in the AntAtl using the knowledge of ecological rules inferred during training from the NOBM itself.

Quantifying coccolithophore biomass has been extremely difficult both in dynamic and diagnostic models. Except for the recently improved PhytoDOAS algorithm (Sadeghi et al., 2012), most bio-optical models show the spatial extent of bloom areas (Alvain et al., 2008; Moore et al., 2012). When analyzing the results of the Dynamic Green Ocean Model, Le Quéré et al. (2005) noted that model calcifiers grow between 40° N and 40° S but they are almost absent poleward of these latitudes. Satellite observations following the method of Brown and Yoder (1994) reveal highest coccolithophore bloom frequencies in the 40–70° latitude band of both hemispheres (Le Quéré et al., 2005, their Fig. 10). Le Quéré et al. (2005) suggests that this is because the traits defined for calcifiers and the zooplankton that graze on them do not give calcifiers a competitive advantage at high latitudes. Compared to other PFTs, they have a lower maximum growth rate, higher light affinity, and lower resistance to darkness. In the NOBM, coccolithophore are very abundant beyond 40° of latitude but only in the North Atlantic. In high latitudes of other basins they are severely outcompeted by diatoms (Gregg and Casey, 2007). Current dynamic model results indicate that we have an insufficient knowledge of either traits of calcifiers (e.g. vital rates), or their protective defenses against zooplankton grazing (Strom, 2002). Even though our PhytoANN considers neither nutrient competition nor zooplankton grazing responses, it provides a more

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realistic and ecologically-consistent picture of coccolithophore distribution in the high latitude regions.

### 3.4 Dramatic shifts in phyto-PFT distribution

We also wanted to check how our diagnostic model performs when it comes to projecting changes under extreme input conditions. Here, we choose to focus on the EEP where interannual El Niño Southern Oscillation (ENSO) contributes the most to temporal variability (Wang and Fiedler, 2006, and references therein). ENSO cycles not only shift total biomass by almost two orders of magnitude but also alter the phytoplankton community composition significantly. In order to really test the PhytoANN response to the extreme 1997–1999 El Niño and La Niña events, we not only excluded the EEP but also the 10.1997–12.1999 period from all data used for confirmatory analysis.

In Fig. 9a we first note that the NOBM captures a sharp decrease in diatoms percentage contribution to total Chl during El Niño. The coincident decrease in chlorophytes contribution is not distinct from the monthly climatological pattern (Fig. 7e). Coccolithophores and cyanobacteria respond positively to the warmer but nutrient poor El Niño conditions. In the PhytoANN, both diatoms and chlorophytes contribute much less to total Chl from October 1997 to summer 1998 than on a long-term average (Fig. 9b). Also, contrary to the NOBM, coccolithophores do not increase their contribution to total Chl in response to El Niño. In turn, cyanobacteria dominate the phytoplankton population as they constitute up to 80 % of total Chl biomass around the peak of El Niño conditions.

In response to the subsequent La Niña event, diatoms in the NOBM restore their high contribution achieving more than 80 % of total Chl by fall 1998. However, we also note that the long-term mean diatom percentage contribution is grossly overestimated. In the NOBM, coccolithophores and cyanobacteria reach their 1997–2005 all time low by the beginning of 1999. In the PhytoANN, we see a different response of the phytoplankton community. All phyto-PFTs return to their long-term average contribution levels but do not reveal the expected equally dramatic changes in phytoplankton

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community composition associated with La Niña high Chl blooms. The substantial increase in Chl is mostly attributed to chlorophytes and only secondly to diatoms. Compared to the NOBM results, the PhytoANN suggests more coexistence of phyto-PFTs even under the extreme La Niña conditions. We also note that along the entire time series we see evidence of coccolithophores becoming the dominant group at times (fall 1999 and 2000) in PhytoANN (Fig. 9b). NOBM simulates interannual variability in coccolithophore biomass as well but in different years (fall 2002 and 2003) and never beyond the diatom levels (Fig. 9a). This suggests that the two models also predict different ecological responses of phyto-PFTs to interannual changes in the ecology of the EEP.

In the bio-optical model (Fig. 9c), haptophytes, which include but are not limited to coccolithophores, dominate the biomass spectrum consistently throughout the entire time series. The sum of *Prochlorococcus* and prokaryotes are on a similar level to cyanobacteria modeled by the PhytoANN, but markedly higher than in the NOBM. Diatoms remain on a very low (less than 10%) level of contribution to total Chl, in close agreement with in situ observations (e.g. Taylor et al., 2011). The bio-optical model reveals a moderate increase in *Prochlorococcus* and prokaryotes contribution to total Chl during El Niño, and it also captures the large increase in diatom contribution during the subsequent La Niña event. During this time diatoms far exceed their long-term average levels. Such a response was also indicated by in situ (Chavez et al., 1999), some remote sensing (Masotti et al., 2011) and dynamic modeling studies (Gorgues et al., 2010), which report that diatoms first decreased and later increased significantly in response to El Niño and La Niña, respectively.

Based on these results, we can conclude that our diagnostic model appears to be moderately sensitive to extreme environmental perturbations and can detect only some significant temporal shifts in phytoplankton community composition. Neither the NOBM nor the PhytoANN fully capture the observed strong changes in diatom contribution during an extreme ENSO cycle. Both models perform better when simulating changes in response to El Niño rather than La Niña conditions. Higher spatial and temporal

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resolution of model estimates would allow to test if our model is also sensitive to high-frequency submesoscale shifts in phyto-PFT distribution attributed to passing tropical instability waves (Palacz and Chai, 2012; Parker et al., 2011).

### 3.5 Assessment of model limitations

First, the ecological indicators used by PhytoANN do not include nutrients concentration and zooplankton grazing explicitly. Effects of changing nutrient concentrations are largely inferred from changes in SST, as SST and nitrate are often closely correlated, e.g. at the BATS site (Nelson et al., 2004). Including field measurements of Si and Fe as separate indicators is expected to improve our model predictive skill. However, using NOBM nutrient fields as inputs did not improve PhytoANN performance in either confirmatory or exploratory analysis in any significant manner. We speculate that some grazing effects are only implicitly included in our model through considering total Chl as an indicator. However, this does not resolve phyto-PFT-specific grazing pressures. Consequently, PhytoANN's interpretation of ecological rules distinguishing between phyto-PFTs is strongly biased towards bottom-up control processes. We are currently developing a prototype of a similar model that includes feedback from key zooplankton PFTs. It is to be implemented in the North Atlantic basin where basin-wide longterm coverage of Continuous Plankton Recorder data enables such an experiment (Raitzos et al., 2008).

Second, the source of several ecological indicators is common to the NOBM and the PhytoANN. This means that PhytoANN's interpretation of some the relationships between phyto-PFTs and the environment is not necessarily independent from NOBM's. However, this does not hinder the verification of our hypotheses, especially considering the large differences in phyto-PFT distributions in regions used only for exploratory analysis.

Third, we assume that the NOBM phyto-PFT distribution used in training most closely represents the in situ conditions in the Atlantic basin. This assumption is based on the fact that NOBM assimilates SeaWiFS Chl data to correct for total PFT biomass levels

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and that it was generally positively evaluated against observations in the North Atlantic (Gregg and Casey, 2007) among other regions. Although in situ data would provide the ideal input, currently available in situ phyto-PFT biomass data, with strong seasonal and geographical bias, is insufficient to represent most of biogeochemical conditions in the open ocean.

Fourth, PhytoANN shows only a surface picture of phyto-PFT biogeography. This limitation will be difficult to overcome considering that all remote sensing indicators provide a surface view themselves, although, there are approaches that try to account for vertical changes in bio-optical algorithms (e.g. Uitz et al., 2006; Brewin et al., 2010).

#### 10 4 Implications

The results of this study highlight the benefits of using advance statistical techniques to unravel complex and highly non-linear ecological interactions, with implications for biogeosciences and marine ecosystem management alike. We demonstrate that through an artificial neural network we can combine remote sensing and dynamic model results to generate new, ecologically-consistent estimates of phyto-PFT distribution in a wide range of biogeographic conditions. If ecological rules can be extracted from weights assigned to connections within PhytoANN, then we can provide biogeochemically-specific parameterizations of phyto-PFT growth functions which are currently too rigid to capture the global variability in phytoplankton vital rates. This approach should help model the global distribution of silicifiers and calcifiers correctly so that we can reduce the uncertainty on how much atmospheric carbon is being fixed into biomass and how much is being exported into the deep ocean (Francois et al., 2002; Rost and Riebesell, 2004; Sarmiento et al., 2002). In turn, this will improve our future projections of global carbon fluxes and climate mitigation plans.

Unlike other diagnostic PFT models, PhytoANN can be used to make future projections under scenarios of climate-induced changes to key environmental indicators. This is because it takes inputs that are also modeled by most coupled NPZD models

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running in forecast mode. In this study, we verify the hypothesis that PhytoANN is able to interpret the complex and nonlinear interactions between phyto-PFTs and the environment, at least to the same extent as the original training model, yet in a fraction of time required to make a dynamic simulation. We speculate that our PhytoANN could be used to interpret similar relationships in an ensemble of coupled models and later applied to future time series of indicators of change. This would provide a novel framework for constructing such ensemble model projections to examine the differences and similarities between them, and eventually lead to better constrained future projections of PFT states.

Furthermore, PhytoANN type of models can further be developed to look at phytoplankton size classes and their dependency on changing environmental conditions, such as temperature. Shifts in community size spectrum in response to rising temperatures are suggested by some studies (e.g. Hilligsøe et al., 2011), yet in others they are shown to depend primarily on total biomass and productivity (e.g. Marañón et al., 2012). Hence, ANNs could prove useful in examining these interactions in the context of variability among biogeochemical provinces rather than global average trends.

Similarly, this approach can be potentially expanded to include higher trophic levels, from zooplankton functional groups to fish species. Depending on data availability among other things, this could first be tested within a single ecosystem. If successful, such a model could provide an alternative framework to the newly proposed General Ecosystem Models (GEMs) that aim at resolving complex and adaptive properties of ecosystems (Purves et al., 2013). ANNs are already heavily relied upon in system control and management applications in other disciplines such as electrical engineering, medical science or business and economics (Anandarajan et al., 2001; Jaeger and Haas, 2004; Khan et al., 2001). Here, we demonstrate that only very few measurable, specific and sensitive indicators of change of phyto-PFTs are sufficient to capture key seasonal to interannual patterns of their distribution. While it is far more difficult to come up with indicators of change of key zooplankton and fish species, such efforts are being currently undertaken under the auspices of the European Union 7th

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**Table 1.** Spatial and temporal characteristics of the input and target data used in confirmatory and exploratory analysis.

	confirmatory analysis	exploratory analysis
spatial resolution	1°	box-average
temporal resolution	monthly	monthly
time series length	Jan 2000–Dec 2004	Oct 1997–Dec 2004

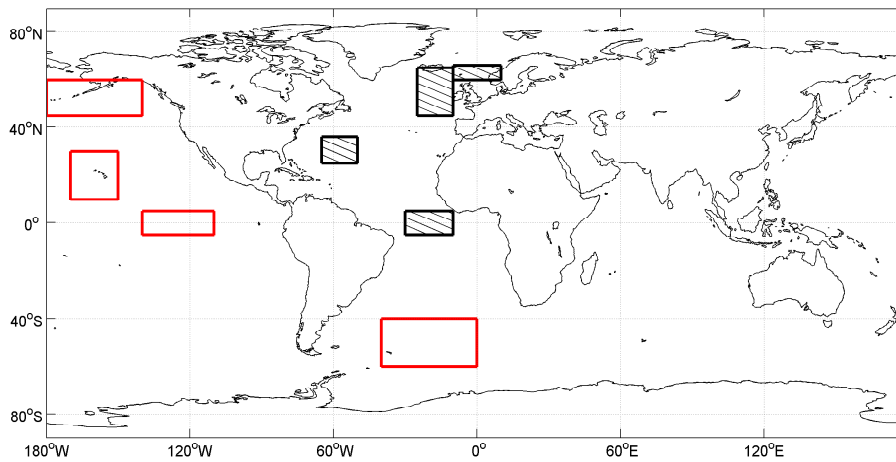
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**Table 2.** Default PhytoANN parameters chosen for this study.

Parameter	Value
ANN type	Feedforward
#hidden layers	1
#neurons in hidden layer	8
#inputs	5
#outputs	4
1st layer transfer function	tangential sigmoidal
2nd layer transfer function	linear
Learning rate parameter	dynamic
Training algorithm	Levenberg–Marquardt
Data division mode	random, every sample
training : validation : testing [%]	70 : 15 : 15
Generalization scheme	early stopping
Performance function	mean squared error

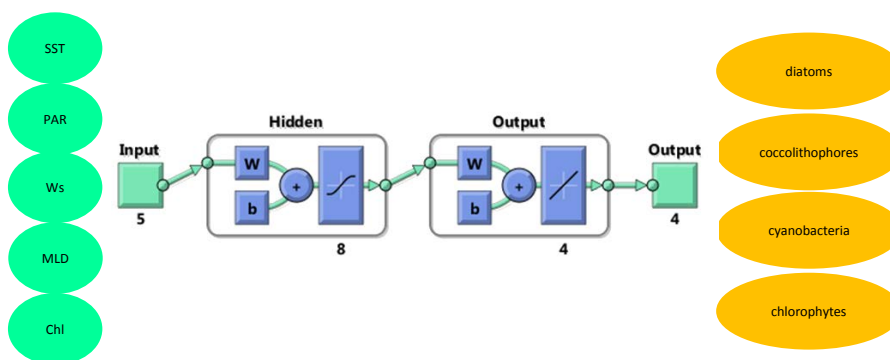
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**Fig. 1.** Location of the areas used for both confirmatory and exploratory analysis (black) and only exploratory analysis (red). Black boxes correspond to the NorwSea, NEAtl, WCAtl and EqAtl when looking from north to south. Red boxes correspond to the NEPac, CPac, EEP and AntAtl when looking from north to south.

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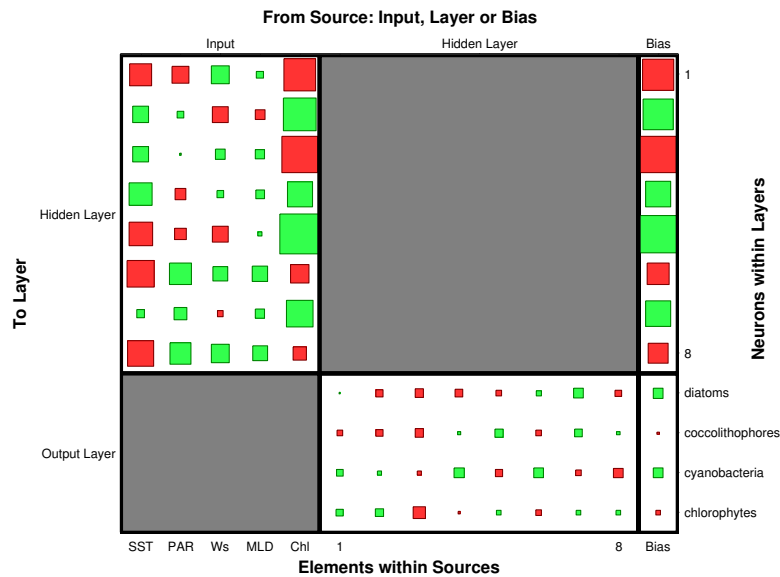


**Fig. 2.** A conceptual model illustrating how the ANN interprets the relationship between input variables, and between input and target variables. Each hidden neuron in layer 1 computes the sum of input values multiplied by connection weights, and calculates the activation value of each neuron (being a nonlinear sum of all inputs). Each output neuron in layer 2 is a linear sum of the activation values from all layer 1 neurons multiplied by respective connection weights.

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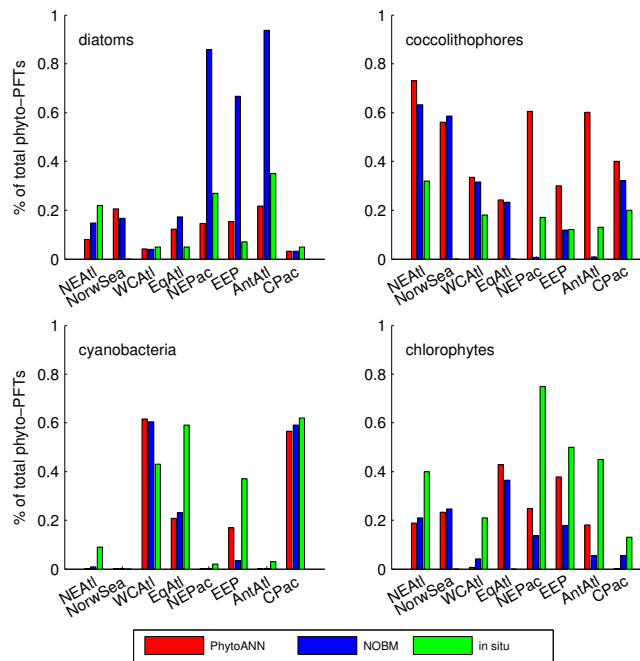






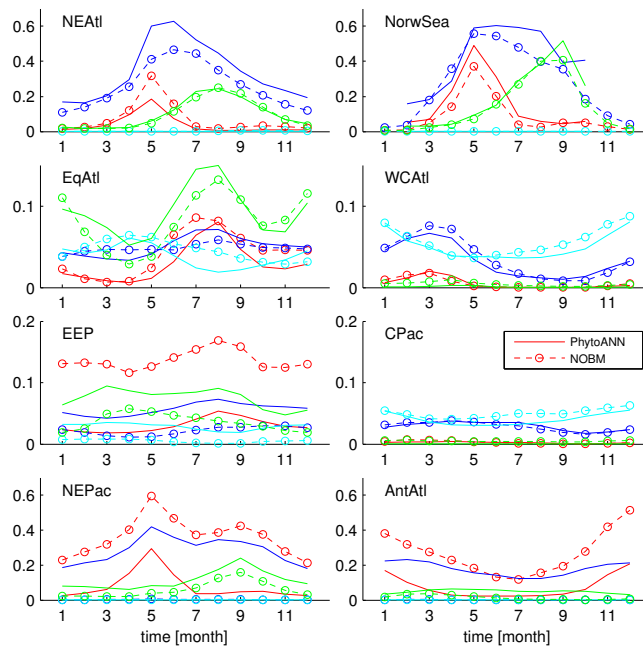
**Fig. 5.** Hinton-weight diagram depicting relative weights assigned to all connections within one of ten PhytoANNs used to create the ensemble. Green color corresponds to positive correlation, red color negative correlation. Area of the square is proportional to the strength of correlation. Left hand size illustrates the connection weights between the input and the hidden layer. The right hand side illustrates the connection weights between the hidden layer and the output layer.

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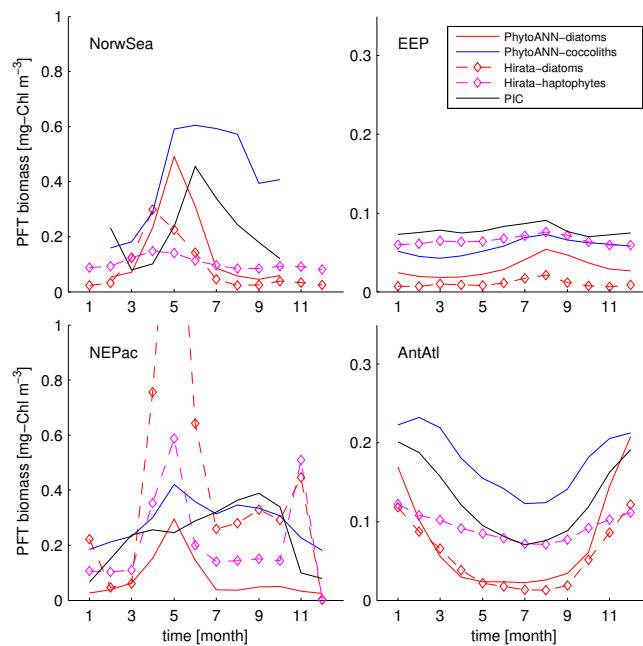
**Fig. 6.** Approximate box or basin-average comparison between longterm (1997–2004) annual mean percentage contribution to total Chl biomass of diatoms, coccolithophores, cyanobacteria and chlorophytes according to NOBM (blue), PhytoANN (red) and in situ (green) estimates. In situ estimates were derived from both C- and Chl-based measurements. Here, estimates were derived from a combination of data collected by Gregg et al. (2003) (<http://gmao.gsfc.nasa.gov/research/oceanbiology/data.php>) and EEP data from Taylor et al. (2011, their Table 3).

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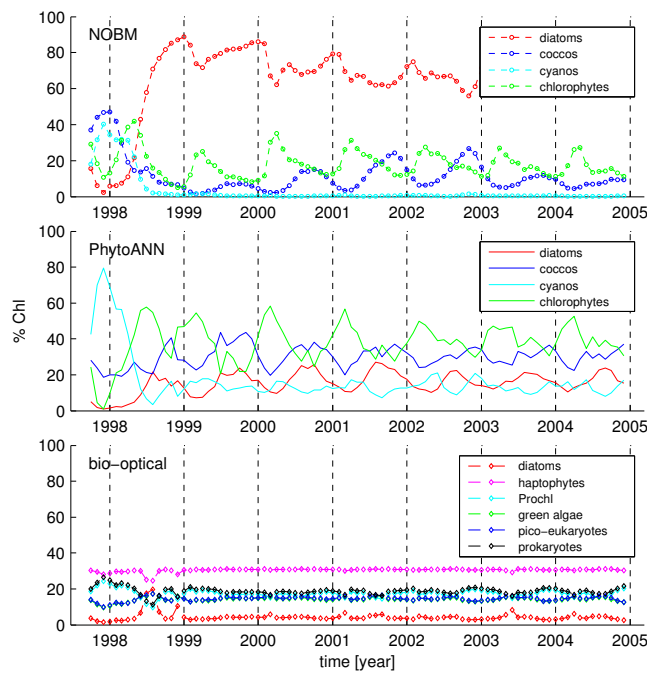
**Fig. 7.** 1997–2004 longterm monthly climatology of NOBM (dashed lines with open circles) and PhytoANN (solid lines) phyto-PFT biomass estimates from confirmatory (**a** NEAtl, **b** NorwSea, **c** EqAtl, **d** WCAtl) and exploratory areas (**e** EEP, **f** CPac, **g** NEPac, **h** AntAtl). Each plot includes four PFTs: diatoms (red), coccolithophores (blue), chlorophytes (green) and cyanobacteria (cyan). All phytoPFT biomass in units of  $\text{mg Chl m}^{-3}$ .

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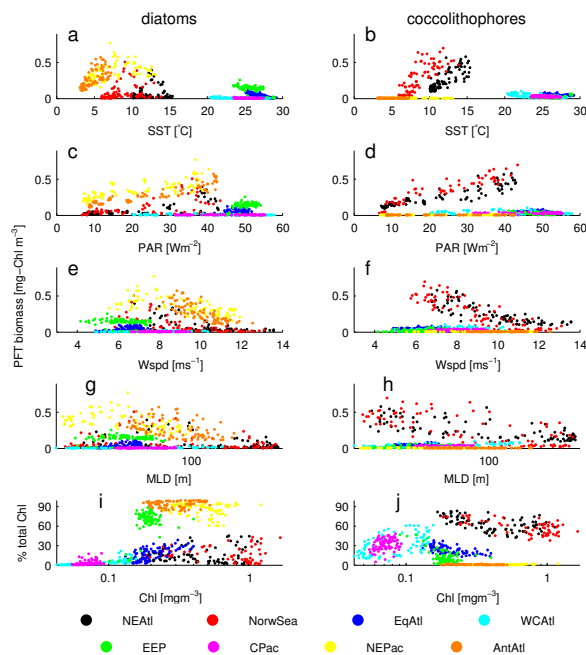
**Fig. 8.** Longterm monthly climatology of diatom (red) and coccolithophore (blue) biomass from PhytoANN compared with diatoms and haptophytes (magenta) from the bio-optical algorithm as well as PIC from space (black) in four areas: (**a**) NorwSea, (**b**) EEP, (**c**) NEPac and (**d**) AntAtl. PhytoANN and PIC results are marked with a solid line, while bio-optical algorithm with dashed line. PIC values are multiplied by an arbitrary factor of 300 to match scales.

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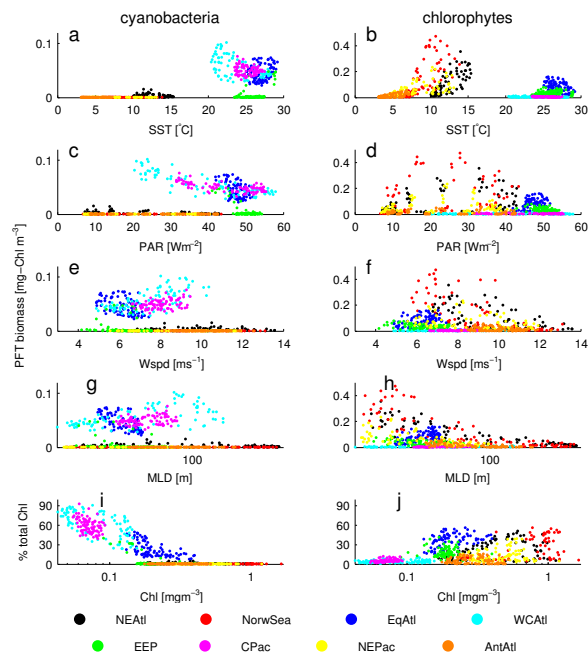
**Fig. 9.** October 1997–December 2004 time series of the percentage contribution to total Chl of four phyto-PFTs in the EEP from (a) NOBM, (b) PhytoANN, and (c) of six phyto-PFTs from the bio-optical algorithm (Hirata et al., 2013).

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**Fig. A1.** Scatter plots showing the emerging relationship between NOBM diatom (left column) and coccolithophore (right column) biomass [ $\text{mg Chl m}^{-3}$ ] (a–h) or % of total Chl (i, j) vs. individual ecological indicators color coded according to domain of origin. On the x-axis from top to bottom: SST, PAR, Wspd, MLD, Chl. All data points come from exploratory analysis only, thus they are box-average and monthly average estimates from October 1997 to December 2004.

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**Fig. A2.** Scatter plots showing the emerging relationship between NOBM cyanobacteria (left column) and chlorophyte (right column) biomass [ $\text{mg Chl m}^{-3}$ ] (a–h) or % of total Chl (i, j) vs. individual ecological indicators color coded according to domain of origin. On the x-axis from top to bottom: SST, PAR, Wspd, MLD, Chl. All data points come from exploratory analysis only, thus they are box-average and monthly average estimates from October 1997 to December 2004.